

PREREMEDIAL DESIGN INVESTIGATION REPORT

for the

Shore Realty Superfund Site Glenwood Landing, New York

VOLUME I

Prepared for:

The Performing Parties Group

Prepared by:

Remediation Technologies, Inc. 9 Pond Lane Concord, Massachusetts 01742

Project No. 3-1033

July 1993



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EXECUTIVE SUMMARY

The Pre-remedial Design Investigation (PDI) was focused on filling data gaps associated with engineering parameters necessary to complete the remedial design of the remedy. The field program consisted of the installation of six well points, one on-site monitoring well, three borings, a time-lag stage-ratio study, the collection of nine soil samples and the collection of sixteen groundwater samples. Laboratory studies included chemical analyses of the environmental samples obtained and treatability studies, including an iron precipitation evaluation, a water pre-treatment study, and a biological treatment study.

The results of the PDI were generally similar to those from the Remedial Investigation (RI). Two conclusions can be drawn from the PDI work. First, a reduction in the number of Volatile Organic Compounds (VOCs) detected occurred between the RI and the PDI, with the levels of the toluene, ethylbenzene and xylene concentrations in the water table samples being greatly reduced. Second, polynuclear aromatic hydrocarbons (PAHs) should not be a concern at the site since they are not present in the groundwater at the site and therefore are not migrating.

The off-site monitoring well was installed and sampled in June of 1993. Due to property access issues this well was not installed during the February PDI as planned. The well has been located on property owned by Nassau County. Analytical results will be submitted with this July's monthly progress report.

Due to the iron concentrations in the soil and groundwater identified at the site during the RI, it was necessary to perform an iron precipitation evaluation to evaluate the potential of fouling the aquifer and thereby inhibiting the implementation and effectiveness of the selected remedy. The study was performed during the course of three months. The soil in the column study did appear darker and there were areas of iron staining; however, the permeability was reduced not because of iron precipitate, but rather from settlement of the soil column which is not indicative of what will occur in the soils at the site when the remedy is implemented. The iron precipitation study has shown that fouling of the aquifer and soil matrix should not cause a significant reduction in permeability of the aquifer which would adversely impact implementation of *in situ* aerobic biological treatment and aquifer aeration.

The time-lag stage-ratio study was performed in place of the traditional pump test in order to determine the permeability of the soils at the site over a large area. The premise of the study is that as the tide rises and falls, a sinusoidal propagation of the water table occurs. Thus the time that the tide takes to travel a known distance can be used to determine the *in situ*

permeability of the soil at the site. The study was performed for one week during a full moon period to ensure maximum tidal fluctuation. The study determined the permeability of the soils at the site to be 57 ft/day (10-2 cm/sec). This value is within reported limits for soil formations on Long Island and is consistent with the values determined during the RI.

The water pre-treatment study was performed in order to determine how effectively iron could be removed from the groundwater prior to treatment or removal of the contaminants. Six treatability runs were performed in order to evaluate various iron removal strategies. Three treatment scenarios effectively reduced the iron concentration in the water to acceptable levels. During the design process itself, an economic analysis will be performed in order to determine which of these three scenarios is the most cost-effective strategy to remove the iron.

The biological treatment study was performed in order to determine if the conditions at the site were conducive to bio-remediation and to determine nutrient additions and pH adjustments necessary to optimize the degradation of the contaminants present at the site. In general, the conditions at the site are appropriate for implementing *in situ* bio-remediation. The microbiology at the site appears to be conducive to *in situ* bio-remediation, since there was a high density of total and VOC degrading bacteria. Additionally, the bacteria showed good oxygen uptake, activity indicating bio-degradation was occurring. The nutrient addition data from the treatability study was conflicting, in that nutrient addition caused an increase in oxygen demand but no increase in population. The soil contaminant analytical data showed that in-situ biodegradation was feasible at the site and that the constituents at the site are biodegradable based on the reduction of contaminant concentrations in the active samples. Since degradation did occur and there was an increase in oxygen uptake during the nutrient addition tests, the reinjected water will be amended with nutrients to optimize degradation rates of the contaminants at the site.

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* Not provided with this report to Technical and Executive Committees. Copy in project file at RETEC's Concord office.

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LIST OF ACRONYMS

AES Applied Environmental Services

ARARS Applicable or Relevant and Appropriate Requirements

BDL below detection limits CFUs colony forming units

Cl₂ chlorine

COIs constituents of interest

CVOCs chlorinated volatile organic compounds
DNAPL Dense Non-Aqueous Phase Liquid

DW deep water Fe²⁺ ferrous iron Fe³⁺ ferric iron

gpd/ft² gallons per day per square foot

H₂O₂ hydrogen peroxide KMnO₄ potassium permanganate

LILCO Long Island Lighting Company LNNL light non-ageous phase liquid

MSL mean sea level NaOCl sodium hypochloride

NYSDEC New York State Department of Environmental Conservation

 O_3 ozone

PAHs polynuclear aromatic hydrocarbons
PCE perchoroethlyene (tetrachloroethylene)
PDI Preremedial Design Investigation

PDIR Preremedial Design Investigation Report

ppm parts per million

RETEC Remediation Technologies, Inc.

RI Remedial Investigation

RI/FS Remedial Investigation/Feasibility Study

ROD Record of Decision
Roux Roux Associates, Inc.

SARA Superfund Amendments and Reauthorization Act of 1986

SOP Standard Operating Procedure SVOCs semivolatile organic compounds

SW shallow water

TCA 1,1,1-trichloroethane TCE trichlorethylene

TSS total suspended solids
VOCs volatile organic compounds

WT water table

July 20, 1993

1.0 INTRODUCTION

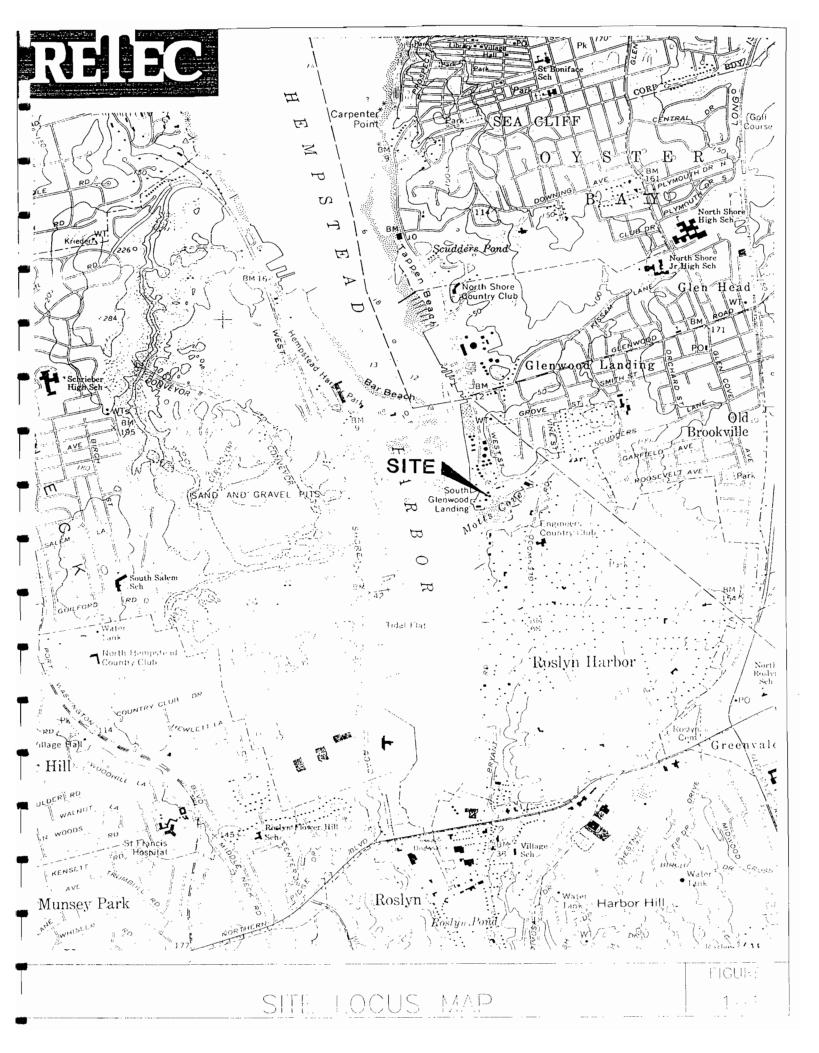
Remediation Technologies, Inc. (RETEC), on behalf of the Performing Parties Group, is submitting this Preremedial Design Investigation Report (PDIR) which summarizes the results of the Preremedial Design Investigation (PDI) undertaken at the Shore Realty Site (the "Site") in Glenwood Landing, Nassau County, New York. The PDIR is intended to present methodologies, procedures, results, and conclusions from the PDI.

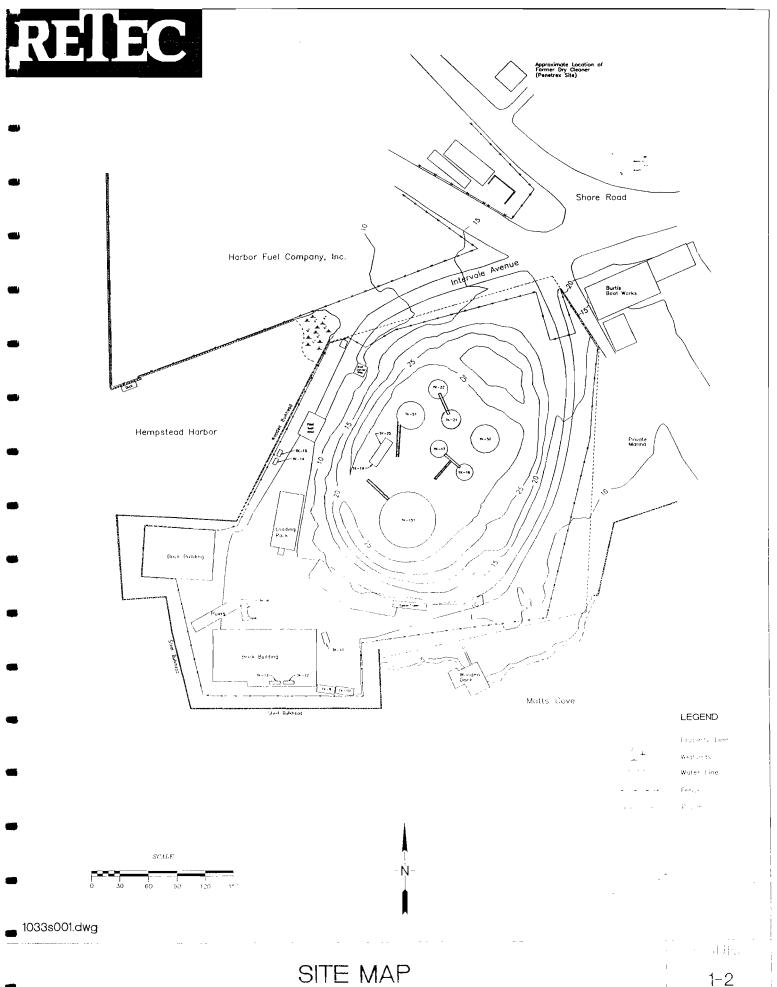
1.1 SITE DESCRIPTION

The Site, located at One Shore Road, Glenwood Landing, New York, is approximately 3.2 acres in size, and surrounded on three sides by water: Motts Cove to the east and south, and Hempstead Harbor to the west (Figures 1-1 and 1-2). Both water bodies and associated intertidal areas are designated tidal wetlands by the State of New York (the State). The Site is at an elevation of approximately 5 to 30 feet above mean sea level (MSL).

The Site is surrounded by industrial, commercial and residential areas. Directly north of the Site on Hempstead Harbor is the Harbor Fuels oil terminal, and 200 feet to the north is an inactive hazardous waste disposal site (the Penetrex Site) which is a former dry cleaner. Approximately 600 feet farther north on Hempstead Harbor, along Shore Road, is the Long Island Lighting Company (LILCO) power station. Directly east of the Site, on Motts Cove, is a private marina, Burtis Boatworks. Approximately 200 feet northeast, upgradient of the Site, is a residential area.

The Site contains three brick buildings which are a pump house/storage building, warehouse, and office/garage. There are seven fixed above ground storage tanks ranging in size from approximately 56,200 gallons to 740,500 gallons. One of these storage tanks is split internally into two compartments. A canopied truck loading rack is located on-site along with the associated piping infra-structure. There are several other surface structures, including a burned trailer, six unmounted storage tanks, a tank trailer, van, boat, and truck. Underground storage tanks used for storing fuel oils, diesel fuels, and other liquids for on-site activities, i.e., building furnaces, site vehicles, etc., are presently in their original locations waiting for sampling, removal and disposal.





1.2 PREREMEDIAL DESIGN INVESTIGATION REPORT ORGANIZATION

This report is organized into seven narrative sections and eight appendices. Section 1.0 is the introduction. Section 2.0 contains site background information, site history, past investigation summaries, a discussion of the constituents of interest, and a description of the selected remedy. Section 3.0 presents recommended objectives for the remedial activities for the Site. Section 4.0 discusses the scope of the investigation. Section 5.0 presents the results of the investigation. Section 6.0 presents the treatability results and Section 7.0 lists references used. Appendix A presents the boring logs from the PDI. Appendix B presents the chain of custodies from the sample shipment. Appendix C presents the monitoring well logs. Appendix D is the soil gas survey field data sheets. Appendix E provides the bio-treatability study data sheets. Appendix F presents the water pretreatment data sheets. Appendix H is the soil analytical data sheets from the laboratory. Appendix G is the groundwater analytical sheets from the laboratory.

2.0 SITE HISTORY

2.1 SITE HISTORY

The Shore Realty property was first used for fuel storage purposes in 1939. In 1974, the site changed hands and was used for the storage and distribution of chemical solvents. Numerous spills of organic chemicals reportedly occurred during this period. In October 1980, the owner's did install monitoring wells and a recovery trench. In 1980, the property was leased to Applied Environmental Services (AES). AES operated the facility for the blending of various chemical waste materials that have a heat value to provide alternate fuel sources. AES also operated a hazardous waste storage facility.

AES continued the monitoring and recovery efforts undertaken, and installed product recovery equipment. The trench containing the product recovery equipment reportedly recovered approximately 500 gallons of liquid chemicals per month during 1981 and 1982. Groundwater samples collected in 1982 were found to contain dissolved concentrations of volatile halogenated and non-halogenated hydrocarbons.

Shore Realty Corporation purchased the Site in October 1983, and evicted AES in January 1984. The State filed suit against Shore Realty and its owner, Donald Leogrande, in February 1984. As a result of that suit, Shore Realty and Leogrande were ordered by the court to undertake remedial actions at the Site. Subsequent thereto, Shore Realty and Leogrande commenced a third party action against numerous defendants, including the prior landowners, prior on-site operators and a number of companies that had allegedly sent chemicals to the Site, while it was operated by AES.

In March 1984, the State inventoried and sampled chemicals contained on-site and collected surface water samples from Hempstead Harbor. From 1985 to 1986, a State contractor removed more than 700,000 gallons of chemicals stored in the above ground tanks and Shore Realty, under State supervision, removed all of the 55-gallon drums stored in the drum storage warehouse. However, drums currently exist on-site from the RI/FS activities. All of the aboveground tanks containing liquids were reportedly emptied and de-contaminated under State supervision.

In February 1987, a group of third-party defendants retained Roux Associates (Roux) to conduct a Remedial Investigation/Feasibility Study (RI/FS) at the Site, which was completed in April 1991. The Record of Decision (ROD) was prepared by New York State Department of

Environmental Conservation (NYDEC) and the U. S. Environmental Protection Agency in June 1991. The Consent Judgement was lodged in June 1992, and entered on August 5, 1992.

2.2 PREVIOUS INVESTIGATIONS

2.2.1 Applied Environmental Services

Groundwater monitoring and product recovery efforts were initiated by AES in 1981. Several monitoring wells and a product recovery trench were installed and sampled for analysis for volatile organic compounds. See Figure 2-1 for the locations of the wells and trench. The data indicated that dissolved volatile and semi-volatile constituents were present in the groundwater and the soils.

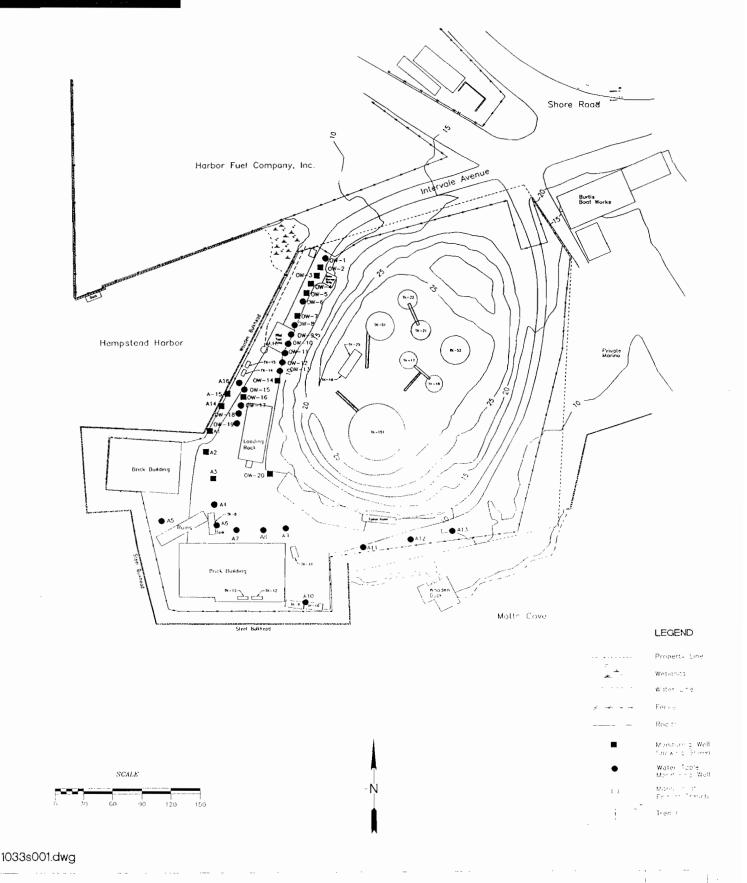
2.2.2 RI/FS by Roux Associates, Inc.

In February 1987, Roux was retained by a group of third-party defendants to conduct a RI/FS for the Site. In May 1987, Roux submitted a Work Plan for the RI/FS to the State of New York. The Work Plan was approved in September 1987. The initial RI began in October 1987, and the supplemental RI began in October 1989.

The RI investigation addressed the mediums of air, soil, sediments and groundwater. See Figure 2-2 for the locations of the RI sampling points. An air monitoring program was conducted to screen ambient air quality conditions to identify health and safety personal protection levels. Additional air monitoring was performed over the mud flats adjacent to the Site. Samples were collected from five monitoring points, six feet above the mud flats, and submitted for volatile organics analysis. Three of these samples were collected west and two were collected south of the Site. Benzene, toluene and ethylbenzene were detected in three of the samples. Only benzene was above New York State's ambient air guideline concentrations.

The soil investigation at the Site was initiated to determine the approximate levels and extents of chemicals in the soil (Roux, 1991a). The investigation included the analysis of soil gas and soil from the Site. A soil gas survey was conducted during the initial RI and during both the winter and spring months of the supplemental RI. The investigative technique used by Roux was limited and could analyze soil gas from a depth of only a few feet below the ground surface. A TIP II meter was used for analyzing the soil gas samples.

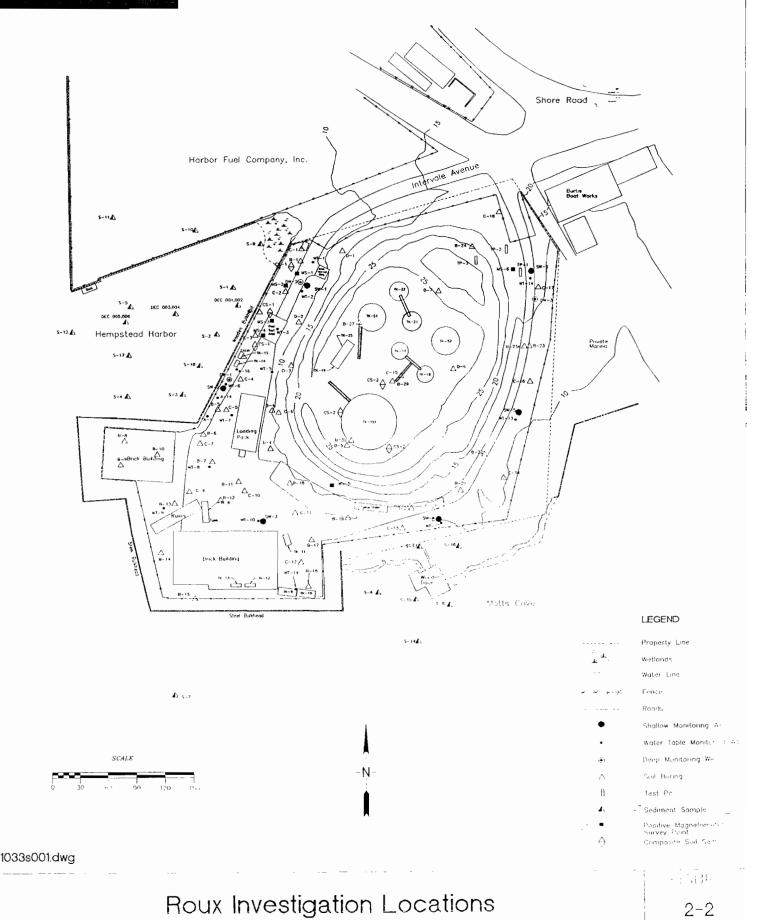
REFEC



Location of Wells By AES

2-1

RELEC



The soil boring program entailed the installation of thirty-eight soil borings and collection of thirty soil samples for chemical analysis during the initial RI. Twenty-five additional borings were installed and thirty-two soil samples were collected for chemical analysis to more fully characterize the lateral and vertical extent of soil contamination during the supplemental RI work.

The soil analytical program entailed the analysis of select soil samples for 129 priority pollutants plus 40 peaks, with the remainder of the samples analyzed for purgeable organics, priority pollution metals, PCBs, and semi-volatile organic compounds. The soil analytical results identified chlorinated volatile organic compounds (CVOCs), non-chlorinated volatile organic compounds, polynuclear aromatic hydrocarbons (PAHs), phthalates, phenols, and metals. The predominant chemicals found at the site were from the non-chlorinated suite of compounds including ethylbenzene, toluene, and xylenes. These three compounds comprises more than 98 percent by mass of the chemical constituents of concern identified at the Site.

The sediment investigation included the collection of samples during both the initial and supplemental RIs. Eight samples were collected during the initial RI. Four of these were analyzed for 129 priority pollutants plus 40 peaks. The last four samples were analyzed for purgeable organics, priority pollutant metals, and PCBs. Seventeen samples were collected from ten locations during the supplemental RI. All were analyzed for volatile organics, semi-volatile organics, and metals.

The groundwater medium was investigated through the Hydrogeologic and Groundwater Quality Investigation. The purposes of the hydrogeologic investigation were to:

- understand the occurrence, movement, and discharge of groundwater beneath the Site;
- determine the potential impact of the Site on groundwater quality; and
- determine whether drinking water resources are presently or potentially impacted by the Site.

Nine monitoring wells were installed for the initial RI. These wells, plus six existing monitoring wells, were sampled during the initial RI. The wells are arranged into three groups, i.e., at the water table (WT), shallow water (SW) and deep water (DW) wells. The WT wells are screened across the water table surface. The SW wells are screened below or near the water table surface. The DW wells are screened deeper into the aquifer. All of these water samples were analyzed for either 129 priority pollutants plus 40 peaks, or U.S. EPA Method 624 for

purgeable organics, priority pollutant metals and PCBs. During the supplemental RI, two additional monitoring wells were installed and sampled, along with the 14 existing monitoring wells.

During December 1990, Roux conducted a well search for the area within a-one-mile radius of the Site, referring to USGS publications for information on existing wells and their use within the researched area. Roux also contacted Sea Cliff Water Company, Jericho Water District, Roslyn Water District and Glenwood Landing Water District to update the published data. Only wells on the east side of Hempstead Harbor have been included in the results of the well search because Hempstead Harbor is a groundwater discharge area which acts as a constant head boundary to the water table aquifer.

In summary, the results of the well search reported in the RI indicated the following:

- there are no supply wells of any kind downgradient of the Site;
- there are no public supply wells within one mile of the Site;
- all potable water is supplied to the Glenwood Landing area by public supply wells located outside of the Glenwood Landing Water District;
- there are no domestic wells in close proximity to the Site which may be impacted by contamination at the Site; and
- the few commercial, irrigation and industrial wells being used in close proximity to the Site are all located upgradient or cross-gradient to the Site, and are therefore not at risk of being impacted by water from the Site.

2.3 SITE CHARACTERIZATION

2.3.1 Regional Geology

The Site is located in the Atlantic Coastal Plain Physiographic Province and is underlain by unconsolidated deposits of Pleistocene and late Cretaceous age. These deposits consist of gravel, sand, silt and clay and are underlain by a relatively impermeable bedrock of early Paleozoic and/or Precambrian age. The deposits form six hydrogeologic units. The units are, from bottom (oldest) to top (youngest), the Lloyd aquifer, Raritan clay, Magothy aquifer, Port Washington aquifer, Port Washington confining unit, and the Upper Glacial aquifer (Kilburn and

Krulikas, 1987). All of these units, with the exception of the Magothy aquifer, are present beneath the Site with a combined thickness of more than 500 feet (Roux, 1991a).

The water table at the Site is within the Upper Glacial aquifer. The Upper Glacial aquifer consists of two geologic units of Pleistocene and Holocene age that overlie the Port Washington confining unit (Kilburn and Krulikas, 1987). The upper Pleistocene deposits are moraine (till) composed of unsorted clay, sand, gravel, and boulders. These deposits may also consist of outwash deposits of stratified brown sand and gravel, and lacustrine and marine deposits consisting of clay, silt, and sand. The thickness of the upper Pleistocene deposits range from 10 to 380 feet. The Holocene deposits vary in thickness from 0 to 50 feet and are composed of sand, gravel, silt, and clay; organic mud, peat, loam, and shells (Kilburn and Krulikas, 1987). The upper surface of the upper glacial deposits comprise present day land surface except in areas such as the Site, where they are overlain by recent Holocene deposits and/or fill materials (Roux, 1991a).

2.3.2 Site Geology

The information presented in this section and section 2.3.4 (Site Hydrogeology) is largely taken from the Remedial Investigation (RI) report, dated April 1991 (Roux, 1991a). The moraine deposits of the Upper Glacial aquifer are estimated to be approximately 110 feet thick beneath the Site. The moraine (till) deposits at the Site consist of four lithotypes, designated Level A through Level D (Roux, 1991a).

Level A immediately underlies the Site and consists of a brown, moderately sorted fine to medium grained sand with intermittent coarse sand and gravel layers. In some areas, Level A may include non-native sediments and debris (fill). Level A ranges in thickness from 27.5 to 1.8 feet. The thickest deposits occur in the area of the storage tanks, whereas the thinnest deposits occur along the western edge near Hempstead Harbor.

Underlying Level A is Level B, which is composed of grey, moderate to well sorted, fine to medium grained sand with thin, intermittent gravel, silt, and clay layers. The bottom of Level B may consist locally of white, well sorted, medium to coarse sand. Level B is thickest in the western portion of the Site near Hempstead Harbor, where the Level is approximately 33 feet thick. Level B is either thin or totally absent in the northeast section of the Site.

Level C, underlying Level B, is a multi-colored (grey, orange, tan, and white), poorly sorted medium to coarse grained sand with many clay, silt, and gravel layers. Level C ranges

in thickness from 55 feet in the northeast section of the Site where Level B is absent, to 22 feet in the west section of the Site where Level B is thickest.

Level D is a grey, silty clay. Level D is at least 5 feet in thickness, and may not be continuous under the entire Site.

2.3.3 Regional Hydrogeology

Precipitation is the source of virtually all the fresh water on Long Island. Precipitation on the island averages 44 inches/year, and evapotranspiration of precipitation averages 21 inches/year. Practically all the precipitation that is not consumed by evapotranspiration recharges the groundwater system. Therefore, the natural groundwater re-charge rate is estimated to be about 23 inches/year (Cohen et al., 1968). The re-charge water is transmitted to the underlying aquifers by the Upper Glacial aquifer (Kilburn and Krulikas, 1987).

Groundwater in the Upper Glacial aquifer occurs under unconfined (water table) conditions. The 1980 water table map of the Upper Glacial aquifer in the northern part of the town of Oyster Bay (Kilburn and Krulikas, 1987) shows that the Site is located at a regional groundwater discharge area. The map also shows a groundwater divide to the east of the Site. In the area of Glenwood Landing, where the Site is located, groundwater moves westward from the divide to discharge in Hempstead Harbor. An upward vertical component of groundwater flow probably occurs from the deeper hydrogeologic units to Hempstead Harbor (Kilburn and Krulikas, 1987). The fact that groundwater discharges into the salt water bodies of Hempstead Harbor (and Motts Cove, as will be explained in Section 2.3.4), prevents the salt water from entering the aquifer.

Previous studies have estimated the hydraulic characteristics (hydraulic conductivity and storativity) of the Upper Glacial aquifer. The specific yield (unconfined aquifer storativity) of the Upper Glacial aquifer in the vicinity of Glenwood Landing is estimated to be 0.10 (Getzen, 1977). The average horizontal and vertical hydraulic conductivities of the Upper Glacial aquifer for Long Island are estimated to be 270 ft/day (2,000 gpd/ft²) and 27 ft/day (200 gpd/ft²) (Franke and Cohen, 1972). McClymonds and Franke (1972) determined the horizontal hydraulic conductivity values of selected lithologic classes in the Upper Glacial aquifer. When the Upper Glacial aquifer is composed of medium, fine, and very fine sand, and sand with silt or clay layers, the horizontal hydraulic conductivity ranges from 53.5 ft/day (400 gpd/ft²) to 240 ft/day (1800 gpd/ft²) (McClymonds and Franke, 1972).

2.3.4 Site Hydrogeology

The RI report shows that the water table configuration is the same under low and high tide conditions. The water table map depicting conditions during low tide on March 19, 1990 shows a shallow groundwater mound near the center of the Site. Groundwater flows radially away from the mound and discharges into Hempstead Harbor and Motts Cove. The mound is caused by the shape and permeability of the high, bermed area where the tanks are located. Precipitation cannot escape the area as surface runoff and there is little vegetation for transpiration. The permeable surface of this area will allow rapid infiltration, which in turn will cause a local mounding of the water table (Roux, 1991a).

Groundwater elevation data from the shallow wells (wells screened from approximately 10 to 20 feet below the water table) and the deep wells (wells screened from approximately 45 to 70 feet below ground surface) shows that at both low and high tide, groundwater enters the Site from the east and northeast and flows to the west and southwest, discharging into Hempstead Harbor. The horizontal hydraulic gradient (slope of the water table) varies from 0.005 under the tank area to 0.05 along the eastern perimeter. The average horizontal gradient south and west of the embankments is 0.02 (Roux, 1991a).

The principal effect of the tidal cycle on groundwater flow is that it reverses the vertical flow direction of the upper few feet of the shallow aquifer. Flow is upward at high tide and downward at low tide. Below the upper 10 to 20 feet of the aquifer, vertical flow is upward at all times (Roux, 1991a). The upward vertical component is a critical factor at the Site because it prevents surface contaminants from migrating into deeper portions of the aquifer.

No Dense Non-Aqueous Phase Liquid (DNAPL) were found in three deep wells at the Site indicates that downward migration of DNAPL has not occurred (Roux, 1991a). Light NAPL (LNAPL), such as those primarily found at the Site, will float on the water table and slowly dissolve. The dissolved components will behave in the same manner as the groundwater in which they are dissolved. At the Site, they will flow horizontally to Hempstead Harbor and Motts Cove. Mounding at the center of the Site will cause a downward flow component beneath the mound to a depth where it is overcome by the upward flow component in the aquifer.

In August 1990, Roux conducted three short-term specific capacity tests to determine the hydraulic conductivity of the shallow aquifer. The obtained hydraulic conductivity values range from 10 gallons per day per square foot (gpd/ft²) (1.3 ft/day) to 225 gpd/ft² (30 ft/day). According to Roux (1991b), these values are less than the actual values because the tested wells were affected by well losses and partial penetration. In February 1992, RETEC utilized the tidal

fluctuation technique of Ferris (1963) to determine the hydraulic conductivities of the shallow and deep aquifers. The obtained values are 57 ft/day (426 gpd/ft²) for the shallow aquifer (Level B), and 53.5 ft/day (400 gpd/ft²) for the deep aquifer (Level C). Results of the hydraulic conductivity based on the tidal fluctuation technique are discussed in section 4.3.2 of this report.

2.3.5 Meteorology

Long Island is located between 40° and 42° north latitude in a temperate-climate belt. Table 2-1 summarizes the average monthly precipitation observed at LaGuardia Airport, New York between the years 1962 and 1991. LaGuardia Airport is the closest location to the Site for which meteorological data is available. During this period, the average monthly precipitation is fairly evenly distributed throughout the year. The mean annual precipitation between 1962 and 1991 is 43.08 inches. Most of the rainfall from May through October comes from thunderstorms. It is usually of brief duration and sometimes intense. For the other months of the year, precipitation is more likely to be associated with widespread storm areas, so that daylong rain, snow or a mixture of both is more common. The maximum 24 hours precipitation event ranges from 2.9 to 7.11 inches, with an average of 3.9 inches.

The mean monthly temperature during the period 1962-1991 is 54.6° F. July is the warmest month with an average monthly temperature of 76.7° F and a maximum recorded temperature of 107°F (1966). January is the coldest month with an average temperature of 32.1° F, and a minimum recorded temperature of -3° (1985).

The prevailing wind direction is northwest during most of the year, except during the summer months when south and southwest winds prevail. Mean monthly wind speed does not vary widely, and averages 12.2 mph for the period 1949-1991 (NOAA, 1991).

2.4 CONSTITUENTS OF INTEREST

The types and concentrations of COIs detected in the soil, groundwater, sediments, and air at the Site are described in detail in the RI report and summarized below and in Section 5.0. A more complete description of these constituents of interest (COIs) and their relative distributions is also provided in the Risk Assessment portion of the Feasibility Study (Roux, 1991b).

TABLE 2-1

Average Monthly Precipitation

Month	Average Precipitation (inches)
January	3.13
February	2.95
March	3.89
April	3.71
May	3.79
June	3.28
July	4.08
August	4.34
September	3.35
October	3.17
November	3.82
December	3.57

^{*} observed at LaGuardia, New York between 1962 and 1991.

2.4.1 Soil and Sediments

The predominant chemicals found in the soils and sediments during the RI at the Site are ethylbenzene, toluene and xylenes. These three compounds comprise more than 98 percent (by mass) of the constituents identified in all samples detected. The sum of these three compounds concentrations ranged from below detection limits (BDL) up to 1 percent in one of the soil samples. Concentrations of xylenes ranged from BDL to approximately 8,400 parts per million (ppm); ethylbenzene from BDL to 1,300 ppm; and toluene from BDL to 2,600 ppm. Approximately 50 percent of the soil samples analyzed contained one or more of these constituents. In general, the highest concentrations of these three compounds were detected in shallow soil (above the water table) along the western portion of the Site (adjacent to the shoreline).

Eight halogenated compounds were detected in 10 out of 45 soil samples submitted for analysis during RI activities. TCA and methylene chloride were the most predominantly detected halogen compounds at the site.

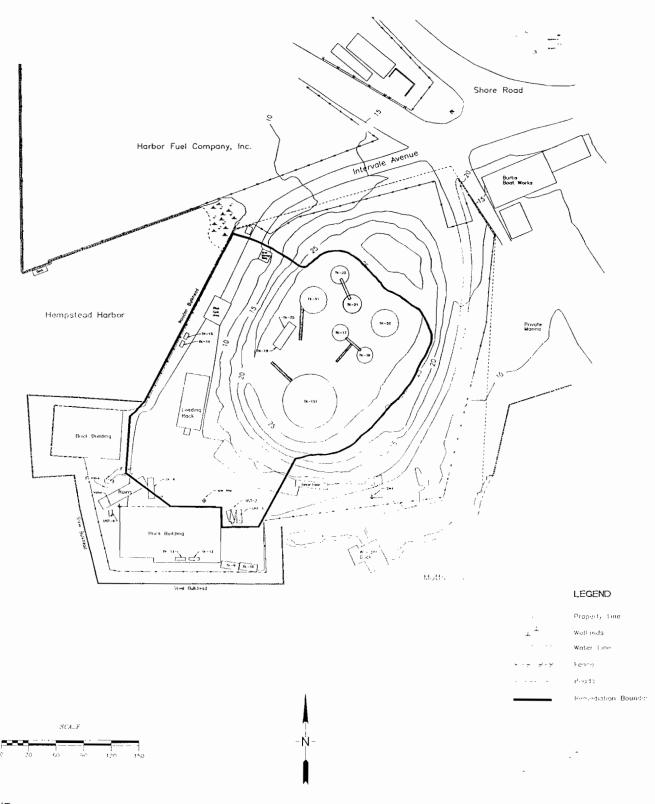
Fifteen PAHs, four phthalates, and one phenolic compound were identified during the RI in Levels A, B, C, and D. These detections were predominantly behind the bulkhead along the western edge of the site and in the tank farm area.

Sixteen PAHs, four phthalates, and no phenolic compounds were detected during the RI in sediments of Motts Cove and Hempstead Harbor adjacent to the site. The detections of these compounds were more frequent, but similar to slightly greater concentrations than the soils on the site. This is particularly true of phthalates which were detected in all eighteen sediment samples collected during the RI. The levels and distributions of organic chemicals and metals in soil and sediments are described in detail in the RI and summarized in Section 5.0. Based on the RI results, the area of soil to be remediated was defined and shown in Figure 2.3.

2.4.2 Groundwater

Six shallow and three deep groundwater monitoring wells (SW- and DW-Series) were installed and sampled as part of the RI. These wells are screened below the water table to ascertain the quality of the groundwater exclusive of any non-aqueous phase liquids or organic sheen floating on the water table. Dissolved constituents within the groundwater consist primarily of ethylbenzene, toluene and xylenes. Other volatile organic compounds that were detected in some samples include benzene, methylene chloride, 1,1-dichloroethane, trans-1,2-dichloroethylene, 1,1,1-TCA, TCE and PCE. These constituents were more prevalent in the

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Area Of Soil To Be Remediated As Identified In The RI 116(11-

2-3

groundwater than in the soils and sediments which is more indicative of an off-site source for these chemicals.

Of all the semi-volatile compounds analyzed during the RI, only phthalates were detected in any of the samples. Two phthalates were detected in five wells. However, all three of the DW series wells detected these compounds. Complete analytical results from these wells are given in the RI report and summarized in Section 5.0.

2.5 SELECTED REMEDY

Based upon the results of the RI/FS and the criteria for selecting a remedy under the applicable laws and regulations, the New York State Department of Environmental Conservation (NYSDEC) and U.S. EPA selected an integrated remedy comprised of *in situ* soil venting, extraction of groundwater and treatment by air stripping and *in situ* biodegradation and monitoring to remediate the Site. The elements of the proposed remedial program are as follows:

- A biotreatability pilot study to determine the type and amount of nutrient and oxygen additives needed to stimulate the growth of indigenous bacteria capable of biodegrading site contaminants.
- 2. A remedial design program to verify the components of the conceptual design and provide the details necessary for the construction, implementation, and monitoring of the remedial program.
- 3. Installation and operation of a soil venting (vapor extraction) system.
- 4. Installation and operation of a groundwater collection and treatment system.
- 5. **A biotreatment program** designed to reduce contaminants in the saturated soils and groundwater to the extent practicable, in conjunction with the other process options employed.
- 6. **A monitoring program** designed to evaluate both the performance of the remedial program while in operation, and its continued effectiveness after discontinuation.

3.0 OBJECTIVES

3.1 SITE REMEDIAL OBJECTIVES

The objectives to be obtained by implementing the proposed remedy as specified in the ROD are:

1. Soil

- a) Reduce the concentrations of benzene and methylene chloride so that the presence of these chemicals at the Site do not present an added risk of cancer of more than one in one million under the most conservative exposure scenario.
- b) Reduce the concentrations of organic contaminants in soils so that, to the extent feasible, contaminants do not leach from soils and contaminate groundwater to levels above standards.
- 2. **Groundwater** Reduce the concentrations of contaminants in groundwater to below New York State groundwater standards as shown in Table 3-1, to the extent technically feasible.
- 3. **Sediments** Indirectly remediate sediments by treating the source of contaminants to the sediments, site soils and groundwater.
- 4. **Air** Eliminate the exceedances of applicable ambient air standards over the mudflats adjacent to the Site.
- 5. Surface Water Eliminate the sheen on surface waters to comply with applicable surface water standards.

If monitoring indicates that continued operation of the remedy is not producing significant reductions in the concentrations of contaminants in soils and groundwater, in accordance with the NCP, the NYSDEC and the U.S.EPA will evaluate whether discontinuance of the remedy is warranted. The criteria for discontinuation will include an evaluation of the operating conditions and parameters as well as a statistical determination that the remedy has attained the feasible limit of contaminant reduction and that further reductions would therefore be impracticable.

TABLE 3-1

Potential Applicable or Relevant and Appropriate Requirements

Clean Water Act Water Quality Criteria for Saltwater Aquatic Life Acute/Chronic	224,000/~	31,200/49	1	10,200/450	2,000/40	12,000/6,400	1	-	113,000/4	224,000/4	5,100/700%	6,300/5,0000	1	430'-0	•••	***	€,800/-0
Clea Water (for Sal Life /						151						6,		,		1:1	
New York State Ambient Water Quality Standards and Guidance Values Part 703 Title 6	5 2	5	\$	\$	5	7	2	5	\$	5	0.7	\$	(p)-	5	\$	5	1(0)
Clean Water Act Water Quality Criteria for Human Health - Fish Consumption	1.85	1,030,000	***	58.8	80.7	15.7	525		ï	:	40	424,000	**	3,250			;
Clean Water Act Water Quality Criteria for Freshwater Aquatic Life Acute/Chronic	11,600/-(19			5,280/840 ^{to}	45,000/21,900 ^(b)	28,900/1,240 ^(b)	***		11,800/20,000 ^(b)	11,600/-69	5,300/- ^(b)	17,500/(6)	:	32,000/-(b)			10.200/2,560 ^(h)
Safe Drinking Water Act Maximum Contaminant Level Goals (MCLGs) 40 CFR 141 & 50 FR 46936	7	200	100	0	0		0				0	1.000	:	700	10.000	100	:
Safe Drinking Water Act Maximum Contaminant Levels (MCLs) 40 CFR 141	7	200	100	5	5	1004	2		1		5	1.000	1	700	10,000	100	;
Constituents of Interest	1.1 dichloroethylene	1.1.1-trichloroethane	Trans-1.2-dichloroethylene	Tetrachloroethylene	Trichloroethylene	Chloroform	Vinyl Chloride	Methylene Chloride	1,1 dichloroethane	1,2 dichloroethylene	Benzene	Toluene	Acetone	Ethylbenzene	Xylene (total)	Styrene	louetd .

--- indicates no specific criteria has been established. All values in pg/l unless otherwise noted.

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TABLE 3-1 (continued)

Potentially Applicable or Relevant and Appropriate Requirements

Constituents of Interest	Safe Drinking Water Act Maximum Contaminant Levels (MCLs) 40 CFR 141	Safe Drinking Water Act Maximum Contaminant Level Goals (MCLGs) 40 CFR 141 & 50 FR 46936	Clean Water Act Water Quality Criteria for Freshwater Aquatic Life	Clean Water Act Water Quality Criteria for Human Health - Fish Consumption	New York State Ambient Water Quality Standards and Guidance \ anter	Clean Water Act Water Quality Criteria for Saltwater Aquatic Life Acute/Chronic	
Naphthalene	-		2,300/620	-	10	2,350/4	
2-methylnaphthalene				••	(p)-	1	
Dibutyl phthalate				154,000	50	!	
Benzoic Acid	-		***		(p)-		
Benzyl Alcohol	1	**	***	•••	(p)-	**	
Di-n-octylphthalate			•••		50	1	
Isophorone	-		117,000/-(b)	520,000	80	12,900/49	
Di Benzo Furan	•••	*	1	•	(p)-		
bis (2-ethylbexyl)phthalate	:	i	1	50,000 ^(c)	1	**	
Acenapthylene	1	:	1	1	(p)-	#	
Acenaphthene	:	:	1,700/5265	1	20	970/500	
Phenanthrene	:	;	1	1	80	ı	
Fluoranthene		;	3,980/-(1)	54	80	40/16 th	
Chrysene	:	;	:	:	.002	*	
2-methylphenol (o-cresol)	***	:	!	!	1	-	
4-methylphenol (p-cresol)	:	:	;		1	1	
2,4 dimethylphenol (xylend)	1	:	2.120/ (1)	400	90	•	

--- indicates no specific criteria has been established. All values in $\mu g/l$ unless otherwise noted.

Potentially Applicable or Relevant and Appropriate Requirements

Constituents of Interest	Safe Drinking Water Act Maximum Contaminant Levels (MCLS) 40 CFR 141	Safe Drinking Water Act Maximum Contaminant Level Goals (MCLGs) 40 CFR 141 & 50 FR 46936	Clean Water Act Water Quality Criteria for Freshwater Aquatic Life Acute/Chronic	Clean Water Act Water Quality Criteria for Human Health - Fish Consumption	New York State Ambient Water Quality Standards and Guidance Values Part 703 Title 6	Clean Water Act Water Quality Criteria for Saltwater Aquatic Life Acute/Chromic
Anthracene			***		90	1
Benzo (a) Anthracene		:	•••		2003	:
Fluorene			•	:	50	
Pyrene		***	***		90	
2.4 Dinitrophenol				14.3	•••	4,850/%
4.6 Dinitro-2-methylphenol				765	***	4,850/-0
N-Nitrosodipropylamine		:	!	!	ı	I

MCL for total trihalomethane concentration

--- indicates no specific criteria has been established. All values in $\mu g/l$ unless otherwise noted.

TBL03-01

^{@ @ @ @ @}

Lowest observed effect level
Value shown is for di-2-ethylhexyiphthalate
Not regulated by State of New York under POC Groundwater standards
Total phenols

3.2 APPLICABLE OR RELEVANT AND APPROPRIATE REQUIREMENTS

Under the Superfund Amendments and Reauthorization Act (SARA) of 1986, remedial actions must comply with Applicable or Relevant and Appropriate Regs (ARARs) unless one or more of six waiver conditions are met (CERCLA Section 121[d][2][A], [d][4]). Applicable requirements are:

Those cleanup standards, standards of control, and other substantive requirements, criteria, or limitations, promulgated under federal or state environmental facility listing laws that specifically address a hazardous substance, pollutant, contaminant, remedial action, location, or other circumstances found at a CERCLA site.

(40 CFR Section 300.5 at 55 Fed. Reg. 8814, USEPA 1990b)

Relevant and appropriate requirements are:

Those cleanup standards, standards of control, and other substantive requirements, criteria, or limitations promulgated under federal, or state environmental or facility listing laws that, while not "applicable" to a hazardous substance, pollutant, contaminant, remedial action, location, or other circumstance at a CERCLA site, address problems or situations sufficiently similar to those encountered at the CERCLA site that their use is well suited to the particular site. Only those state standards that are identified in a timely manner and are more stringent than federal requirements may be relevant and appropriate.

(40 CFR Section 300.5 at 55 Fed. Reg. 8817, USEPA 1990b)

The remedy will continue until such time that compliance with the substantive technical requirements of the ARARs listed in Tables 3-2 and 3-3 or conditions indicate that a waiver of these ARARs is justified based upon conditions given in the ROD.

TABLE 3-2

Listing of Potential Federal ARARs and TBCs

WATER

Safe Drinking Water Act [42 U.S.C. 300(f)]

40 CFR 141.11-16 Maximum Contaminant Levels 40 CFR 141.50-52 Maximum Contaminant Level Goals 40 CFR 144-147 **Underground Injection Control Regulations** 40 CFR 122-125 National Pollutant Discharge Elimination System

40 CFR 403 Pretreatment Standards

Clean Water Act (33 U.S.C. 1251)

40 CFR 230 Guidelines for Specification of Disposal Sites for Dredged or Fill Materials

40 CFR 231 Restriction of Disposal Si 's for Dredged Materials

40 CFR 131 Water Quality Criteria

Rivers and Harbors Act

Dredge and Fill Requirements Section 10

"Quality Control for Water, 1986" - EPA 44/5-86-001, May 1, 1986, 51 FR 43665

Health Advisories, EPA Office of Water

*Developing Requirements for Direct and Indirect Discharge of CERCLA Wastewaters, 1987" - USEPA Office of Water Guidance **Documents**

AIR

Clean Air Act (42 U.S.C. 7401)

40 CFR 50 National Primary and Secondary Ambient Air Quality Standards 40 CFR 61 National Emissions Standards for Hazardous Air Pollutants 40 CFR 60 New Source Performance Standards

HAZARDOUS WASTE

Resource Conservation and Recovery Act

Identification and Listing of Hazardous Wastes 40 CFR 264.18 Location Standards and Prohibitions for TSD Facilities 40 CFR 264.90-109 Groundwater Protection and Monitoring 40 CFR 264.110-120 Closure and Post-closure 40 CFR 264.170-176 Containers 40 CFR 264.190-199 Tanks 40 CFR 264.270-299 Land Treatment Landfills 40 CFR 264.300-339 40 CFR 264.340-999 Incinerators

40 CFR 268.1-50 Land Disposal Restrictions

40 CFR 264 Subpart S Corrective Action at Hazardous Waste Management Facilities (Proposed)

USEPA RCRA Guidance Documents - Design Guidelines

Land Treatment Units

40 CFR 264

Landfill Design

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USEPA Technical Resource Documents

Hazardous Waste Land Treatment

Review of In-Place Treatment Technologies for Contaminated Surface Soils, Vol.2, USEPA-540/2-84-0036, November 1984.

Department of Transportation

49 CFR 107, 171, 172 Hazardous Materials Transport

Toxic Substances Control Act (15 U.S.C. 2601)

40 CFR 761.60-79 Storage and Disposal of PCBs PCB Spill Clean-up Policy Rule 40 CFR 761.120

TABLE 3-2 (continued)

Listing of Potential Federal ARARs and TBCs

MISCELLANEOUS

Coastal Zone Management Act of 1972 (16 U.S.C. 1451)

15 CFR 930, 923.45

Air and Water Pollution Control Requirements

Endangered Species Act of 1973 (16 U.S.C. 1531)

50 CFR 81, 225, 402

Fish and Wildlife Coordination Act (16 U.S.C. 661)

Marine Protection, Research and Sanctuaries Act (33 U.S.C. 1401)

Occupational Safety and Health Act (29 U.S.C. 651)

29 CFR 1910

Requirements for Workers Engaged in Response Activities

Integrated Risk Information System (IRIS), USEPA 1990

Carcinogenic Potency Factors (CPF)

Reference Doses for Chronic Exposure (RfD)

Health Effects Assessments (HEAs), USEPA 1985

Executive Orders 11988 (Floodplains) and 11990 (Wetlands)

U.S. EPA's Policy on Floodplains and Wetlands Assessment: for CERCLA Actions, August 6, 1985, (40 CFR 6, Appendix A)

1033PDIR/TBL03-02 June 8, 1993

TABLE 3-3

Listing of Potential New York State ARARs/SCGs and TBCs

	WATER							
6 NYCRR 701	Classifications and Standards of Quality and Purity, and Appendix 31							
6 NYCRR 701.15	Derivation of Effluent Limitations; empowers State to enforce guidance values for surface							
6 NYCRR 702	water where no standards exist Special Classifications and Standards							
6 NYCRR 703	Groundwater Classifications, Quality Standards and Effluent Standards and/or Limitations							
6 NYCRR 750-757	Implementation of NPDES Program in NYS							
6 NYCRR 885	Classifies Hempstead Harbor Class SB Waters							
10 NYCRR 5	Public Water Supply MCLs							
10 NYCRR 170	Water Supply Sources							
TOGS 1.1.1 (9/24/90) TOGS 2.1.2 (4/1/88)	Ambient Water Quality Standards and Guidance Values Underground Injection/Recirculation (UIR) at Groundwater Remediation Sites							
1003 2.1.2 (4/1/88)	Onderground injection Recirculation (Olik) at Groundwater Remediation Sites							
AIR								
6 NYCRR 257 Air Quality Standards								
6 NYCRR 212	General Process Emission Sources							
Air Clean-up Criterion, January 19	990, Ambient Guideline Concentrations							
	HAZARDOUS WASTE							
6 NYCRR 371	Identification and Listing of Hazardous Waste							
6 NYCRR 372	Hazardous Waste Manifest System and Related Standards							
6 NYCRR 373	Location and Design Standards for TSD Facilities							
6 NYCRR 373-2	Final Status Standards for Owners and Operators of Hazardous Waste							
CNIVORD 274	Treatment/Storage/Disposal Facilities							
6 NYCRR 374	Standards for the Management of Specific Hazardous Wastes and Specific Types of Hazardous Waste Management Facilities							
MISCELLANEOUS								
Department of State Coastal Manag State Coastal Policies	gement Program							
Division of Marine Resource 6 NYCRR 661	Chapter 10 Tidal Wetlands, Land Use Regulations							
Division of Fish and Wildlife 6 NYCRR 182	Endangered Species of Fish and Wildlife							
Sediment Criteria (NYS 1989)								

4.0 SCOPE OF PRE-DESIGN INVESTIGATION

This section is the summary of site activities performed for the PDI from February 22 through March 12, June 7 and June 22, 1993. During this period, on-site activities included soil borings, soil gas survey, well points, monitoring well installation, time-lag stage-ratio study, and groundwater sampling. These activities are described below.

4.1 SOIL EXPLORATION

The soil explorations were conducted using a CME 550 drilling rig with 4-1/4" I.D. hollow stem augers. The drilling contractor was Warren George, Inc. of Jersey City, New Jersey.

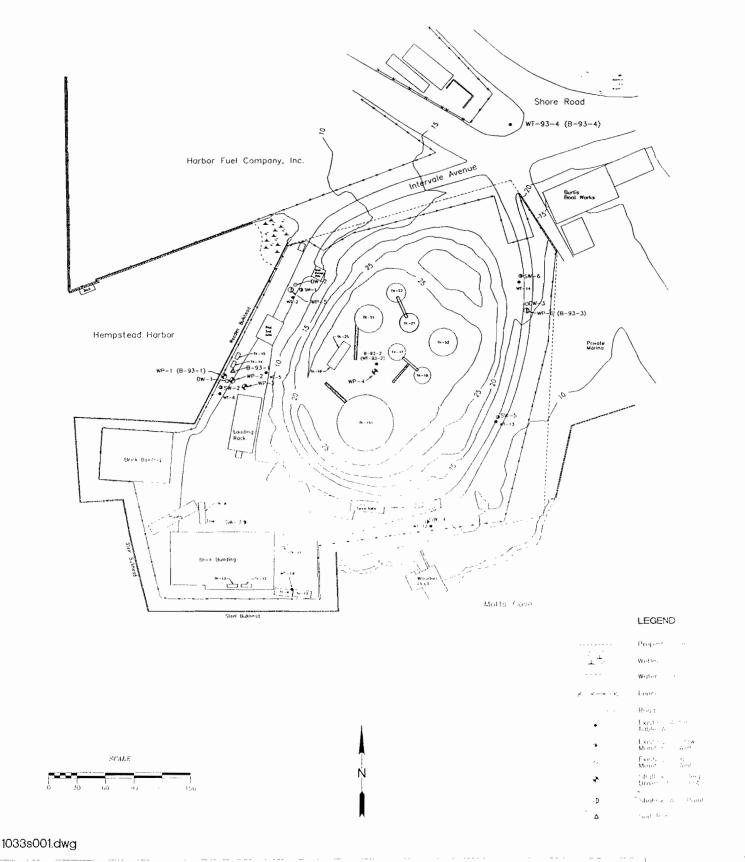
4.1.1 Soil Borings

Soil borings were conducted in 3 locations on-site and one location off-site, as shown in Figure 4-1. Logs are presented in Appendices A and C. The first boring (B-93-2), located in the center of the tank farm area, was drilled to 52 feet and 4" diameter split spoons were collected approximately every 5 feet. Soils collected were screened using an OVA, and 9 soil samples were collected for analysis. See Table 4-1. Samples for vertical hydraulic conductivity were attempted, but the sandy soil resulted in no suitable sample recovery. Soil samples for the iron precipitation study were collected from 31' to 35' using 4" diameter split spoons lined with plastic sleeves. The boring was then completed as monitoring well WT-93-2.

The second boring (B-93-3) was located in the northeast corner of the Site near the Burtis Boat Works. The boring was advanced to 44 feet and 4" diameter spilt spoons were collected approximately every 5 feet. Soil collected was screened using an OVA, and 9 soil samples were collected for analysis. Samples for vertical hydraulic conductivity were collected using 4-inch split spoons with plastic liners at depths 22'-24' and 47'-49'. The boring was then completed as well point WP-6.

The third boring was located along Hempstead Harbor adjacent to the truck loading rack. The boring was advanced to 14 feet and split spoons were driven continuously. Soil collected was screened using the OVA and 9 soil samples were collected for analyses. The boring was

RELEC



PDI Investigation Locations

51110

TABLE 4-1

Summary of Headspace Readings Shore Realty Superfund Site Glenwood Landing, New York

W 34 C C 8 8 7 C C C C C C C C C C C C C C C			OVA Reading
Boring I.D.	Sample I.D.	Depth	(ppm)
B-93-1	B-93-1A,B,C	0'-2'	0
		2'-4'	12
	B-93-1D,E,F	4'-6'	72
		6'-8'	>1,000
	B-93-1G,H,I	'8'-10'	350
		10'-12'	315
		13'-15'	275
B-93-2		0'-2'	0
D - 7.7 - 2		5'-7'	0
		10'-12'	2
	B-93-2A,B,C	12'-14'	200
	B-93-2C,D,E	15'-17'	>1,000
	B-93-2F,G,H	19'-21'	65
		33'-35'	10
		40'-42'	5
		45'-47'	0.5
B - 93 - 3	B-93-3A,B,C	3'-5'	0
	B-93-3D,E,F	7'-9'	2.5
	B-93-3G,H,I	11'-13'	N/A
		15'-17'	4
		25'-27'	4
		30'-32'	2
		35'-37'	8
		40'-42'	6.5
WΓ-93-4		0'-2'	0
W 1-93-4		2'-4'	0
		2 - 4 4' - 6'	0
		6'-8'	0
		8'-10'	0
		10'-12'	0
		12'-14'	0
		14'-16'	0
		16'-18'	0

then completed as well point WP-1. The off-site boring was located on the northern side of Intervale Avenue adjacent to the Harbor Fuels Facility. The boring was advanced to 18.5 feet and split spoons were driven continuously. Soil collected was screened with an OVA. No samples were sent for chemical analysis. The boring was completed as WT-93-4.

4.1.2 Soil Samples

Soil samples, identified in Table 4-1, were collected from each of the three borings. Sampling tools were decontaminated after each use with soap, water and methanol. Each sample involved splitting the contents of the split spoon into three or four subsamples. The soil was divided into four sub-samples for the following analyses: individual constituents, total petroleum hydrocarbons, bio-treatability, and field screening. After the samples were collected, they were stored in coolers and depending on the number of samples collected, they were delivered or picked up by the analytical laboratory (NYTEST Environmental Inc.) daily. If the samples were not picked up, they were stored in coolers packed with ice. The chain of custody used to track sample shipments is presented in Appendix B. The results of the volatile organic chemical analyses are presented in Table 4-2. Total CVOCs in Table 4-2 include:

- methylene chloride;
- 1,1-dichloroethylene;
- 1,1-dichloroethane;
- 1,2-dichloroethylene (total);
- 1,2-dichloroethane;
- chloroform;
- 1,1,1-TCA;
- TCE;
- 1,1,2-TCA;
- PCE; and
- 1,1,2,2-PCE.

The results of the SVOCs analyses are presented in Table 4-3. The results of the inorganic chemical analyses are presented in Table 4-4.

Table 4-2Soil Samples for TCL Volatile Organics (mg/kg)

Shore Realty Superfund Site Glenwood Landing, New York

Sample ID		932A		-93-2D	-	93-2G		-93-3A		93-3D		-93-3G		93-1A -		-93-1D	7.000	93-1G
Laboratory] ID	15	89701	1	1589702	15	89703	1.5	592201	15	92202	15	92205	15	97701		1597702	159	7703
% Moisture		7		19		18		9		22		20		21	'	16		20
Diluiton Factor		1.00		1		√1		1.00		1		1		1.00		10	L	1.00
Chloromethane	<	0.011 U	<	1.5 U	<	0.012 U	<	0.011 U	<	0.013 U	<	0.012 U	<	1.6 U	<	15 UJ	<	1.6 U
Bromomethane	<	0.011 U	<	1.5 U	<	0.012 U	<	0.011 U	<	0.013 U	<	0.012 U	<	1.6 U	١ <	15 UJ	<	1.6 L
Vinyl Chloride	<	0.011 U	<	1.5 U	<	0.012 U	<	0.011 U	<	0.013 U	<	0.012 U	<	1.6 U	<	15 UJ	<	1.6 U
Chloroethane	<	0.011 U	<	1.5 U	<	0.012 U	<	0.011 U	<	0.013 U	<	0.012 U	<	1.6 U	<	15 UJ	<	1.6 T
Methylene Chloride	\ <	0.008 U	<	11 U	<	0.016 U	<	0.011 U	<	0.014 U	<	0.013 U	٠, ۲	3.7 U	<	16 UJ	<	2.3 t
Acetone	1	0.1	<	1.5 U	<	0.012 U	<	0.011 U	<	0.017 U	<	0.027 U	<	1.6 UJ	<	15 UJ	<	1.6 t
Carbon Disulfide	<	0.005 U	<	0.77 U	<	0.006 U	<	0.005 U	<	0.006 U	<	0.006 U	<	0.79 U	<	7.4 UJ	<	0.78 T
1,1-Dichloroethylene	<	0.005 U	<	0.77 U	<	0.006 U	<	0.005 U	<	0.006 U	<	0.006 U	<	0.79 U	<	7.4 UJ	<	0.78 t
1,1-Dichloroethane	<	0.005 U	\ <	0.77 U	<	0.006 U	<	0.005 U	<	0.006 U	<	0.006 U	<	0.79 U	<	7.4 UJ	<	0.78 T
1,2-Dichloroethylene (total)	<	0.005 U	<	0.77 U	<	0.006 U	<	0.005 U	<	0.006 U	<	0.006 U	<	0.79 U	<	7.4 UJ	<	0.78 T
Chloroform	\ <	0.005 U	<	0.77 U	<	0.006 U	<	0.005 U	<	0.006 U	<	0.006 U	<	0.79 U	<	7.4 UJ	<	0.78
1,2-Dichloroethane	<	0.005 U	<	0.77 U	<	0.006 U	<	0.005 U	<	0.006 U	<	0.006 U	<	0.79 U	<	7.4 UJ	<	0.78
2-Butanone		0.023	<	1.5 U	<	0.012 U	<	0.011 U	<	0.013 U	<	0.012 U	<	1.6 U	<	15 UJ	<	1.6 U
1.1.1-Trichloroethane	<	0.005 U	<	0.77 U	<	0.006 U	<	0.005 U	<	0.006 U	<	0.006 U	<	0.79 U	<	7.4 UJ	<	0.78
Carbon Tetrachloride	<	0.005 U	\ <	0.77 U	<	0.006 U	<	0.005 U	<	0.006 U	<	0.006 U	<	0.79 U	<	7.4 UJ	<	0.78
Vinyl Acetate	<	0.011 U	<	1.5 U	<	0.012 U	<	0.011 U	<	0.013 U	<	0.012 U	<	1.6 UJ	<	15 UJ	<	1.6 1
Bromodichloromethane	<	0.005 U	<	0.77 U	<	0.006 U	<	0.005 U	<	0.006 U	<	0.006 U	<	0.79 U	<	7.4 UJ	<	0.78 U
1,2-Dichloropropane	\ <	0.005 U	<	0.77 U	<	0.006 U	<	0.005 U	<	0.006 U	<	0.006 U	<	0.79 U	<	7.4 UJ	<	0.78 T
cis-1,3-Dichloropropene	<	0.005 U	<	0.77 U	<	0.006 U	<	0.005 U	<	0.006 U	<	0.006 U	<	0.79 U	<	7.4 UJ	<	0.78
Trichloroethylene	<	0.005 U	<	0.77 U	<	0.006 U	<	0.005 U	<	0.006 U	<	0.006 U	 <	0.79 U	<	7.4 UJ	<	0.78
Dibromochloromethane	<	0.005 U	\<	0.77 U	<	0.006 U	<	0.005 U	<	0.006 U	<	0.006 U	<	0.79 U	<	7.4 UJ	<	0.78
1,1,2-Trichloroethane	<	0.005 U	<	0.77 U	<	0.006 U	<	0.005 U	<	0.006 U	<	0.006 U	<	0.79 U	<	7.4 UJ	<	0.78
Benzene	<	0.005 U	<	0.77 U	<	0.006 U	<	0.005 U		0.006 J	1	0.004 J	1	28 J	<	7.4 UJ	<	0.78 1
Trans-1,3-Dichloropropene	<	0.005 U	٧.	0.77 U	<	0.006 U	<	0.005 U	<	0.006 U	<	0.006 U	<	0.79 U	<	7.4 UJ	<	0.78
Bromoform	<	0.005 U	<	0.77 U	<	0.006 U	<	0.005 U	<	0.006 U	<	0.006 U	<	0.79 U	<	7.4 UJ	<	0.78
4-Methyl-2-Pentanone	<	0.011 U	<	1.5 U	<	0.012 U	<	0.011 U	<	0.013 U	<	0.012 U	<	1.6 U	<	15 UJ	<	1.6
2-Hexanone	<	0.011 U	<	1.5 U	<	0.012 U	<	0.011 U	<	0.013 U	<	0.012 U	<	1.6 U	<	15 UJ	<	1.6 T
Tetrachloroethylene	<	0.005 U	<	0.77 U	<	0.006 U	<	0.005 U	<	0.006 U	<	0.006 U	<	0.79 U	<	7.4 UJ	<	0.78
1.1.2.2-Tetrachloroethane	<	0.005 U	<	0.77 U	<	0.006 U	<	0.005 U	<	0.006 U	<	0.006 U	<	0.79 UJ	<	7.4 UJ	<	0.78 1
Toluene	<	0.005 U		37 D	1	0.047	<	0.005 U	1	0.003 J		0.002 J	1	0.15 J		720 J	1	23
Chlorobenzene	<	0.005 U	<	0.77 U	<	0.006 U	<	0.005 U	<	0.006 U	<	0.006 U	<	0. 7 9 U	<	7.4 UJ	<	0.78
Ethylbenzene	<	0.005 U		61 D		0.1	<	0.005 U	\ <	0.006 U	<	0.006 U		1.8 J		260 J		6.8
Styrene	<	0.005 U	<	0.77 U	<	0.006 U	<	0.005 U	١,	0.006 U	<	0.006 U	<	0.79 U	<	7.4 UJ	<	0.78
Total Xylenes		0.007		530 D		0.6	<	0.005 U	<	0.006 U	<	0.006 U	L	27 J		1600 J		41
Total CVOCs		0		0		0		0		<u> </u>		0		0	ζ			0
Total BTEX		0.007		628		0.747		0		0		6	1	56.95		2580		70.8

U means the material was analyzed for, but not detected.

J means the associated numerical value is an estimated quantity.

UJ means the material was analyzed for, but not detected. The sample quantitation limit is an estimated value.

D means there was an analysis at a secondary dilution factor.

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Table 4-3 Soil Samples for TCL Semivolatile Organics (mg/kg)

Shore Realty Superfund Site Glenwood Landing, New York

Sample II)	B-93-2B	B-93-2E	B-93-2H	B-93-3B	B-93-3E	B-93-3H	B-93-1B	B-93-1B	B-93-1H
Laboratory ID	1589704	1589705	1589706	1592206	1592207	1592210	1597706	1597707	1597708
% Moisture	5	18	17	7	20	17	15 -	20	18
Dilution Factor	1.00	1.00	1.00	1.00	2.00	2.00	4.00	4.00	1.00
Polynucicar Aromatic Hydrocarbo									
bis(2-Chloroethyl)ether	< 0.035 U	< 0.4 U	< 400 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
1,3-Dichlorobenzene	- 0.035 U	< 0.4 U	■ 0.4 U	< 0.35 U	< 0.82 U	4 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
1,4-Dichlorobenzene	• 0.035 U	4 0.4 U	→ 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	4 1.6 U	< 0.4 U
Benzyl Alcohol	→ 0.035 U	< 0.4 U	< 0.4 ₹₹	< 0.35 U	< 0.82 U	< 0.8 U	4 1.6 U	4 1.6 U	4 0.4 U
1,2-Dichlorobenzene	< 0.035 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	4 1.6 U	< 1.6 U	< 0.4 U
bis(2-Chloroisopropyl)ether	< 0.035 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
N-Nitroso-di-n-propylamine	< 0.035 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
Hexachloroethane	< 0.035 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
Nitrobenzene	< 0.035 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
Isophorone	< 0.035 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
Benzoic Acid	< 1.7 U	< 2 U	< 1.9 U	< 1.7 U	4 U	< 3.9 U	< 7.5 U	< 8 U	< 2 U
bis(2-Chloroethoxy)methane	< 0.035 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
1,2,4-Trichlorobenzene	< 0.035 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
Naphthalene	< 0.035 U	1.1	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	2.2	0.17 J
4 - Chloroaniline	< 0.035 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
Hexachlorobutadiene	< 0.035 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
2-Methylnaphthalene	< 0.035 U	2	0.031 J	< 0.35 U	< 0.82 U	0.008 J	0.13 J	1.8	0.39 J
Hexachlorocyclopentadiene	< 0.035 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
2-Chloronaphthalene	< 0.035 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
2-Nitroaniline	< 1.7 U	< 2 U	< 1.9 U	< 1.7 U	4 U	< 3.9 U	< 7.5 U	4 8 U	< 2 U
Acenaphthylene	< 0.035 U	0.069 J	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	0.034 J	< 1.6 U	< 0.4 U
2,6-Dinitrotoluene	< 0.035 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
3-Nitroaniline	< 1.7 U	4 2 U	< 1.9 U	< 1.7 U	4 U	< 3.9 U	< 7.5 U	< 8 U	4 2 U
Acenaphthene	< 0.035 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	0.016 J
Dibenzofuran	< 0.035 U	0.12 J	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	0.04 J	0. 01 9 J
2,4-Dinitrotoluene	< 0.035 U	< 0.4 U	< 0.4 U	< 0.35 U │	< 0.82 U	< 0.8 U ↓	< 1.6 U	< 1.6 U	< 0.4 U
4-Chlorophenyl-phenylether	< 0.035 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U │	< 1.6 U	< 1.6 U	< 0.4 U
Flourene	< 0.035 U	0.26 J	< 0.4 U	< 0.35 U	< 0.82 U \	< 0.8 U	0.041 J	0.17 J	0.032 J
4-Nitroaniline	4 1.7 U	4 2 U	4 1.9 U	< 1.7 U	4 U	< 3.9 U	▼ 7.5 U	< 8 U	< 2 U
N-Nitrosodiphenylamine	< 0.035 U	< 0.4 U	< 0.4 U	→ 0.35 U	< 0.82 U	< 0.8 U	0.059 J	0.59 J	< 0.4 U
4-Bromophenyl-phenylether	< 0.035 U │	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
Hexachlorobenzene	< 0.035 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
Pentachlorophenol	< 1.7 U	< 2 U	4 1.9 U	< 1.7 U	< 4 U	< 3.9 U	< 7.5 U	< 8 U	* 2 U
Phenanthrene	< 0.035 U	0.25 J	0.016 J	0.079 J	0.1 J	0.35 J	0.11 J	0.41 J	0.035 J
Anthracene	< 0.035 U	< 0.4 U	0.002 J	0.02 J	0.02 J	0.08 J	0.034 J	0.1 J	0.004 J
Flouranthene	< 0.035 U	< 0.4 U	< 0.4 U	0.21 J	0.17 J	0.5 J	< 1.6 U	0.3 J	< 0.4 U
Pyrene	< 0.035 U	< 0.4 U	< 0.4 U	0.15 J	0.92 J	0.29 J	0.12 J	0.29 J	< 0.4 U
3-3'-Dichlorobenzidine	< 0.069 U	< 0.81 U	< 0.8 U	< 0.71 U	< 1.6 U	< 1.6 U	< 3.1 U	< 3.3 U	< 0.81 U
Benzo (a) anthracene	< 0.035 U	< 0.4 U	< 0.4 U	0.12 J	< 0.82 U	0.24 J	< 1.6 U	< 1.6 U	< 0.4 U
Chrysene	< 0.035 U	< 0.4 U	< 0.4 U	0.11 J	< 0.82 U	0.22 J	< 1.6 U	< 1.6 U	< 0.4 U
Benzo(b)flouranthene	< 0.035 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
Benzo(k)flouranthene	< 0.035 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
Benzo(a)pyrene	< 0.035 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
Indeno(1,2,3-cd)pyrene	< 0.035 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	4 1.6 U	< 0.4 U
Dibenz(a,h)anthracene	< 0.035 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
Benzo(g,h,i)perylene	< 0.035 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U _	< 0.4 U

PDIR(TAB4-3,WK1 15-Jun 95

Table 4-3 (continued) Soil Samples for TCL Semivolatile Organics (mg/Kg)

Shore Realty Superfund Site Glenwood Landing, New York

B-93-2B B-93-2H B-93-3B B-93-3E B-93-3H B-93-1B B-93-1B B-93-1B

Laboratory ID	1589704	1589705	1589706	1592206	1592207	1592210	1597706	. 1597707	1597708
% Moisture	5	18	17	7	20	17	15 : [20	18
Dilution Factor	1.00	1.00	1.00	1.00	2.00	2.00	4.00	4.00	1.00
Phthalates									
Dimethylphthalate	< 0.35 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
Diethylphthalate	< 0.35 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
Di-n-butylphthalate	< 0.35 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	4 1.6 U	< 1.6 U	< 0.4 U
Butylbenzylphthalate	< 0.35 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	4 1.6 U	< 1.6 U	< 0.4 U
bis(2-Ethylhexyl)phthalate	0.38	0.27 J	0.042 J	< 0.12 U	< 0.16 U	< 0.59 U	0.75 J	3.9	0.16 J
Di-n-octylphthalate	0.041 J	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	0.013 J	< 1.6 U	< 1.6 U	< 0.4 U
Phonois									
Phenol	< 0.35 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
2-Chlorophenol	< 0.35 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
2-Methylphenol	< 0.35 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	4 1.6 U	< 1.6 U	0.045 J
4-Methylphenol	< 0.35 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	0.063 J
2-Nitrophenol	< 0.35 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
2,4-Dimethylphenol	< 0.35 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
2,4-Dichlorophenol	< 0.35 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
4-Chloro-3-methylphenol	< 0.35 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
2,4,6-Trichlorophenol	< 0.35 U	< 0.4 U	< 0.4 U	< 0.35 U	< 0.82 U	< 0.8 U	< 1.6 U	< 1.6 U	< 0.4 U
2.4.5-Trichlorophenol	< 1.7 U	< 2 U	< 1.9 U	< 1.7 U	< 4 U	< 3.9 U	< 7.5 U	< 8 U	4 2 U
2,4-Dinitrophenol	< 1.7 U	< 2 U	< 1.9 U	< 1.7 U	< 4 U	< 3.9 U	< 7.5 U	< 8 U	< 2 U
4-Nitrophenol	< 1.7 U	< 2 U	< 1.9 U	< 1.7 U	4 4 U	< 3.9 U	< 7.5 U	< 8 U	< 2 U
4,6-Dinitro-2-methylphenol	< 1.7 U	< 2 U	< 1.9 U	< 1.7 U	< 4 U	< 3.9 U	< 7.5 U │	< 8 U	4 2 U
Pentachlorophenol	< 1.7 U	< 2 U	< 1.9 U	< 1.7 U	< 4 U	< 3.9 U	< 7.5 U	< 8 U	< 2 U

Notes:

U means the constituent was not detected.

J means constituent detection below quantifiable detection limit.

E means constituent detection above quantifiable detection limit.

ND means constituent not detected.

NA means constituent was not analyzed for.

15 - Jun - 93

Table 4-4 Soil Samples for Inorganics (mg/kg) Shore Realty Superfund Site Glenwood Landing, New York

Sample ID	B-932B	B-932E	B-932H	B932KB	B932KT	B-933B	B-933B	B-933H	B93~1B	B93-1E	B93-1H
Laboratory ID	589704	589705	589706	589708	589707	592206	592207	592210	597706	597707	597708
Aluminum	NA	NA NA	NA	NA							
Antimony	NA										
Arsenic	NA	NA	NA NA	NA	NA	NA	NA	NA	NA	NA	NA
Berium	NA NA	NA	NA NA	NA	NA NA	NA NA	NA NA	NA NA	NA	NA NA	NA
Beryllium	NA	NA	NA	NA	NA)	NA	NA	NA NA	NA	NA NA	NA
Cadmium	NA NA	NA	NA NA	NA	NA	NA	NA	NA	NA NA	NA NA	NA
Calcium	NA	NA	NA	, NA	NA	NA	NA	NA	NA NA	NA	NA
Chromium	NA	NA	NA	NA NA	NA	NA	NA	NA NA	NA	NA	NA
Cobalt	NA	NA	NA NA	NA NA	NA	NA NA	NA	NA	NA	NA	NA
Copper	NA	NA NA	NA	NA	NA	NA	NA	NA NA	NA	NA	NA
Iron	3.23	2.21	1.34	0.307	1.91	4.4	4.05	5.27	2.78	5.77	3.81
Lead	NA	NA	NA) NA	NA NA	NA	NA	NA	NA NA	NA	NA
Magnesium	NA	NA NA	NA	NA							
Manganese	NA	NA	NA	NA	NA NA	NA	NA NA	NA	NA	NA	NA NA
Mercury	NA	NA	NA	NA	NA	NA	NA NA	NA	NA	NA	NA
Nickel	NA NA	NA	NA NA	NA NA	NA	NA NA	NA	NA	NA	NA	NA NA
Potassium	NA NA	NA	NA NA	NA NA	NA	NA NA	NA	NA	NA	NA	NA
Selenium	NA										
Silver	NA										
Sodium	NA	NA	NA NA	NA	NA	NA	NA	NA	NA	NA	NA
Thallium	NA										
Vanadium	NA										
Zinc	NA										
Cyanide	NA										
	1.50					0.47		(05)			
Ammonia Nitrogen	4.56	< 1	< 1	NA	NA NA	8.47	5.54	6.05	< 1	< 1	< 1
Total Nitrogen	42.5	< 2	18.5	NA	NA NA	37	52	111	72.5	< 2	< 2
Orthophosphate	< 0.8	< 0.8	< 0.8	NA	NA NA	1.38	1.18	3.41	< 0.8	1.09	2.12
Total Phosphate	46.4	101	47.2	NA	NA	160	147	124	59.3	160	19.5
Nitrate	< 0.8	< 0.8	< 0.8	NA	NA	<0.8	<0.8	<0.8	< 0.8	< 0.8	< 0.8
Nitrite	< 0.2	< 0.2	< 0.2	NA	NA	<0.2	<0.2	<0.2	0.26	< 0.2	< 0.2

Sample followed by KB is from Iron Precipitation Study.
Sample followed by KT is from Iron Precipitation Study.
U means the constituent was not detected.

J means constituent detection below quantifiable detection limit.

E means constituent detection above quantifiable detection limit.

ND means constituent not detected.

NA means constituent was not analyzed for.

TABLE 4-5

Well Point Screens

Well I.D.	Water Elevation	Upper Screen Interval	Lower Screen Interval
WP - 1	5.0'	2.0'-7.0'	8.0'-13.0'
WP - 2	approx 5'	2.0'-7.0'	8.0'-13.0'
WP - 3	approx 5'	2.0'-7.0'	8.0'-13.0'
WP - 4	18.4'	16.0'-21.0'	22.5'-27.5'
WP - 5	2.1'	1.0'-6.0'	7.0'-12.0'
WP - 6	7.80'	5.0'-10.0'	11.0'-16.0'

4.1.3 Well Points

Six groundwater elevation monitoring points were constructed on-site, see Table 4-5. Four well points were constructed by auguring down to depth, and the remaining two were constructed in boring locations. The well points were constructed by nesting two sections of 1.25" diameter PVC pipe. The screen zone was selected based on the location of the water table. The deeper screen was from 3'- 8' below the water table and the shallow screens were from 2' below the water table to 3' above. The annulus of the augured hole was allowed to naturally collapse as the augers were removed. After the screens were set based on the water table, the well point was finished off with either a cemented flush mount road box or a cemented protective casing. The logs of the well point construction are presented in Appendix C.

4.1.4 Groundwater Monitoring

4.1.4.1 Monitoring Well

One new well was installed on-site, WT-93-2 and one off-site, WT-93-4. The on-site well was completed in boring B-93-2. The off-site well was completed in B-93-4. The boring was augured to 52 feet, and the augers were pulled back to 25 feet and the hole was allowed to collapse below the augers. The off-site well was augered to a depth of 18.5 feet. The annulus of the wells was filled with #2 Morie sand. The sand interval was from 12 feet to 26 feet below the ground surface for WT-93-2 and 6 feet to 18 feet for WT-93-4. The wells were constructed with 2 inch diameter PVC and 10 feet of #20 slot screen. The screen was set from 24' to 14' for WT-93-2 and 8 to 18 feet in WT-93-4 and sand was placed 2 feet above the screen. The area above the sand pack was then filled with 2 feet of Baroid 3/8" bentonite pellets. The remainder of the well were filled with a bentonite-cement slurry to the surface and completed with cemented flush mounted road-boxes. The log of these wells are presented in Appendix C.

The on-site well was developed on March 3 and the off-site well on June 7 using the drill rig pump and a decontaminated intake hose. The hose was inserted into the well and the intake was placed at the bottom of the well. The pump was run at 4 gpm and the intake of the hose was moved up and down the well screen. After allowing the rig to pump water for 30 minutes the pump flow was varied as the intake line was moved up and down the screen. The rig pump removed 200 - 300 gallons of water from each well during development. The water quality at that time had stabilized and the clarity greatly increased.

4.1.4.2 Groundwater Sampling

The groundwater samples were collected during March from 16 wells around the Site. and on June 22 for the off-site well, see Table 4-6. The sampling was started by measuring the depth to water and the depth to bottom using a interface probe. Using the measured water levels, well volumes were calculated. The wells were purged using either a bailer or submersible pump. The well volume removed was monitored for: pH, conductivity, dissolved oxygen, temperature and turbidity. After each well volume was removed the parameters were measured and recorded. Table 4-6 summarizes these results. After a minimum of three well volumes were removed or after the parameters stabilized within 10%, the well was sampled. Sampling was performed using a decontaminated teflon bailer. The necessary sample bottles were filled and stored in a cooler. At the end of each day or every other day, the samples were picked-up or delivered to the analytical laboratory. Those samples not picked up each day, were stored on ice. Chain of custodies were completed and accompanied each sample delivery to track the samples. These are presented in Appendix B. Table 4-7 is a summary of the volatile organic chemical analyses of the groundwater samples. Table 4-8 is a summary of the semivolatile organic chemical analyses. Table 4-9 is a summary of the inorganic chemical analyses of the groundwater. The off-site well, WT-93-4, sampling results will be submitted with the July monthly progress report.

4.2 SOIL GAS SURVEY

Soil gas measurements by Dräger gas detector tubes were performed at the site using the PRESITSM methodology. The purpose of the investigation was to assess the lateral extent of contamination caused by VOCs in soils. Soil gas measurements were performed at various depths during the period of March 1, 1993 through March 5, 1993.

4.2.1 Measurement Procedures Applying Gas Detector Tubes

In vadose zone soils, the total concentration of volatile organic compounds analyzed from a soil sample describes a summation of constituents existing in three phases. Contaminants are present as vapors filling the soil pores, dissolved in the water making up the soil moisture, and adsorbed to soil solids. Partition co-efficients can be used to describe the distribution of contaminant concentrations between the different phases under equilibrium conditions.

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TABLE 4-6

Groundwater Sampling Parameters Shore Realty Superfund Site GIrnwood Landing, New York

Well I.D.	Date(s) Sampled	Temperature	Hd	Conductivity	Turbidity	Dissolved Oxygen
		<u>ي</u>		MHOS	NTUs	mg/l
WT-2	3/11/93	8.9	6.33	N/A	> 200	3.88
WT-5	3/10/93	7.8	5.80	N/A	> 200	3.08
DW-2	3/11/93	11.5	6.75	N/A	> 200	9.51
9-LM	3/9/93	0.9	7.24	N/A	> 200	3.90
SW-1	3/10/93	9.5	68.9	N/A	>200	N/A
WT - 14	3/5/93	10.9	6.28	435	> 200	3.97
DW-1	3/9/93	10.0	7.21	N/A	34.2	10.00
DW-3	3/8/93	12.3	5.21	20	69.5	8.75
WT-93-2	3/10/93	11.4	6.42	N/A	> 200	1.93
WT-13	3/8/93	7.8	8.09	482	> 200	2.88
SW-5	3/8/93	11.4	7.78	102	> 200	8.32
9-MS	3/8/93	12.8	4.28	160	> 200	8.06
WT-12	3/9/93	6.7	6.42	N/A	> 200	5.79
SW-2	3/9/93	9.6	6.55	N/N	> 200	8.75
SW-3	3/9/93	8.6	89.9	125	> 200	86.9
SW-4	3/9/93	10.0	98.9	48	>200	10.21
WT - 93 - 4	6/22/93	8.3	5.50	477	A/N	N/A

Turbidity and dissolved oxygen meter inoperable during field activities on June 22, 1993. NOTE: Conductivity Meter inoperable during field activities on March 9, 10, and 11, 1993.

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Groundwater Samples for TCL Volatile Organies (ug/l)
Shore Realty Superfund Site
Glemwood Landing, New York Table 4-7

Sample ID	7	DW-3	SW-2	13	SW-S	[.	SW-4		SW-5	SW-6	-	WI-12	WT-13	-	DW-1	DW-2	SW-1	7	WT-2	WT-S		WT-6	WT-93	3-2	WT-14
Laboratory ID	168	1603802	1603807	_	1603806	<u>ر</u>	1608805	_	1603803	1603801		1603808	1603804	I_	1607403	1607405	 -	1607404	1607407	1607408		1607401	1607402	22	1601001
% Moisure	Z.	-	×		ž		Ϋ́		Y.	X		۲ ۲	χ		V.	Ϋ́	_	٧×	٧	Y Y	_	٧×		Y.A	Š
Dilution Factor	1.8	Q	1.00		1.8		1.00	\exists	0.1	1.00		1.00	8		-			_	-	-		9.1		_	9:1
Chloromethane	•	10 UJ	• 10	10 UJ	• 10 UJ	=	10 UJ	·	10 UJ	• 10	Į.	10 UJ	• 10	·	10 UJ	10 UJ	Ŀ	10 UJ	· 100 UJ	• 10 UJ	ī		•	TO OU	· 10 U
Bromomethane	Ţ	10 C	• 10	10 U	10 U		10 U	•	10 01	• 10	Þ	10 U	• 101	'n	10 U	10 U	v	10 U	100 U	• 10	'n	200 C	·	100 U	10 U
Vinyl Chloride	•	10 U	• 10	10 01	10 U	-	10 U	•	10 C	• 10	Þ	· 10 U	0 .		10 U	10 O	_	10 U	100 U	• 10	• •	200 U	•	2 0 0	• 10 U
Chloroethane	-	10 U	• 10	10 U	0 OI .	5	10 C	_	10 C	01	Þ	· 10 U		·	10 C	. 10 U	•	10 U	· 100 U	01	<u>.</u>	200 C		D 00	10 U
Methylene Chloride	v	0 CJ	*	8 UJ	in o	17	SUJ	•	9 0.1	•	5	10 6 ·		5	34 UJ	· 14 UJ	•	10 11	130 J	20 UJ	5	960 J		120 J	• 5 U
Acetone	•	10 0.1	• 10	10 UJ		i	10 UJ	_	10 UJ	100	5	10 UJ	101	5	66 J	13 UJ	•	23 UJ	250 J	15 U	<u></u>	810 J	·	86 UJ	10 U
Carbon Disulfide		S U	٠	2 0	٠	· D	SU	•	5 U	٠,	- -	. 5 0	,	-	5.3	. s u	,	s or	· 50 UJ	٠.	5	250 UJ	·	50 UJ	· 5 U
1,1 -Dichloroethylene	•	5 U	٠	D		· D	SU	•	5 U	٠	D	. 5 U	,	·	SU	. S.U.	_	-	· 50 U	٠,	'n	250 U	·	20 O	• 5 U
1,1 -Dichloroethane	•	S U	٠	Þ	Š	·	SU	_	-	7	-	. 5 U	٠		SU	. 5 L	_	-	19 J	٠	· •	250 U	·	20 C	1.
1,2-Dichloroethylene (total)	•	S U	•	D	٠	÷	SU	•	SU	٠	ב	. 5 0	٠		SU	. 50	_	69	17 J	٠	_ _	44 J	•	20 C	23
Chloroform	•	S U	٠	:	Š	· :	S U	•	2 C	٠	: :	. 5 0	Š	·	SU	, ,	_	-	· 50 U	٠	<u>-</u>	250 U	·	20 C	. 30
1,2-Dichloroethane		SU	٠,	Þ	. 2	-	SU	•	SU	٠,	;	. 5 0	٠	·	S U	. 5 U	•	SU	· 50 U	٠,	<u>.</u>	250 U	·	20 C	· S U
2-Butanone	-	0.0	•	10 01	101 •	·	10 U	•	10 U	• 10	;	10 U	01	·	10 U	10 U	٠	10 U	· 100 U	2	· -	200 C	·	D 01	· 10 U
1,1,1-Trichloroethane	•	S U	٠	s u	7	_	SU		2.5	~		2 3	-	-	S U	. 5 0	_	20	23 J	٠,	· - -	250 U	•	20 U	11
Carbon Tarachloride		3 5	٠.	ם		· D	S U	•	2 13	٠,	ם	. 5 U		-	S U	. 5.		2 C	· 50 U	•	· n	250 U		20 O	. 50
Vinyl Acetate	-	0 0	. 10	10 U	01 .	· >	10 U	٠	.0 C	. 10		. 10 C		·	10 O	· 10 U		D 01	· 100 C	0.	· - -	200 U		2 0 0	10 U
Bromodichloromahme	v	5 U	٠		٠	· :	S U	_	S U	٠,	;	2 0 0	٠	· -	S U	٠ ، د	<u>·</u>	2 0	· 50 U	•	· - >	250 U	٠	20 C	. 50
1,2-Dichloropropme	•	SU	٠	þ	٠	· =	SU	•	S U	٠.	ב	. 5 U	,	· >	S U	. 5 1		S U	· 50 U	٠,	· >	250 U	•	20 C	· \$ U
ds-1,3-Dichloropropene	v	2 C	٠	n	٠	· =	5 U	•	3 C	٠,	n	. 5 0	٠	, D	2 0	<u>.</u> ۲		2 C	· 50 U	•	n	250 U	•	20 O	· \$ U
Trichloroethylene	v	3 0	٠	ה	٠ ،	• >	S U	•	2 C	_	<u> </u>	· 5 U	2		S U	, 5 t	_	13	· 50 U	٠ -	<u>·</u>	250 U	٠	20 O	3 J
Dibromochloromethane	·	S U	٠	ב	٠	·	5 U	_	SU		: :	. 5 U	٠	·	S U	٠ 5 ل	•	S U	· 50 U	•	· p	250 U	•	20 O	\$ O
1,1,2-Trichloroethane	•	s U	٠.	;	٠	·	S U	•	2 C	,	ם	. 5 U	٠ 5	· •	2 C	, 51	<u>.</u>	S U	· 50 U	•	Þ	250 U	•	20 O	· 5 U
Benzene	·	o	٠	Þ	٠.	· :>	2 C		SU	•	_ _	200	٠,	<u>.</u>	2 0	, 51		S U	· 50 C	٠ •	-	150 J		20 C	· \$ U
Trans-1,3-Dichloropropene	٧	2 C	٠	:	٠,	· :	5 U	•	S U		:	. s u	٠	·	SU	٠ ٢٠	•	S U	· 50 U	٠,	· - -	250 U		20 C	
Bromoform	•	2 C	٠	2	٠	'n	SU	•	SU	•		. 50	٠,	· -	S U	. \$ 0	•	2 0	• \$0 U	۰	· n	250 U	•	20 C	\$ O
4-Mahyl-2-Pentanone	·	n 0.	• 10	10 U	· 10 U	· -	10 U	•	10 U	• 10	Þ			· -	10 U	10 U	-	10 U	· 100 U	10 U	<u>·</u>	200 U	·	D 001	· 10 U
2-Hexanone	•	10 01	01	10 01	· 10 UJ	5	10 01	_	10 01	01		· 10 UJ	• 10	'n	10 U	. 10 U		0 01	· 100 U	10 C	· p	200 C	·	D 001	· 10 U
[Fetrachloroethylene]	<u>.</u>	2 0	•	SU	4	<u>.</u>	SU	•	2 0	•	- -	. s u		<u>.</u>	2 0	٠ ۲	_	20	101	٠,	· - -	250 U		20 C	18
1,1,2,2 - Tetrachlorochane		s u	٠	ם	۰.	·	SU	•	2 C	•	Þ	. s U	,	· •	2 0	, ,1	_	S U	· 50 U	۰.	· p	250 U	v	20 U	. 50
Toluene	•	s u	٠	D	۰	· >	SU	•	2 C	٠,	>	. s U	٠ د	D	S U	, ,1		S U	490	٠	Þ	15000 D		60	• S U
Chlorobenzene	. .	s	٠	ם	٠,		SU	_	2 C	٠	ם	. 5 U	٠ >	· -	S U	. 5 1	•	S U	• SO U	•	'n	250 U	٠	20 C	. SU
Ethylbenzene	•	S U	٠	ם :	٠,	· >	S U	_	2 C	٠.	ວ	. 5 U	٠ >	· >	S U	. 51	•	S U	2800 D	•	Þ	4700 J	_	88	, 5 U
Styrene	v	5 U	•	SU	. SU	÷	SU		3 C	٠	D	. 5 U	٠	·	SU	. 5 t		SU	· 50 U	•	Þ	210 J	·	20 U	, 5 U
Total Xylenes		SU	٠	0	٠	· n	SU	_	S U	٠	D	. 5 U	۰	·	5 U	, 5 L		SU	15000 D	٠	ם	25000 J	4	1500 J	· 5 U
Toyour I was I					4		c			4		,	,		34	c		24	100	0,0		2001		021	\$
Tari bara				_	•	_	• •	_				• 0					_		20031			15670	_		2
idal Blev			•		>]	,	_	>	,]	>	>]	>	>]	2	12220	2]	2000	2	3	>

U means the material was analyzed for, but not detected.

D means there was an analysis at a secondary dilution factor.

J means the associated numerical value is an estimated quantity.

UJ means the material was analyzed for, but not detected. The sample quantitation limit is an estimated value.

Table 4–8
Groundwater Samples for TCL Semivolatile Organics (ug/l)
Shore Realty Superfund Site
Glenwood Landing, New York

									1								
Sample ID	DW-3	SW-2	SW-3	SW-4	SW-5		WT-12	WT-13	DW-1	DW-2	SW-1	WT-2	WT-5	MT-6	WT-93-2		WT-14
Laboratory ID	1603802	1603807	1603806		1603803		1603808	_	1607403	1607405	1607404	1607407	1607408	1607401	1607402	16	1601001
% Moisture	¥ S	¥ ?	V S	Y S	Š,	ΨX.	Y S	¥ S	X,	Ž,	X Y	Ą.	Å,	¥ S	Ϋ́,	ž ·	
Dullion Factor	8.1	8:	8.1	3	3	8	3	3	•	-	-	1	1	8.1	-	_	3.
Polyauclear Aromatic Hydrocarbons																	
bis(2-Chloroethyl)cther	· 10 U	• 10 U	· 10 C	•	· 10 C			• 10					• 10 U		• 10 U	٠	10 C
1,3 - Dichlorobenzene	· 10 C	10 0	· 10 U	•	200			10								٠	10
1,4-Dichlorobenzene	. 10 U	100	· 10 C	•	· 10 U			• 10						10 U	10 U	٠	
Benzyl Alcohol	· 10 U	· 10 U	· 10 U	10				• 10								٠	
1,2-Dichlorobenzene	10 U	10 U	· 10 U	•				• 10			• 10 U			· 10 U		٠	10 C
bis(2-Chloroisopropyl)ather	· 10 U	· 10 U	· 10 U	• 10			• 10 U	• 10	· 10 U	• 10 U	· 10 U	• 10 U		• 10 U	• 10 U	•	10 U
N-Niroso-di-n-propylamine	· 10 C	• 10 U	· 10 U	٠	· 10 U		• 10 U	10		• 10 U	· 10 U	↑ 10 U		• 10 U	1 7	<u>•</u>	10 U
Hexachloroethane	, 10 U	10 U) O C	•	· 10 U	• 10 U	• 10 U	• 10	· 10 C	5 00 00	• 10 U	• 10 U	· 10 U	• 10 U	10 U	•	
Niro benzene	· 10 U	• 10 U	· 10 U	•	10 U			0.	• 10 U					• 10 U	• 10 U	•	10 U
Isophorone	· 10 U	, 10 U	· 10 U	•	· 10 U			• 10								•	10 C
Benzoic Acid	. S0 U	· 50 U	0.00	•	O 05			. 20							. So U	<u>.</u>	50 U
bis(2-Chloroethoxy)methane	· 10 U	7 10 C	· 10 U	•	. 10			• 10			· 10 C		. 10 U		• 10 U	-	5
1.2,4 - Trichlorobenzene	2 :	01 :	10 0	•	. 10			2 9				10 0.		10 01	, 10 t	•	D :
Naphthalene	0 ;	0 :	0 :	•	9 :			2 :		0 :				32	9 9	•);
4-Chloroaniline	100	100	0 :	. 10				2 :						0 0 0		<u>•</u>	0:
Hexachlorobutadiene	0:00	0:00:00:00:00:00:00:00:00:00:00:00:00:0	0 :		2:	. 10		2 9		2 :			•		10 5	<u> </u>	
2 - Methylnaphthalene	0 0 0	0 :	0 :	•	5 : 0 :			0 9		2 :				91	717	•	2 ;
Hexachlorocyclopentadiene	01	10 1	. 10 U	•	. 10			01 :		0 :	9 9				0 :	•	5 5
2-Chloronaphthalene	2 :	01 :		•	0 0			2 9		0 0	2;				0 0 0	•	0 :
2-Nitroanline	2 :	2 :	2 5	•	2 .			0, 5		2 5	2 :				2000	·	2 5
Acenaphthylene	2 :			-				•							100	•	2 :
2.6 - Dinitrotoluene		2 8		2 5	ت: و و			2 5) : •	2 3		2 5			•	2 5
3-Nitroaniline	S :	200	2 :	•	2 :			2 .		2 :	2 :		2 :		000	•	2 :
Acenaphthene	2 :	2 :			2 :		2 :	2 5					2 9	10 0		•	2 :
Dibenzoluran	2 :	2 :	2 :	•	9 9			2 9		2 :) or :		• -	2 :
2,4-Dinitrotoluene	0 5			2 5	9 5		•	• •	2 5	2 5	2 5	•	2 5	100	•	•	2 5
4-Cmorophenyi-phenyiquei	2 5				2 5			, ,	, ,		, ,					, ,	1 2
4 - Nitroaniline	20.00	2000	2000		. 8			. 20		20.00	20 05				\$0 C	•	S 8
N-Nerosodiphenylamine	· 10 U	. 10 U	· 10 U	•	10 CI			• 10	· 10 U	· 10 U	· 10 U		10 U	→ 10 U	1 7	•	10 U
4-Bromophenyl-phenylether	· 10 U	→ 10 U	• 10 U	1 · 10 U	· 10 U	· 10 U	· 10 U	10 U	· 10 U	• 10 U	• 10 U	• 10 U	· 10 U	• 10 U	. 10 U	•	10 U
. Hexachlorobenzene	· 10 U	• 10 U	· 10 U	•	· 10 U			• 10		• 10 U	• 10 U			• 10 U	• 10 U	•	10 U
Phenanthrene	· 10 U	• 10 U	· 10 U	•	· 10 C			10		• 10 U	• 10 U				· 10 U	•	10 U
Anthracene	· 10 U	. 10 U	· 10 C	•	· 10 U			• 10		• 10 U					· 10 U	<u>•</u>	10 U
Flouranthene	2 .	. 10 U	10 10 1	•	2 2			•		10 0	• 10 U				10 U	•	2 5
Pyrene	. 10	10 0	10 1	•				2 :		10 0	10 0			10 0	, 10 U	•	10 C
3-3'-Dichlorobenzidine	20 C	20 U	20 U	•				. 20		20 C	· 20 U			₹ 20 U	20 C	<u> </u>	70 C
Benzo (a) anthracene	10 0	. 10 U	. 10 .	• 10				2 9	0:0:		10 0					•	0 :
Chrysene	201	700	0 0 0	•				2 :							100	•	
Benzo(b)flouranthene) : 2 :							0 9	2 :					1000		•	
Benzo(x)monranthene			10 11								100			100		, ,	1 5
Denzo(a) pyrene Indeno(1,2,4-ed) rytene	1 2 2	10.01	10 0					2 2		10 Cl	10 11					, •	
Dibenz(a,b)anthracene	10 U	10 U	· 10 U	٠				• 10	10 U		10 U			10 U		•	
Denzo(g,h,i)perylene	· 10 U	10 U	· 10 U	1 10 U		· 10 U	· 10 U	• 10 U	• 10 U	• 10 U	◆ 10 U	• 10 U	· 10 U	• 10 U		_	

PDIR/TAB4-8.WK1

Table 4-8 (continued)
Groundwater Samples for TCL Semivolatile Organics (ug/l)
Shore Realty Superfund Site
Glenwood Landing, New York

Sample ID	DW-3	SW-2	SW-3	5W-4	SW-5	SW-6	WT-12	WT-13	DW-1		SW-1	WI-2	WT-5	- WT-6	WT-93-2	13-2	WT-14
Laboratory ID	1603802	1603807	1603806	1603805	1603803	1603801	1603808	1603804	1607403	1607405	1607404	1607407	1607408	1607401	160	02	1601001
% Moisture	٧Z	۲ ۲	۷ Z	Ϋ́	۷.	٧×	٧Z		Ϋ́		Ϋ́	Ϋ́	Ϋ́ X	4 Z	ž	-	٧X
Diluion Factor	1.00	1.00	1.00	1.00	1.00	1.00	1.00		1		-	-	1	1.00		-	1.00
l'hihalates																	
Dimethylphthalate	· 10 U	10 U	10 U	10 U	· 10 U			•					• 10	10 U	·	200	10
Diethylphthalate	· 10 U	· 10 U	10 U	10 U	· 10 U		• 10 U	10	· 10 U				٠	10 U		1.1	10
Di-n-butylphthalate	· 10 U	→ 10 U	· 10 U	· 10 U	· 10 U			• 10					• 10	10 U	•	0.0	10
Butylbenzylphthalate	· 10 U	→ 10 U	• 10 U	10 U	· 10 U			• 10					10	10 U	•	0.0	10
bis (2 - Ethylbexyl) phthalate	· 10 U	→ 10 U	· 10 U	· 10 U	· 12 U	→ 10 U	10 U	· 10 U	1 1	v 10 U	· 10 U	• 10 U	_	· 10 U	•	10 U	• 10 U
Di-n-octylphthalate	3 1	5 J	· 10 U	1.	• 10 U		4		5 J					· 10 U	·	10 C	10
Phenois																	
Phenol	· 10 U	→ 10 U	• 10 U	10 U	10 U	→ 10 U	→ 10 U		· 10 C					4 J	·	200	10
2-Chlorophenol	· 10 U	→ 10 U	· 10 U	10 U	• 10 U	• 10 U	· 10 U	1 · 10 U	· 10 U	•	10 U	10 U					
2-Methylphenol	· 10 U	10 U	· 10 U	· 10 U	· 10 U	· 10 U	· 10 U	• 10	· 10 U			30	10	920 E		9 1	10 U
4-Methylphenol	· 10 U	10 U	• 10 U	• 10				11	6 3	310 E		6 J	10				
2-Nitrophenol	· 10 U	10 U	· 10 U	• 10					• 10	· 10 U	•	10 11	10				
2,4-Dimethylphenol	. 10 C	· 10 U	→ 10 U	· 10 U	· 10 U			• 10						440 臣		=	10
2,4-Dichlorophenol	· 10 U	10 U	10 C	· 10 U	· 10 U	O 01 ·	· 10 U	• 10					•	• 10 U		10 11	10
4-Chloro-3-methylphenol	· 10 U	· 10 U	10 U	· 10 U	• 10 U	O 01		• 10					10	→ 10 U	•	0 0	10
2,4,6-Trichlorophenol	· 10 U	· 10 U	10 U	· 10 U	· 10 U	10 U		• 10					• 10	→ 10 U	•	0 11	10
2,4,5-Trichlorophenol	→ 50 U	• 50 U	→ S0 U	1 SO U	• 50 U	• 50 U	· 50 U	\$ 20					• 50	→ 50 U		20 12	50
2,4 - Dinitrophenol	· 50 U	→ 50 U	· 50 U	· 50 U	• 50 U			•				• 50 U	•	→ 50 U		2 J	50
4 - Nitrophenol	· 50 U	• S0 U	1 SO U	• S0 U	• 50 U			• 50					\$ 50	• 50 U	·	20 02	\$
4,6-Dinitro-2-methylphenol	· 50 U	· 50 U	1 SO U	· 50 U	• 50 U	v S0 U	• 50 U	• 50					• 50	50 U		1.	50
Pentachlorophenol	11 05	11 05	11 05	. 50 11	50 11			\$,	50 11		11 03	05

U means the constituent was not detected.

J means constituent detection below quantifiable detection limit.

E means constituent detection above quantifiable detection limit.

ND means constituent not detected.

NA means constituent was not analyzed for.

Groundwater Samples for Inorganics (ug/L)
Shore Realty Superfund Site
Glenwood Landing, New York Table 4-9

WT-14T	601001	NA	٧X	VA	٧z	- AN	Y X	NA AA	۲×	Y.V	Ϋ́	14800	Y.	Y Z	NA	٧X	ΥN	٧Z	V.	٧Z	Y X	٧Z	Y X	ΑN	NA	11.5	15900	40	100	14500	10	13.1
	607402	AN	Y'N	AN	Y X	AN	- AX	NA VA	NA AN	Y.	NA	101000	NA	Y'N	NA	ď Z	NA	Y X	NA	Y Y	Y X	NA AN	A'N	NA	ν _χ	1480	1100	17 4	120	D 4	4	8850
WT-6		AX	۲ ۲	Y V	Ϋ́Z	NA VA	Ϋ́	NA V	٧×	ĄX	NA VA	4990	Y Y	Y V	AN	٧z	NA AN	ĄZ	Ϋ́Α	Ϋ́Z	Ϋ́	Ϋ́	Y'A	Ϋ́Υ	Y Z	2840	2610	11 7	400	04.	7	1260
WT-5	607408	NA	٧z	AN	Ϋ́Z	Ϋ́	Ϋ́	ΑN	Ϋ́	Ϋ́	Ϋ́Υ	40600	Ϋ́	٧X	٧X	Ϋ́	Ϋ́Α	٧×	ΑN	Ϋ́Z	Ϋ́Α	Ϋ́	ΑN	ΑN	Y Y	2450	2590	7	•	D.4	S	1890
WT-2	607407	NA	٧X	A'N	٧z	٧X	٧x	Ϋ́	٧X	٧X	Ϋ́	19300	Ϋ́	٧ ٧	Ϋ́	۲ ۲	٧X	٧X	Ϋ́Z	٧X	٧X	Ϋ́	٧X	ΥZ	Y Y	280	810	17	110	. t U	· 1U	1450
SW-1	607404	NA NA	Ϋ́	Ϋ́	₹Z	Ϋ́Α	Ϋ́	Ϋ́Х	×z	٧X	٧X	974	Ϋ́	۷ ۲	ΑN	۲X	٧X	٧Z	٧Z	۲Z	Ϋ́	Ϋ́Z	Ϋ́	Υ _N	Ϋ́Z	N S	13000	7	SU	12100	1 U	10 U
DW-2	607405	NA	٧X	AN	Ϋ́Z	A'N	Š	Ϋ́	Ϋ́	Ϋ́Z	Ϋ́χ	26600	Ϋ́	٧X	ź	۲ ۲	Ϋ́	ž	Ϋ́,	Ϋ́Х	۲ ۲	۷ ۲	۷ Z	AN AN	X	v 5 U	1480	, A		1440	1 1	10 U
DW-1	607403	NA	Y.	ΑΝ	Ϋ́Α	NA AN	Y X	Ϋ́	Ϋ́	Ϋ́Α	Ϋ́	112	Ϋ́	Ϋ́	Ϋ́Z	Ϋ́	٧X	۷.	Ϋ́	ΝΑ	Ϋ́Α	Ϋ́	Y Z	A'N	A'N	. S U	2000		U	1880	· 1 U	· 10 U
WT-13	603804	NA	Ϋ́	Y.	Ϋ́	Ϋ́	٧X	Ϋ́	Y.	Ϋ́,	A'N	51300	ΑN	ζ Z	Ϋ́	Ϋ́Z	Ϋ́	Y.	Ϋ́	Ϋ́	Υ _N	Y X	Y V	ΑN	X V	3 0	6980	. T	2970	5370	1 U	15.6
WT-12	603808	NA	Ϋ́Α	Ϋ́	Ϋ́	Ϋ́	A Z	YZ.	۷ ۲	۷ ۷	Y.	1810	Ϋ́Α	Ą Z	Ϋ́Α	۲ ۲	Ϋ́Α	Y Z	NA NA	۷×	۲ ۲	Ϋ́	Y X	Ϋ́	₹ Z	U S	2710	, D	140	2350	· 1 U	181
SW-6		NA	Ϋ́	V.V	Ϋ́	××	Ϋ́.	Ϋ́ν.	Ϋ́ X	V.	۲ ۲.	· 59000 U	Y.	V.	٧×	Υ.	Y.	Ϋ́	Y.Y	٧.	Y.	Y.Y	Ϋ́Х	Ϋ́,	X V	. S U	7270	4	290	7010	. 10	< 10 U
S-WS	603803	NA NA	٧	Ϋ́	Ϋ́Z	Ϋ́Α	Ϋ́	Ϋ́Α	Y.	ΥN	ΥN	67000	Ϋ́	٧X	Ϋ́	٧X	ΑN	۷ Z	ΥN	Š	Υ _Z	٧X	٧X	٧×	۲X	\$ U	4750	D 4	170	4410	· 10	10 U
SW-4	603805	NA AN	Ϋ́	ΝΑ	Ϋ́Z	Ϋ́	٧X	Ϋ́	Ϋ́	Υ _N	ΥN	30300	ΑN	Y X	VA	ď	ΥN	Ϋ́	ΝΑ	Ϋ́	٧X	Y.	۷ ۲	A'N	₹ Z	s U	2580	7	180	2120	· 1U	· 10 U
SW-3	603806	Ϋ́	Y Z	NA A	۷ ۲	N.	Y Y	Y X	Y Z	××	Y Z	1380	Ϋ́	Y Y	NA A	Ϋ́	N.A	Z A	Y.	Y X	Ϋ́	Y.	ÝZ	VX	ν V	\$ U	3890	7	ς,	3950	· 1 U	• 10 U
SW-2	603807	NA NA	۲ Z	Ϋ́	۲ ۲	٧X	Y.	Ϋ́	٧X	Y Z	××	8620	Y.	Y Y	٧X	۷ Z	Ϋ́	٧X	Y Z	Y X	¥Z	Y Z	Š Z	۷ ۷	X X	· 5 U	3730		160	3220	1 U	· 10 U
DW-3	603802	Ϋ́	۷ ۷	Y Y	Y Z	Υ _X	٧X	V.	V	Y X	٧Z	105	Y X	٧X	٧X	Y X	Y V	Y Z	Y Y	Y Z	×	٧	۲ ۲	ž	< Z	• S U	2120	4		2510	. 1 U	· 10 U
Sample ID	Laboratory ID	Aluminum	Antimony	Arsenic	Barium	Beryllium	Cadmium	Calcium	Chromium	Cobalt	Copper	Iron	Lead	Magnesium	Manganese	Mercury	Nickel	Potassium	Selenium	Silver	Sodium	Thallium	Vanadium	Zinc	Cyanide	Ammonia Nirogen	Total Narogen	Orthophosphate	Total Phosphate	Natrate	Narite	Dissolved Iron

U means the constituent was not detected.

J means constituent detection below quantifiable detection limit.

E means constituent detection above quantifiable detection limit.

ND means constituent not detected.

NA means constituent was not analyzed for.

· ` _ _____

Soil gas concentrations indicative of relevant source areas should be expected to be within the ppm range. Substance-specific gas detector tubes are designed to operate in this range, thus allowing an instantaneous evaluation and delineation of source areas in a cost-effective manner. The PRESITSM methodology uses gas detector tubes for *in situ* soil gas analyses at ambient pressures, and requires relatively small volumes of gas for analysis. This method, as applied at this facility, requires a borehole of 1 inch in diameter established with a grooved boring rod driven into the ground by means of a jackhammer. A probe containing a gas detector tube in the tip is inserted into the boring. Using a small, hand-operated bellows pump, a defined volume of gas from the bottom of the boring is drawn through the detector tube. The gas volume required is dependent on the type of tube used, and the indicating range desired; for thetubes used, the volume varied from 500 to 1,000 ml. The tube contains a reagent that changes color in the presence of a specific chemical. The length of the color band in the tube indicates the concentration of the chemical vapor. The instantaneous reading enables the field crew to adapt the investigation program in terms of locations and depths to the actual contamination pattern.

Accuracy of the measurement is a function of both incidental and systematic error. Where incidental error is a measurement of fluctuation when several measurements are taken of a precisely defined concentration, this incidental error is quantified and reported as standard deviation by the manufacturer. This standard deviation, which is actually a co-efficient of variation (i.e., relative standard deviation), is given as a percentage and relates to the mean value. In contrast, systematic error such as miscalibration, storage effects, equipment malfunctions, and "cross-sensitivity" cannot be calculated using statistical methods, but can be avoided. For some tubes, measurement beyond detection range can be taken by using more or fewer pump strokes, although, more precise readings can be expected at concentrations within the detection range. The Dräger detector tubes used for this survey, together with their respective range are presented in Table 4-10.

Different gas detector tubes show different degrees of cross-sensitivity. They not only indicate the substance they are designed and calibrated for, but also some other compounds of similar chemistry. For example, the benzene tube is sensitive to alkyl benzenes (toluene) and petroleum hydrocarbons; the methylene chloride tube to TCE, PCE, TCA, octane, and toluene; the PCE tube and the TCE tube to free halogens, hydrogen halides, and easily cleaved halogenated hydrocarbons; and the petroleum hydrocarbon tube to volatile petroleum hydrocarbons including n-octane, n-hexane, n-heptane, iso-octane, n-nonane, benzene, toluene, and xylene. Hence, the strength of the method is in the ability to determine the relative change in the readings as a site is traversed.

TABLE 4-10
Summary of Dräger Tubes Used During Survey

Title (Tube Type)	Detection Range
Benzene (5/b)	5-50 ppm
Methylene Chloride (100/a)	100-200 ppm
Perchloroethylene (2/a)	2-40/20-300 ppm
Trichloroethylene (2/a)	2-50/20-250 ppm
Trichloroethane (50/d)	50-600 ppm
Petroleum Hydrocarbon (100/a)	100-1,200 ppm

The highest readings indicate the potential source area as distinguished from other areas where readings are relatively low or non-detectable. Because the method is not intended to quantify the exact chemical composition of the soil gas, however, results are considered to be only semi-quantitative. The manufacturer gives a relative standard deviation of 10-15% for benzene tubes, 10-15% for methylene chloride tubes, 15-20% for PCE tubes, 10-15% for TEE tubes, and 10-15% for petroleum hydrocarbon tubes, if only a single substance is involved.

4.2.2 PRESIT Soil Gas Survey Results

A drawing showing the site, the location of each soil gas survey point, and the results of the soil gas measurements is presented as Figure 4-2. Table 4-11 lists the results of the measurements, soil characterization, total depth, water level, and any additional comments. Copies of the field data sheets prepared on-site during the investigation are presented in Appendix D.

4.3 HYDROGEOLOGIC INVESTIGATION

4.3.1 Permeability Samples

The permeability samples were collected from two of the borings, B-93-2 and B-93-3. The samples were intended to be collected using thin walled shelby tubes, but due to the sandy soil, success was minimal.

At B-93-2, shelby tubes were attempted at 25'-27', 27'-29' and 50'-52'. The three attempts had no soil recovery due to the lack of cohesiveness in the soils. At WP-4, next to B-93-2, a shelby tube sample was successfully collected from 13'-15'. The recovery was about 2/3 full, but the sampling tube stuck into the sampling head. The tube was cut off at the sampling head and sealed with end caps and wax. A shelby tube failed at 18'-20'. The sample was then collected from 20'-22' using a split spoon and plastic sleeve. A shelby tube sample was collected from 22'-24'. All tubes were capped, sealed with wax and stored upright.

The sampling efforts at B-93-3 were unsuccessful using the shelby tubes. Samples were collected using split spoons with liners from 22'-24' and 47'-49'. The samples were capped and sealed with wax, and stored in an upright position. Table 4-12 summarizes the vertical hydraulic conductivity analysis results. Due to poor recovery and the samples being collected with a



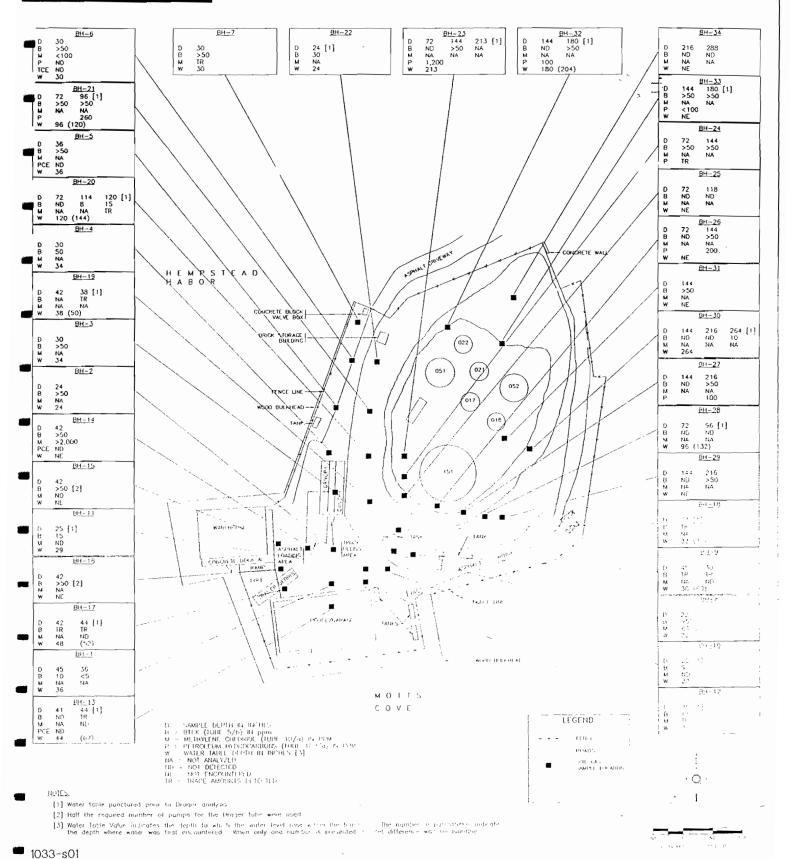


TABLE 4-11

Summary PreSIT** Soil Gas Survey Results

Comments		Moved hole twice	44"-77" heavy odor	30"-42" heavy odor		tried perchloroethylene - ND	tried perchloroethylene - ND, trichloroethane - ND, 15"-44" heavy odor	27" - 42" oil sheen, heavy odor	18"-42" faire odor
Water Level		36"	24	34."	34.	36.	30	30	22
Total Depth		-21	72*	42"	44-	44"	44"	42*	42"
Soil Characterization		0-6" concrete 6"-74" dark sand and gravel 44"-72" tan sand, some dark spots	0-8" asphalt 8"-24" brown sand 24"-44" brown sand to tan sand to grey sand 44"-72" grey sand	0-2" organics (weeds) 2"-30" reddish brown sand per gravel 30"-42" dark grey sand	0-2" organics (weeds) 2"-29" dark brown sand 20"-39" dark brown sand and light tan sand 39"-44" light tan sand	0-2" organics 2"-38" dark brown sand and gravel 38"-44" charcoal color sand	0-2" organics 2"-15" brown sand and gravel 15"-44" black sand and some gravel	0-2" organics 2"-27" brown sand and gravel 15"-44" dark discolored sand and gravel	0-4" asphalt 4"-18" dark brown to black sand and gravel 18"-42" dark grey brown sand
Methylene Chloride 100/a Range 100 to 200 ppm	1:						< 100 ppm	TR	< 100 ppm
Benzene 5/b Range 5 to 50 ppm		10 ppm <5 ppm	> 50 ppm	> 50 րրու	50 ppm	> 50 ppm	> 50 ppm	>50 ppm	>50 ppm
Depth	1, 1993	45° 36°°	24"-	30".	30	36."	30	30	22
I.D.	MARCH 1, 1993	BH-1	BH-2	BH-3	BH-4	BH-5	BH-6	BH-7	BH-8

TABLE 4-11 (continued)

Summary PreSIT** Soil Gas Survey Results

Comments	move hole due to concrete; water table 60"	water table 32"				perchloroethylene - ND water table 67"	perchloroethylene - ND heavy odor 18-29*	Drager tubes '/2 number of pumps 17*-36* heavy odor
Water Level	30	20.	67	35				
Total Depth		42"	43"	43.		.29	42.	42.
Soil Characterization	0-6" asphalt 6"-20" light brown to black sand and gravel 20"-33" dark brown sand 33"-44" grey sand 44"-45" grey brown sand	0-4" asphalt 4"-24" dark brown to black sand and gravel 24"-42" dark charcoal sand	0-4" asphalt 4"-21" black sand and gravel 21"-36" dark sand and gravel 36"-43" dark sand	0-4" asphalt 4"-24" tan sand 24"-38" dark brown sand 38"-43" grey sand		0-6" concrete 6-19" dark sand and gravel 19"-30" brown sand 30"-39" black sand 39"-58" brown sand 38"-67" grey sand	0-10° asphalt 10°-18° dark sand and gravel 18°-29° dark sand 29°-42° tan sand	0-4' asphalt 4'-17' brown sand 17'-36" grey brown sand mix 36"-42" light tan sand
Methylene Chloride 100/a Range 100 to 200 ppm	ND	ND	ΩN	TR		ND	> 2.000 ppm	>2,000 ppm
Benzene 5/b Range 5 to 50 ppm	TR TR	mdd 6	15 ppm	< 5 ppm		ND TR	>50 ppm	-> 50 ppm
Depth	30	20	25 ***	35".	2, 1993	41"	42"	45
I.D.	ВН-9	BH-10	BH-11	BH-12	MARCII 2, 1993	BH-13	BH-14	BH-1,5

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TABLE 4-11 (continued)

Summary PreSITs* Soil Gas Survey Results

	Benzene 5/b Range 5 to	Methylene Chloride 100/a		Total	Water	
[S0 ppm	Range 100 to 200 ppm	Soil Characterization	Depth	Level	Comments
	>50 ppm		0-6" asphalt 6"-42" dark brown and grey sand	42"		Dräger tubes 1/4 number of pumps 6"-42" heavy odor
	XT XT	Q N	0-10" asphalt 10"-19" sand and gravel fill 19"-36" brown sand and gravel 36"-53" brown sand	.29	** 84	water table 52"
	TR		0-2" organic weeds 2"-42" brown sand	42.	33**	water table 35"
	ND TR		0-2" organic weeds 2"-67" brown sand		38"	water table 50"
	ND 8 ppm 15 ppm	TR TR	0-144" tan sand 144"-168" grey sand	168*	120**	water table 144"
	> 50 ppm > 50 ppm		0-120° brown sand 120°-144° grey sand	144*	96	water table 120° petroleum hydrocarbon 260 ppm
	30 ppm		0.2" organic weeds 2"-41" dark sand	41"	24	
	ND >50 ppm		0-6" dark organic sand 6"-36" sand and gravel 36"-72" sand 72"-140" sand 140"-240" grey sand	216-	*213*	petroleum hydrocarbon 213" 1,200 ppm heavy odor 1407-213" water table 213"

TABLE 4-11 (continued)

Summary PreSITs" Soil Gas Survey Results

Comments	petroleum hydrocarbon TR	118" - 120" refusal moved hole twice	benzene at 144°, black petroleum hydrocarbon 200 ppm heavy odor 6°-144°	heavy odor 180°-216° petroleum hydrocarbon 100 ppm	water table 132"	189*-216* heavy odor	water table 264"
Water Level					96		264"•
Total Depth	144"	120*	144"	216"	144"	216*	288.
Soil Characterization	0-6" dark organic sand 6"-60" orange and grey sand mix 60"-109" grey sand 109"-139" light to organic sand 139"-144" tanish brown sand	0-68" sand and gravel 68"-72" white sand 72"-118" sand gravel	0-6" black sand 6"-72" brown sand 72"-144" brown sand	0-6" organic and dark sand 6"-180" brown sand 180"-216" dark brown to black sand	0-3" organic sand 3"-67" brown sand 67"-72" dark brown sand 72"-138" dark brown sand 138"-144" dark brown sand	0-3" organic 3"-180" brown sand 180"-216" black sand	0-3" organic sand 3"-82" brown sand 72"-180" dark brown sand 180"-216" black sand 216"-288" grey sand
Methylene Chloride 100/a Range 100 to 200 ppm							
Benzene 5/b Range 5 to 50 ppm	mdd 05 <	ND ND	ND > 50 ppm	ND >50 ppm	O Z Z	, ND >50 ppm	DN ND 10 ppm
Depth	72" 144"	72"	72" 144"	144" 216"	72" 96"	144"	144" 216" 264"•
I.D.	BH-24	BH-25	вн-26	ВН-27	BH-28	ВН-29	BH-30

TABLE 4-11 (continued)

Summary PreSITs* Soil Gas Survey Results

		Benzene 5/b	Methylene Chloride				
1.D.	Depth	Range 5 to 50 ppm	100/a Range 100 to 200 ppm	Soil Characterization	Total Depth	Water Level	Comments
MARCH 4, 1993	4, 1993						
BH-31	144"	> 50 ppm		0-3" organic 3"-120" brown sand 120"-144" tan sand	144"		120°-144° heavy odor
вн-32	180"	ND >5 ppm		0-6" organic sand 6"-120" brown sand 120"-144" grey sand 144"-216" whte sand	216"	180**	water table 204" petroleum hydrocarbon - 144"-100 ppm
вн-33	144"	>50 ppm		0-6" organics and sand 6-77" brown sand 72"-109" tan sand 109"-120" red sand	144"		petroleum hydrocarbon <100 ppm
BH-34	216" 288"	N ON ON		0-6" organics and sand 6'-288" brown sand	288"		

TR ND *

Trace Not Detected After punching through water table.

Summary of Vertical Hydraulic Conductivity Analysis
Shore Realty Superfund Site
Glenwood Landing, NY Table 4-12

Notes	Mot Domoldad	INOL NEITHONGE	Remolded	Remolded	Remolded	Not Remolded
Permeability (cm/sec)	23-10-2	01 X C.2	9.2 x 10 ⁻⁴	6.2 x 10 ⁻⁴	3.8 x 10 ⁻⁵	2.0 x 10 ⁻⁶
Effective Confining Stress (psi)	XX	WM	NA	NA	NA	4.7
Confining Stress (psi)	-	1.1	1.3	1.2	1.2	7.5
Water Content (%)		4.3	15.4	19.1	16.6	17.3
Dry Unit Weight (pcf)	6.20	74.3	97.3	113.2	112.6	109.2
Wet unit Weight (pcf)	7 00	78.4	112.3	134.8	131.3	128.1
Depth		13-13	20-22	22-24	47-49	24-26
Sample	3 4 5 5	W1'-4	WP-4	13-03-31	13-93-3K	B-93-2J

Samples that were remoided could not be retneved with shelby tubes due to sandy soil and had to be retneved by driving a split spoon with a plastic sleeve inside it.

NA - Data not analyzed for or not applicable

pcf – pounds per cubic foot psf – pounds per square foot

driven split spoon instead of a shelby tube, the samples had to be remolded in order to have a sufficient for the analysis.

4.3.2 Determination of Hydraulic Conductivity Based on Cyclic Water Level Fluctuations

The tidal fluctuation technique of Ferris (1963) was used to determine the transmissivity and the horizontal hydraulic conductivity at the site. A hydraulic profile was developed for Level B and Level C perpendicular to the edge of Hempstead Harbor. The profile consisted of three driven wellpoint clusters: WP-1, WP-2, and WP-3, installed at distances of 17, 35, and 50 feet from the harbor (see Figure 4-1). Each cluster consisted of two wellpoints, a shallow one screened in Level B and a deep one screened in Level C. Two data loggers and nine pressure transducers were used to record water level fluctuations over a seven-day period. However, one of the data loggers experienced mechanical problems during the first day of the study, so two of the well points have an additional day of data. In addition to monitoring water level fluctuations in the aquifer, tidal fluctuations of the Hempstead Harbor were also measured using the same data logger and pressure transducer system. The tidal fluctuations at the harbor, and the water level response to these fluctuations at the wellpoints, were monitored simultaneously and continuously for a period of seven days. In order to maximize the tidal fluctuations, the monitoring period was chosen to include three days before and three days after a full moon. The system was also set up before a storm which caused local flooding, and was operated from March 4 to March 11, with water levels monitored every 10 minutes.

According to Ferris (1963), in an aquifer bounded by a body of tidal water, the water level in the aquifer will respond to the tidal fluctuations. This is known to be the case at the Site, based on the tidal fluctuation data shown in Figures 4-3, 4-4, and 4-5. When the tides fluctuate as a simple harmonic motion, a train of sinusoidal waves is propagated through the aquifer. Because the amplitude of each transmitted wave decreases as the distance from the boundary (tide source) increases and the time lag of a given maximum or minimum increases as the distance from the boundary increases, the transmissivity of the aquifer can be found by either the stage-ratio method or the time-lag method (Ferris, 1963). Both methods were employed to determine the transmissivity at the Site.

4.3.2.1 Stage-Ratio Method

The stage-ratio method can be summarized as follows. The ratio of the groundwater fluctuation to the tidal fluctuation is computed for the rising and falling limb of each cycle, and the averages of the ratios for rising and falling limbs are calculated for each wellpoint. The average length of the period of the tidal fluctuation is computed and the averages of the ratios

for rising and falling limbs are plotted on semi-logarithmic graph paper against the distances of the observation wells from the edge of the tide source (Hempstead Harbor). The transmissivity is calculated from the following equation:

$$T = \frac{4.4 (\Delta x)^2 S}{t_o} \qquad where;$$

T = Transmissivity,

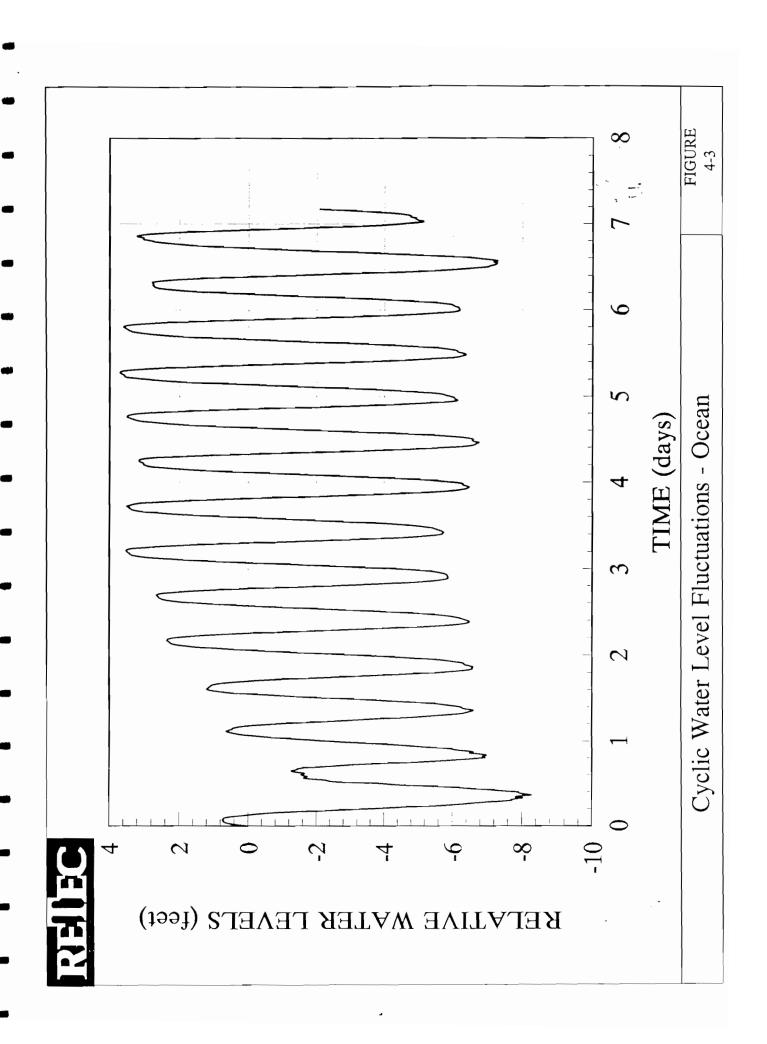
 Δx = Distance From Edge of Surface Water Body,

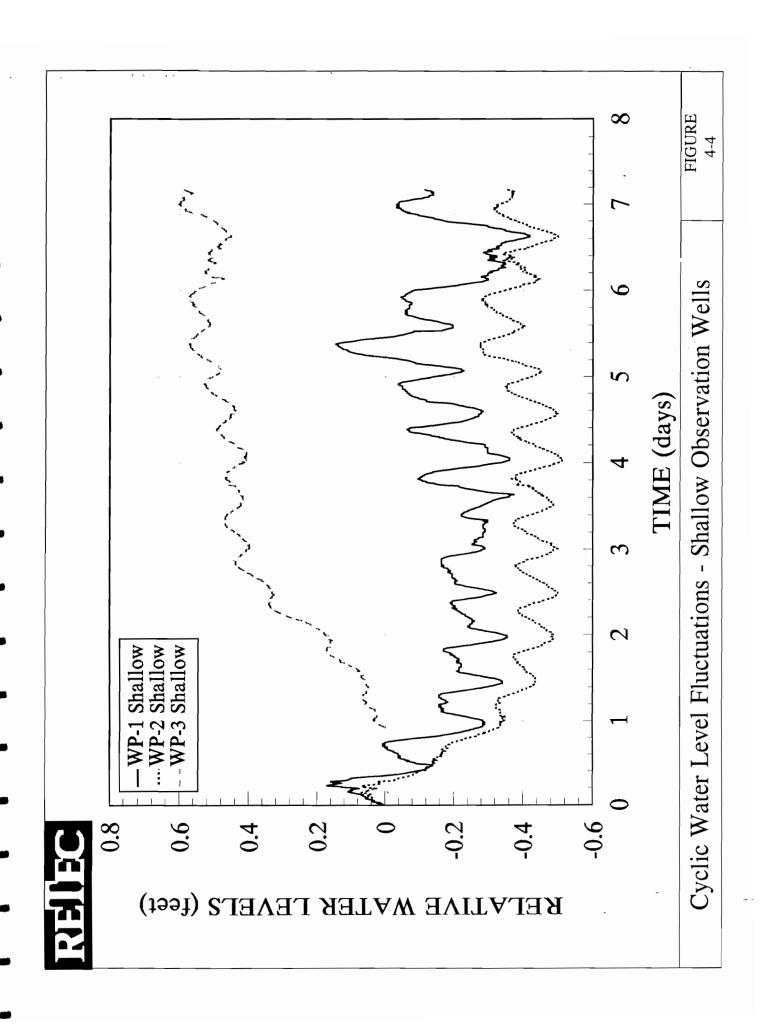
S = Storativity, and

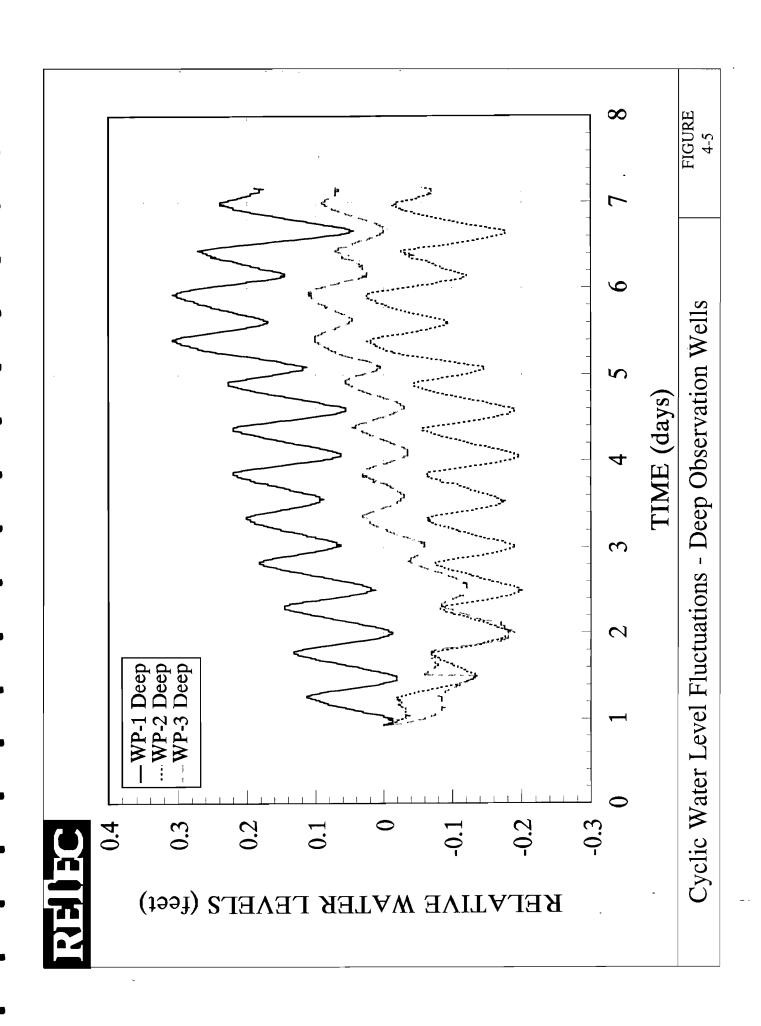
t_a = Periodicity of Tidal Fluctuation

Figures 4-3, 4-4, and 4-5 depict normalized plots of water levels for Hempstead Harbor, the shallow wellpoints, and the deep wellpoints. The shallow wellpoints were screened in Level B, and the deep wellpoints were screened in Level C. Using the water level data of Figures 4-3, 4-4, and 4-5, the ratio of groundwater fluctuation to Hempstead Harbor stage fluctuation was computed for the rising and falling limb of each cycle. These ratios are listed in Table 4-13. The length of the period of the Hempstead Harbor fluctuation averaged 0.51 days. The averages of the ratios for rising and falling stages were calculated and are included in Table 4-14. Figures 4-6 and 4-7 plot these average ratios against the distances of the observation wells from Hempstead Harbor's edge.

As shown on these figures, for one log cycle, $\Delta x = 59$ feet in case of the shallow wellpoints, and $\Delta x = 101$ feet in case of the deep wellpoints. The U.S. Geological Survey determined the storativity of the Upper Glacial Aquifer in the vicinity of Glenwood Landing to be 0.10 (Getzen, 1977). Thus, the transmissivity is 400 ft²/day (3,000 gpd/ft) for Level B, and 1,175 ft²/day (8,800 gpd/ft) for Level C. If the saturated thickness of Level C is 22 feet (Roux, 1991a), and the saturated thickness of Level B is 7 feet in the vicinity of WP-1, WP-2, and WP-3, then the horizontal hydraulic conductivities of Level B and Level C are 57 ft/day (426 gpd/ft²) and 53.5 ft/day (400 gpd/ft²), respectively. Well logs (soil boring B-93-1) show that Level B







wells WP-1, WP-2, and WP-3 to the corresponding range in stage of the Ratio of the range in water level fluctuation in observation Hempstead Harbor Table 4-13

Comment of Curyes	3011	Chal	Challon Obsernation		J.	Doon Oheanmtion	
of cyclic		Ollan	Wells		3	Wells	•
fluctuations		WP-1	WP-2	WP-3	WP-1	WP-2	WP-3
Rising Stage	1-2	1	1 1	1	0.0164		
Falling Stage	2-3	 	1	0.0018	0.0181	 	0.0080
Rising Stage	3-4	 	1 1	0.0149	0.0190	0.0087	0.0099
Falling Stage	4-5	 	0.0149	0.0017	0.0182	0.0153	0.0153
Rising Stage	2-6		0.0114	0.0207	0.0171	0.0121	0.0114
Falling Stage	2-9	1 1	0.0130	0.0014	0.0145	0.0138	0.0043
Rising Stage	7-8		0.0134	0.0120	0.0183	0.0141	0.0095
Falling Stage	8-9	!!!	0.0136	0.0045	0.0140	0.0136	0.0030
Rising Stage	9-10	l !	0.0144	0.0075	0.0147	0.0137	0.0096
Falling Stage	10 - 11	1	0.0125	0.0063	0.0122	0.0125	9900.0
Rising Stage	11 - 12	0.0299	0.0132	0.0056	0.0142	0.0125	6900.0
Falling Stage	12 - 13	0.0269	0.0154	0.0064	0.0154	0.0135	2900.0
Rising Stage	13-14	0.0313	0.0153	0.0093	0.0163	0.0147	0.0083
Falling Stage	14-15	0.0232	0.0135	0.0058	0.0165	0.0135	0.0074
Rising Stage	15-16	0.0238	0.0150	0.0088	0.0169	0.0144	0.0081
Falling Stage	16-17	0.0200	0.0113	0900.0	0.0120	0.0109	0.0050
Rising Stage	17 - 18	0.0377	0.0182	0.0097	0.0201	0.0175	0.0097
Falling Stage	18 - 19	0.0344	0.0134	0.0057	0.0143	0.0121	0.0057
Rising Stage	19-20	 	0.0129	0.0058	0.0142	0.0123	0.0065
Falling Stage	20-21			0.0105	0.0167	0.0147	0.0088
Rising Stage	21 - 22			0.0079	0.0143	0.0104	0.0050
Falling Stage	22-23	 		0.0064	0.0231	0.0154	0.0077
Rising Stage	23-24	 	1	0.0147	0.0190	0.0160	0:0092
Falling Stage	24-25	 	 	0.0046	!!	1	
	-						
Average Ratio		0.0284	0.0138	0.0077	0.0163	0.0134	0.0079

Table 4-14

Time lag, in hours, between the minimum and maximum stages of the Hampstead Harbor and the corresponding minimum and maximum water levels in observation wells WP-1, WP-2, and WP-3 Shore Realty Superfund Site

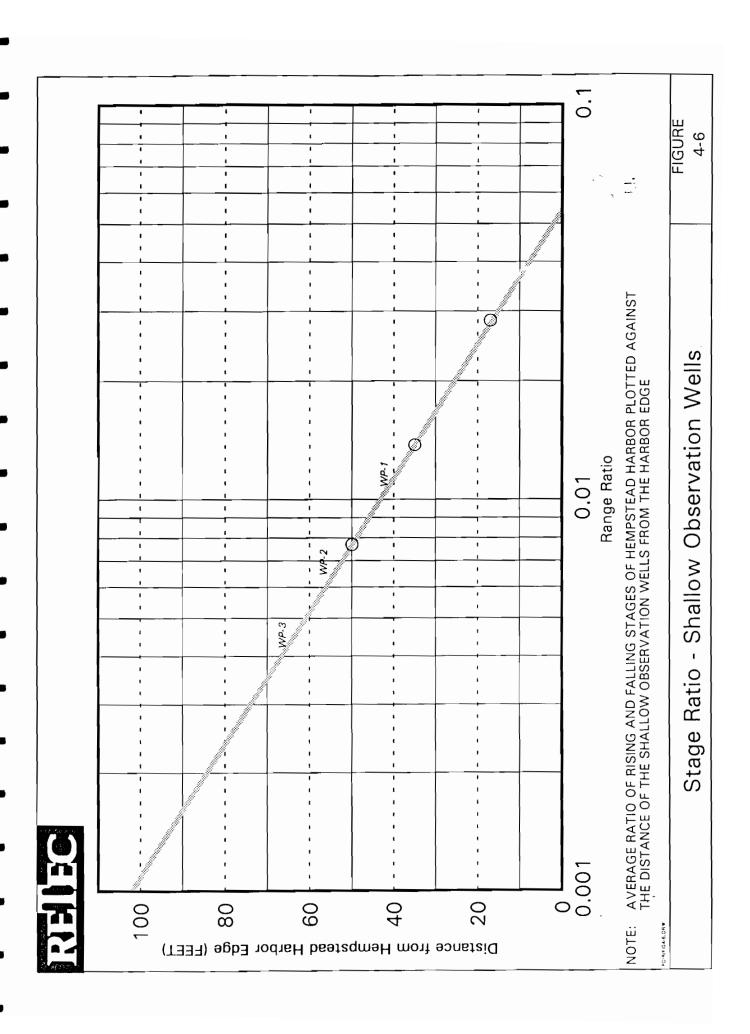
Glenwood Landing, New York

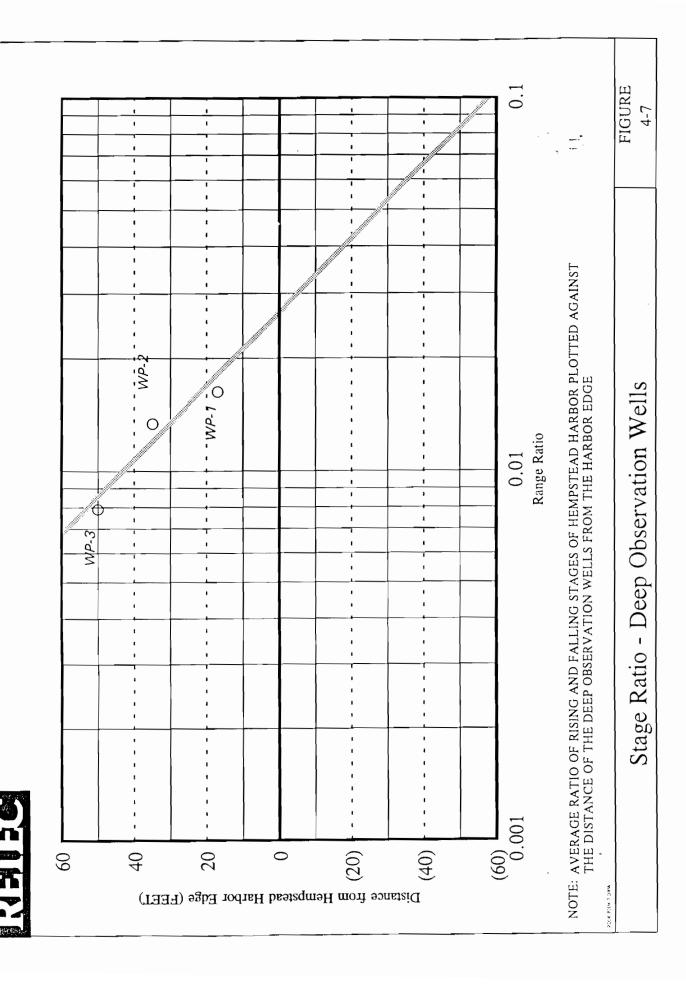
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of cyclic		Wells	
fluctuations	WP-1	WP-2	W

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WP-3

Minimum 1	1 1		1	1		
Maximum 2	1 1		1	3.17	3.17	1
Minimum 3	1.38	2.06	1	2.17	4.00	3.33
Maximum 4	1	3.43		3.67	3.83	1
Minimum 5	2.06	2.06	-	2.83	3.17	2.83
Maximum 6	1	2.06		2.67	3.17	4.17
Minimum 7	2.06	2.06	1	2.67	2.50	4.33
Maximum 8	1	2.74	4.11	2.83	3.17	3.50
Minimum 9	1	2.74	4.11	2.67	2.83	3.17
Maximum 10	0	2.06	3.43	2.83	3.00	3.83
Minimum 1	1	2.05	4.11	3.17	2.67	4.50
Maximum 12		2.06	3.43	2.83	3.33	4.17
Minimum 1	13 2.06	2.06	4.11	2.67	3.00	4.67
Maximum 1	14 2.06	2.06	3.43	2.17	2.83	3.00
Minimum 15		2.06	4.12	2.67	3.33	4.17
Maximum 1	3.43	3.43	4.11	2.67	3.33	4.17
Minimum 1	17 2.74	2.74	3.43	3.33	3.33	3.33
Maximum 1	18 2.06	2.06	4.12	2.67	2.50	4.00
Minimum 1	19 2.05	2.05	4.11	2.67	2.83	4.00
Maximum 20	0	2.06	1	3.00	3.17	3.67
Minimum 21		3.43	1 1	3.00	3.50	4.67
Maximum 22	2	2.05	1 -	2.67	2.83	2.83
Minimum 23	3	2.06	1	2.17	2.17	2.67
Maximum 24	1 1	2.74		2.67	2.83	3.50
Average Time Lag	Lag 2.18	2.37	3.89	2.78	3.07	3.74





and Level C are composed of medium and fine sand, silt and clay in the vicinity of WP-1, WP-2, and WP-3. McClymonds and Franke (1972) reported that the horizontal hydraulic conductivity of the Upper Glacial Aquifer when composed of medium, fine, and very fine sand, and sand with silt or clay layers, ranges from 400 gpd/ft² (53.5 ft/day) to 1,800 gpd/ft² (240 ft/day). Therefore, the determined horizontal hydraulic conductivities compare very well with the lower end of the range obtained by McClymonds and Franke (1972).

4.3.2.2 Time-Lag Method

The same data collected for the stage-ratio analysis can also be analyzed using the time-lag method, however, Ferris (1963) implies, that this methodology is less reliable than the stage-ratio method and gives a number of explanations including differences in the effective screen resistance of the observation wells could tend to distort observations of the timing of maximum and minimum water levels. This rationale could explain the large range in the measured lag of maxima and minima listed in Table 4-14.

In the time-lag method, the time lag is determined between the minimum and maximum of the tidal fluctuations and the corresponding minimum and maximum fluctuations in the wellpoints. The averages were calculated for the time lag of the minima and maxima for each wellpoint, and then plotted against the distances of the observation wells from the edge of Hempstead Harbor. The slope of the line through these plotted values (x/t) is determined. The transmissivity was calculated as follows:

$$T = \frac{0.60 \ x^2 \ S \ t_o}{t^2} \qquad where;$$

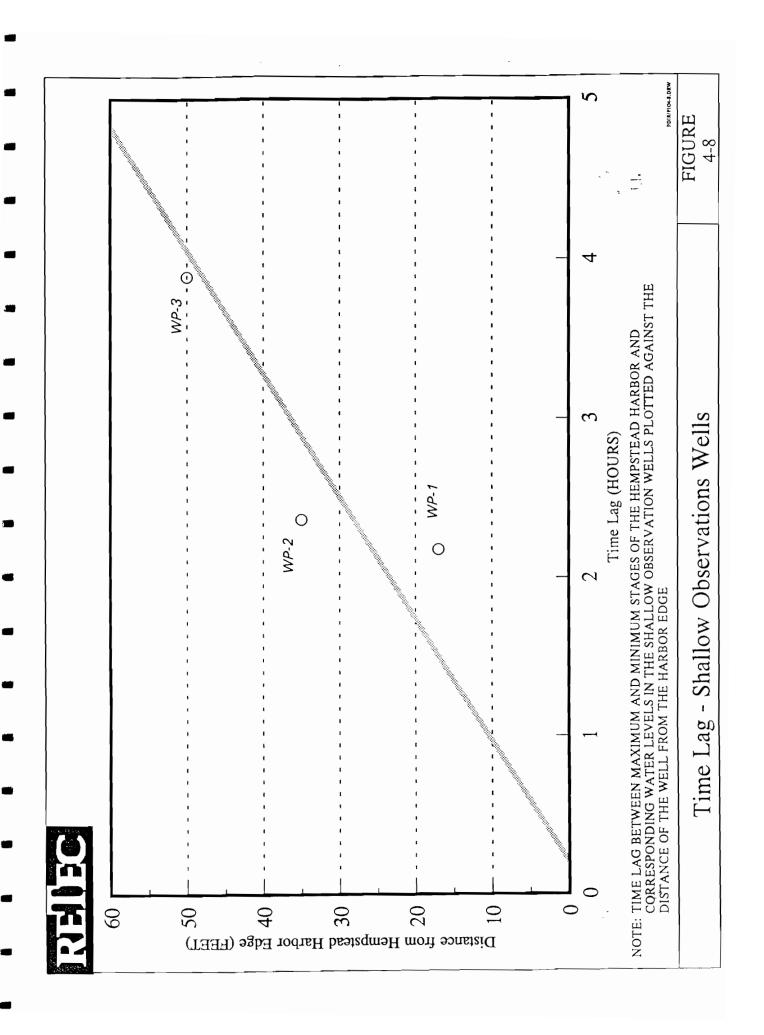
T = Transmissivity,

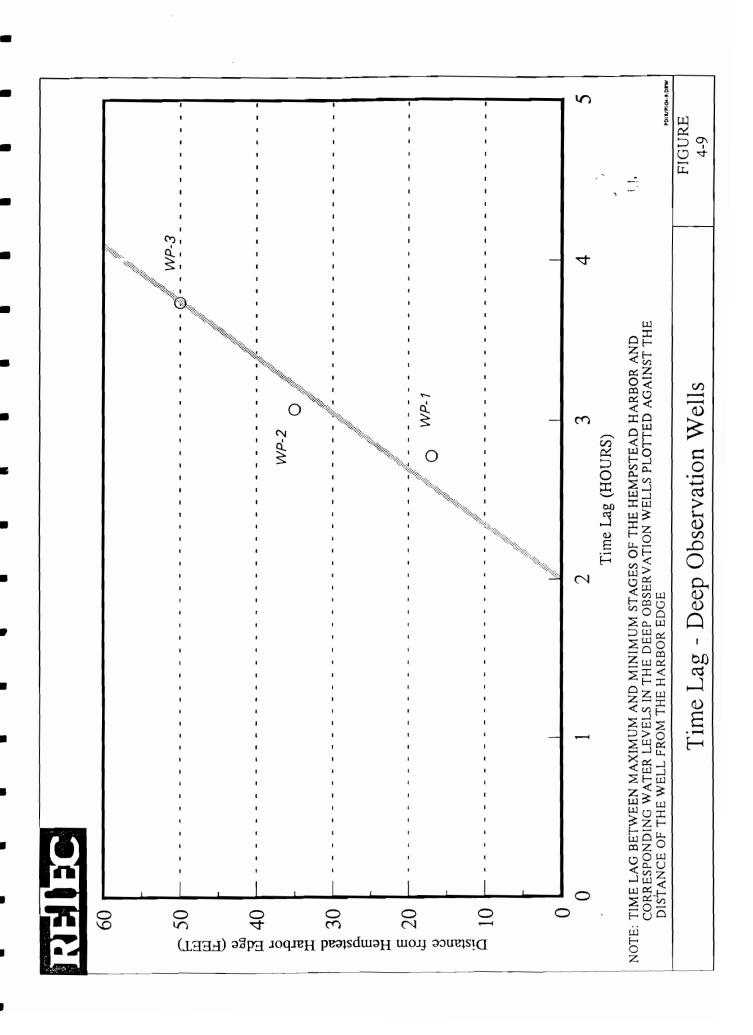
 $\frac{x}{t}$ = Slope of the Distance – Time Lag Plot,

S = Storativity,

t_o = Periodicity of Tidal Fluctuation

The average values of the time lag, t, are plotted against the distance of the wellpoints from the harbor's edge, x, in Figures 4-8 and 4-9. The slope of the line passing through the plotted values is x/t, where: x = 13 feet, t = 1 hour for the shallow wellpoints, and x = 17 feet, t = 0.6 hours for the deep wells. Substituting these values into the above equation with





 $t_o = 0.51$ days, and S = 0.10 (Getzen, 1977), gave a transmissivity of 400 ft²/day (3,000 gpd/ft) for Level B, and 1,900 ft²/day (14,150 gpd/ft) for Level C. Although the transmissivity value determined for Level B is the same as that obtained by the stage ratio method, the time-lag method is not considered as reliable as the stage-ratio method, as indicated above. Therefore, the transmissivity and horizontal hydraulic conductivity of Level B and Level C are considered to be those determined by the stage-ratio method.

4.3.3 Summary

The horizontal hydraulic conductivity as determined by the stage-ratio method is approximately 2 X 10⁻² cm/sec. The values for vertical hydraulic conductivity for the samples collected from the shelby tube were 2.3 X 10⁻² and 2.0 X 10⁻⁶ cm/sec. Remolded samples of sand and silty sand typically decrease in hydraulic conductivity due to the compaction that takes place as the soil is put into a plastic sleeve. It is also typical for the driving action of the split spoon to compress a soil sample. For this reason, the 2.0 X 10⁻⁶ cm/sec result seems unrepresentative of the site based upon the known background data for these formations on Long Island and the results of the time-lag stage-ratio test. Neglecting this sample, the laboratory hydraulic conductivity analyses support the results of the stage-ratio method.

5.0 NATURE AND EXTENT OF CONTAMINATION

The nature and extent of contamination at the Site was identified and described during the RI/FS and refined during the PDI. The stratigraphy of the Site is divided into four levels related to the water table. Level A is greater than five feet above the water table. Level B is five feet above the water table to three feet below the water table. Level C is three to fifteen feet below the water table, and Level D is greater than fifteen feet below the water table.

5.1 RESULTS OF SOILS INVESTIGATIONS

The analytical results from the RI were divided into six groups: CVOCs, non-chlorinated volatile organics (primarily ethylbenzene, toluene and xylenes), PAHs, phthalates, phenols, and total metals. Three of these six groups - PAHs, phthalates, and phenols - are subsets of the semivolatile organic compounds (SVOCs). Table 5-1 presents a summary of the various compounds identified in Zone A soils during the RI & PDI. Table 5-2 presents a summary of the various compounds identified in Zone B soils during the RI & PDI. Figure 5-1 is a summary of the results from the RI and PDI. The concentrations shown are for CVOCs, BTEX, and SVOCs.

5.1.1 Vadose Zone (Level A) Soils

Only a portion of the site has vadose zone soils: the tank farm, berms, and the northeastern portion of the access road adjacent to the Marina. The soils in this zone are moderately impacted. Two of the three samples collected from this zone during the RI were impacted by non-chlorinated volatile organics constituents ranging from 0.077 to 274.9 ppm. CVOCs detected during the RI and PDI include:

- methylene chloride;
- 1,1-dichloroethylene;
- 1,1-dichloroethane;
- 1,2-dichloroethylene:
- chloroform;
- vinyl chloride;
- 1,1,1-trichloroethane;
- 1,1,2-trichloroethane;

Table 5-1
Summary Of Compounds Identified in Zone A Soil Samples During the RI and PDI (mg/kg)
Shore Realty Superfund Site

Glenwood Landing, New York

	2007 1000 100 N		Zone A Soil		11 1 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2
CHEMICAL NAME	No. of	No. of	Range of	Geometric	‡ ucl
<u></u>	Detects	Samples	Levels	UCL	+ ucr
Volatile Organic Compounds					
Vinyl Chloride	0	29	ND	ND	ND
Methylene Chloride	2	29	0.006 - 0.37	10	5
Acetone	1 1	29	1	10	9
1,1-Dichloroethane	1	29	0.064) 9	5
trans-1,2-Dichloroethylene	1	26	0.1	10	5
1,1,1-Trichloroethane] 3	29	0.074 - 7.6	10	5
1,2-Dichloropropane		29	ND	ND	ND
Trichloroethylene	1 1	29	0.009	9	5
1,1,2-Trichloroethane	1 1	29	0.011	9	5
Benzene	1 2	29	0.005 - 28	10	6 .
Tetrachloroethylene		29	ND	ND	ND
Toluene	26	29	0.002 - 2,600	44	492
Ethylbenzene	22	29	0.001 - 1,300	28	188
Xylenes	24	29	0.01 - 8,400	48	1,183
	No. of	No. of		Average	Geometric Avg.
	Detects	Samples	Range of Levels	Level	Level
Advantage America W. Horney (1)	Detects	Samples	Develo	Live	Deva
Polynuclear Aromatic Hydrocarbons(1) Benzoic acid	0	4	ND	ND	ND
	5	5	0.089 - 12	4.34	1.99
Napthalene	5	5	0.089 - 12	5.13	2.46
2-Methylnapthalene	1 1	4	0.13 - 13	0.53	0.24
Acenaphthene	2 0	5	ND	1	
Acenaphthylene			1	ND 0.17	ND
Dibenzofuran	2	4	0.12 - 0.22	0.17	0.16
Fluorene	2	4	0.23 - 0.3	0.61	0.39
Phenanthrene	2	4	0.33 - 0.45	0.69	0.49
Anthracene	1 1	5	0.048	0.63	0.32
Fluoranthene	2	5	0.08 - 0.23	0.63	0.36
Pyrene	2	4	0.024 - 0.17	0.55	0.26
Benzo(a)anthracene	1 1	5	0.076	0.63	0.35
Chrysene	1	5	0.12	0.64	0.39
Benzo(b)fluoranthene	2	5	0.036 - 0.14	0.61	0.28
Benzo(k)fluoranthene	0	5	ND	ND	ND
Benzo(a)pyrene	2	5	0.025 - 0.065	0.59	0.22
Indeno(1,2,3-cd)pyrene	1	5	0.033	0.62	0.30
Dibenzo(a,h)anthracene	0	5	ND	ND	ND
Benzo(g,h,i)perylene	1	5	0.033	0.62	0.30
Phthalates(1)					
Di-n-butylphthalate	3	5	0.037 - 1.5	1.00	0.59
Butylbenzylphthalate	2	3	0.033 - 0.047	0.10	0.07
bis(2-Ethylhexyl)phthalate	4	5	0.74 - 12	4.55	3.41
Di-n-octylphthalate	2	. 5	0.08 - 3.3	1.11	0.45
Phenols(1)					
2-Methylphenol	ī	4	1.5	0.53	0.33
4-Methylphenol	0	4	ND	ND	ND
			_		
2,4-Dimethylphenol	0	4	ND	ND	ND

Notes

Range and average values reported in this table include all constituents detected in samples including estimated values and constituents detected in method blanks.

(1) - Polynuclear Aromatic Hydrocarbons, Phthalates, and Phenols statistics are based on RI data only.

ND - Not Detected

NA - Not Applicable

UCL - Upper 95% Confidence Limit

TAB5-1.WK1 15-Jul-93

Table 5-2 Summary Of Compounds Identified in Zone B Soil Samples During the RI and PDI (mg/kg) Shore Realty Superfund Site

Glenwood Landing, New York

25 July 1980 Committee Com	14/1/2		Zone B Soil		11.00
CHEMICAL NAME	No. of	No. of	Range of	Geometric	
	Detects	Samples	Levels	UCL	UCL
/olatile Organic Compounds					
Vinyl Chloride	0	16	ND	ND	ND
Methylene Chloride	3	16	0.005 - 20	18	5
Acetone	1	16	0.10	ND	ND
1,1-Dichloroethane	0	16	ND	ND	ND
trans-1,2-Dichloroethylene	0	12	ND	ND	ND
1,1,1-Trichloroethane	0	16	ND	ND	ND
1,2-Dichloropropane	0	16	ND	ND	ND
Trichloroethylene	0	16	ND	ND	ND
1,1,2-Trichloroethane	0	16	ND	ND	ND
Benzene	o l	16	ND	ND	ND
Tetrachloroethylene	1 1	16	0.004	11	1
Toluene	8	16	0.009 - 140	24	31
Ethylbenzene	8	16	0.003 - 87	37	31
Xylenes	12	16	0.007 - 530	79	217
Aylenes					
	No. of	No. of	Range of	Average	Geometric Avg
	Detects	Samples	Levels	Level	Level
olynuclear Aromatic Hydrocarbons(1)					
Benzoic acid	0	1	ND	ND	ND
Napthalene	1	2	6.2	3.19	1.04
2-Methylnapthalene	1 1	2	9.6	4.89	1.30
Acenaphthene	0	2	ND	ND	ND
Acenaphthylene	0	2	ND	ND	ND
Dibenzofuran	1	2	0.22	0.22	0.22
Fluorene	1	2	0.3	0.24	0.23
Phenanthrene	1 .	2	0.41	0.29	0.27
Anthracene	0	2	ND	ND	ND
Fluoranthene	0	2	ND	ND	ND
Pyrene	1	2	0.024	0.10	0.06
Benzo(a)anthracene	0	2	ND	ND	ND
Chrysene	0	2	ND	ND	ND
Benzo(b) fluoranthene	0	2	ND	ND	ND
Benzo(k)fluoranthene	0	2	ND	ND	ND
Benzo(a)pyrene	0	2	ND	ND	ND
Indeno(1,2,3-cd)pyrene	0	2	ND	ND	ND
Dibenzo(a,h)anthracene	0	2	ND	ND	ND
Benzo(g,h,i)perylene	0	2	ND	ND	ND
hthalates(1)					
Di-n-butylphthalate	1	2	0.16	0.17	0.17
Butylbenzylphthalate	0	2	ND	ND	ND
bis(2-Ethylhexyl)phthalate	i	2	0.74	0.46	0.36
Di-n-octylphthalate	0	2	ND	ND	ND
	L		J		
Phenols(1)			ND	ND.	ND
2-Methylphenol	0	2	ND	ND	ND
4-Methylphenol	0	2	ND	ND	ND
2,4-Dimethylphenol	0	2	ND	ND	ND
2,4-Dinitrophenol	0	2	ND	ND	ND

Notes:

Range and average values reported in this table include all constituents detected in samples including estimated values and constituents detected in method blanks.

15-Jul-93

(1) - Polynuclear Aromatic Hydrocarbons, Phthalates, and Phenols statistics are based on RI data only.

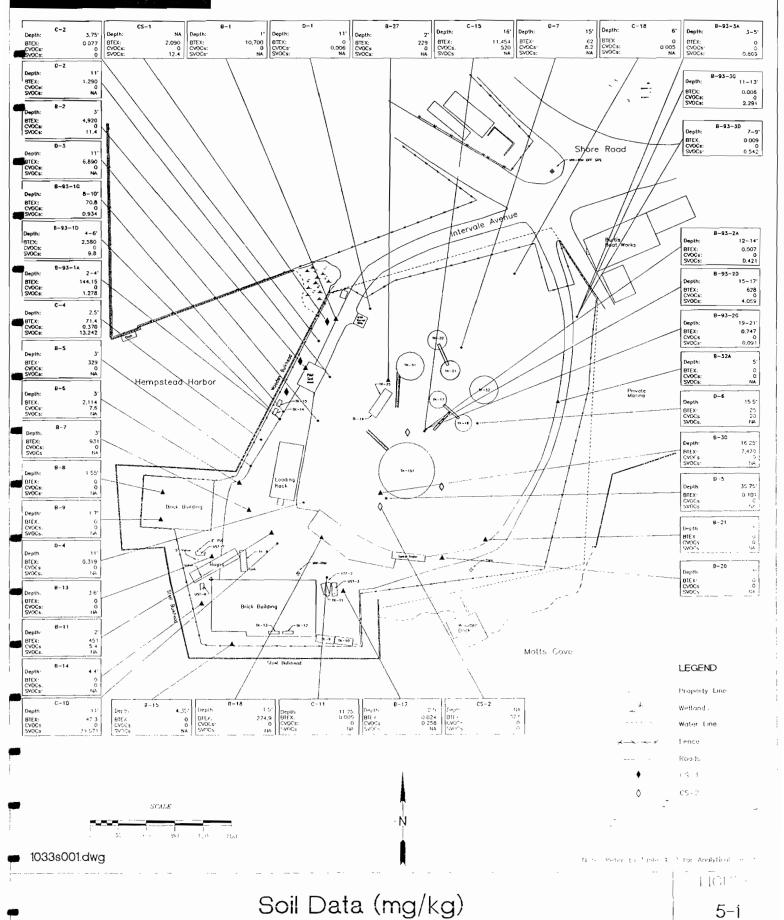
ND - Not Detected

NA - Not Applicable

UCL - Upper 95% Confidence Limit

TAB5-2.WK1

REFEC



5-1

- trichloroethylene; and
- tetrachloroethylene.

No CVOCs or SVOCs were detected in Level A soils during the RI. Total metals analysis was performed on samples from this level, and four total metals were detected ranging from 0.0072 ppm of copper to 0.224 ppm of zinc.

Two samples were rellected from this level during the PDI. One sample detected non-chlorinated volatile organics at 0.009 ppm. No CVOCs, phthalates, or phenols were detected. Both samples detected PAHs at concentrations ranging from 0.382 ppm to 0.689 ppm. The samples were analyzed only for total iron from the metals group.

5.1.2 Saturated Soils

The saturated soils at the Site were divided into three levels during the RI.

5.1.2.1 Level B

Level B had the most detections during the RI, and the most samples, since it contains the water table surface where most of the contamination is expected to be present. Seven CVOCs were detected in six samples ranging from 0.009 ppm of TCE to 20 ppm of methylene chloride. Eighteen soil samples detected non-chlorinated volatiles in the form of ethlybenzene, toluene and xylenes ranging from 0.024 ppm to 10,700 ppm. Fifteen PAHs were detected in five samples ranging from 0.025 ppm of benzo(a)pyrene to 13 ppm of 2-methyl naphthalene. Phthalates were detected in four samples ranging from 0.033 ppm of butylbenzylphthalate to 12 ppm of bis(2-ethylhexyl)phthalate. One sample identified 2-methyl phenol at a concentration of 1.5 ppm. Total metals were detected in all of the samples, but in no pattern that would indicate that site operations have increased total metals concentrations.

Five samples were collected from Level B during the PDI. Non-chlorinated volatile compounds, CVOCs, PAHs, phthalates, and phenols were detected at similar levels in similar areas of the Site as identified in the RI. There were no SVOC results from the tank farm during the RI. However, samples were analyzed for SVOCs from the center of the tank farm during the PDI. PAHs in the two samples from this area ranged from non-detect to 3.799 ppm. Phthalates in the two samples ranged from 0.270 ppm to 0.421 ppm. No phenols were detected in the two samples.

5.1.2.2 Level C

Í

Ten samples were collected from Level C during the RI. CVOCs were detected in the form of methylene chloride in two samples at concentrations from 0.006 ppm to 0.37 ppm. Nine samples detected non-chlorinated volatiles in the form of ethlybenzene, toluene and xylenes ranging from 0.005 to 71.4 ppm. Acetone was also detected in one sample at a concentration of 1 ppm. Two PAHs were detected in one sample ranging from 0.089 ppm of naphthalene to 0.13 ppm of 2-methyl naphthalene. Two phenolic compounds were detected in one sample ranging from 0.037 ppm of di-n-butylphthalate to 1.6 ppm of bis(2-ethylhexyl)phthalate. No total metals were detected.

Two samples were collected from Level C during the PDI. The concentrations of the various constituents was similar to those found during the RI along the bulkhead. However, a sample was collected from this level in the center of the tank farm. Chlorinated compounds were detected in the form of methylene chloride at 0.016 ppm. Non-chlorinated compounds were detected in the form of ethlybenzene, toluene and xylenes at 0.747 ppm. PAHs and phthalates were detected at 0.049 ppm and 0.042 ppm, respectively. No phenols were detected.

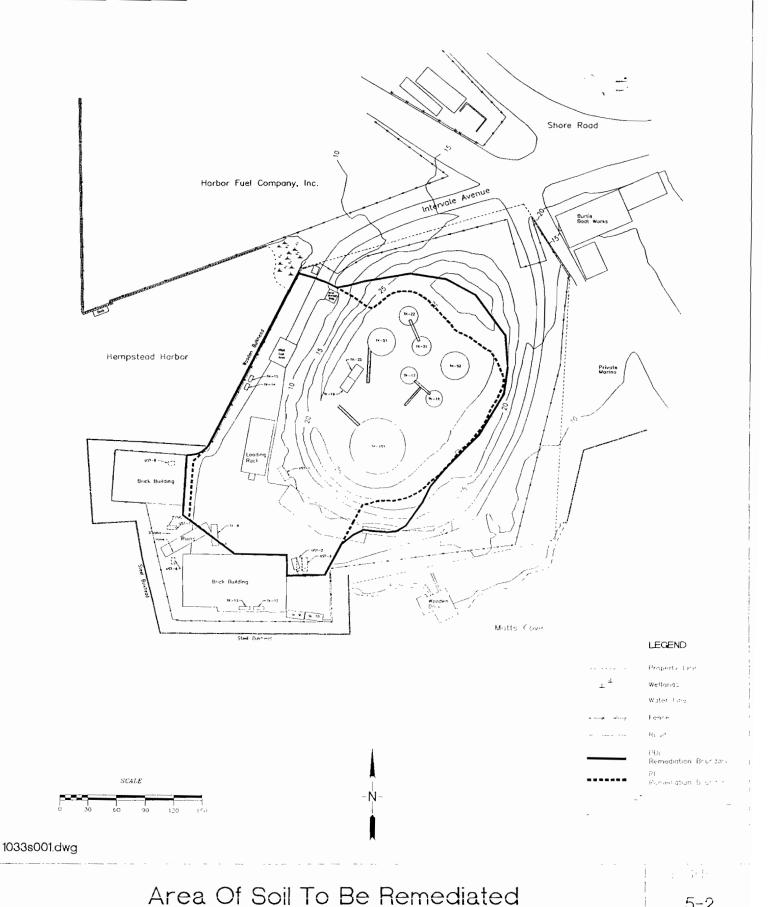
5.1.2.3 Level D

Five soil samples were collected from the Level D during the RI. The only CVOC detected was PCE at 0.004 ppm. Ethlybenzene, toulene, and xylenes was detected in all of the soil samples ranging from 0.002 ppm to 2.87 ppm. None of the samples were analyzed for PAHs, phthalates, phenols or total metals. No samples were collected from this level during the PDI.

5.1.3 Limits of Soil to be Remediated

The soils investigation conducted during the PDI refined the limits of soil to be remediated presented in Figure 5-2. This area is expanded from the RI due to the results of the PDI soil gas survey which was able to collect samples from just above the water table in the areas of higher elevation at the site. The area identified in the RI was expanded northward and directly southward of the tank farm.

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5.2 RESULTS OF SEDIMENTS INVESTIGATION

The analytical results from the RI were divided into the same six groups as those for the soil: CVOCs, non-chlorinated volatile organics, PAHs, phthalates, phenols, and total metals. Table 5-1 presents a summary of the various compounds identified in sediments during the RI.

The chlorinated suite had only two detectable constituents, methylene chloride and PCE. Methylene chloride was considered a laboratory-introduced contaminant and PCE was an estimated value. The non-chlorinated suite included ethlybenzene, toluene, xylenes and acetone. Ethylbenzene, toluene and xylenes were detected in four samples ranging from 0.013 ppm toluene to 1.4 ppm xylenes. Five samples detected acetone ranging from 0.017 ppm to 0.051 ppm. PAHs were the most frequently detected organics in the sediment samples. Thirteen of the samples detected PAHs ranging from an estimated value of 0.03 ppm fluorene to 1.2 ppm fluoranthene. Sixteen PAH compounds were detected in the sediment samples. Phthalates were detected in all of the sediment samples ranging from 0.038 ppm of butylbenzylphthalate to 8.1 ppm of bis(2-ethylhexyl)phthalate. Phenols were not detected in any sediment samples. Total metals were detected in all of the sediment samples.

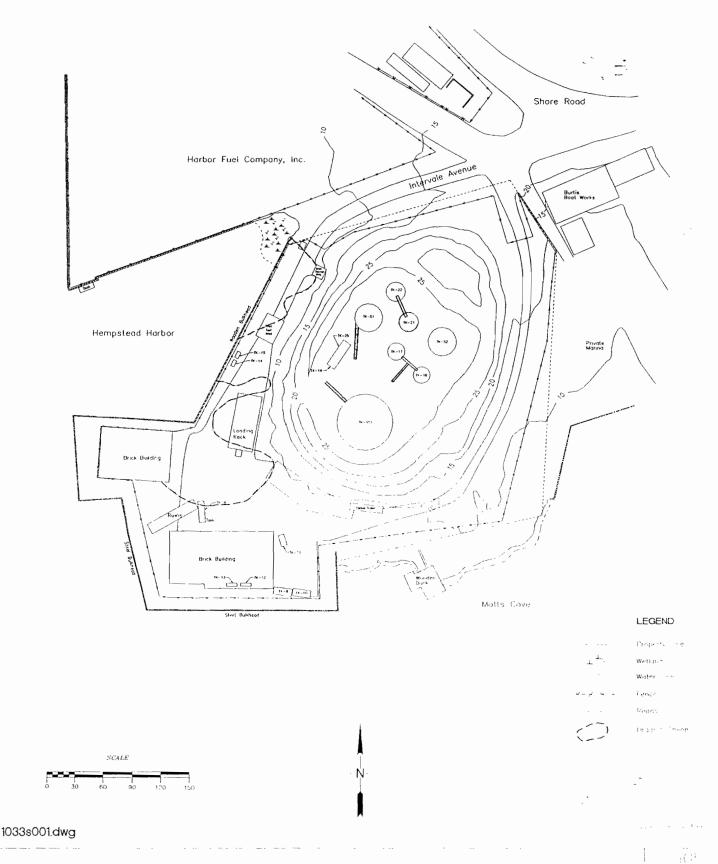
5.3 WATER TABLE OBSERVATIONS

During the RI and PDI, the water table fluctuations were observed and the organic sheen on the water table was noted. Figure 5-3 presents the location of the organic sheen observed on the water table in November of 1990 during the RI. The water table at the Site propagates in a sinusoidal fashion at the Site due to the fluctuation in the tides. As the tide rises and falls, it sends this propagation through the Site from the south and west. This propagation, in combination with a majority of the contaminants (by mass) floating on the water table, explain the reason for the highest result of contamination being present in Level B. Section 4.0 discusses the water fluctuation data more thoroughly.

5.4 GROUNDWATER SAMPLES

The monitoring wells at the Site were divided into three groups during the RI; WT, SW, and DW. The summary of the compounds identified in Zone A groundwater and water table samples during the RI & PDI is presented in Table 5-3. Table 5-4 summarizes the compounds

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Location Of Organic Sheen

5-3

TABLE 5-3
Summary of Compounds Identified in Zone A Groundwater Samples during the RI and PDI (ug/l)
Shore Realty Superfund Site

Glenwood Landing, New York

	Oleliwood Laik	1116, 11011 10	· K		
86 (A)		2007	Zone A Ground	lwater	· .
CHEMICAL NAME	No. Of	No. Of	Range Of	Geometric	
	Detects	Samples	Levels	UCL -	UCL
Volatile Organic Compounds					
Vinyl Chloride	0	19	ND	ND	ND
Methylene Chloride	6	19	1 - 34	10	832
Acetone		19	66	ND	ND
1,1-Dichloroethane	4	19	1 - 4	8	830
trans-1,2-Dichloroethylene		14	7 - 10	13	1,147
1,1-Dichloroethylene		19	1	8	830
1,1,1-Trichloroethane	5	19	2 - 20	9	832
1,2-Dichloropropane	0	19	ND	ND	ND
	3	19	4 - 13	9	831
Trichloroethylene		19	ND	_	
1,1,2-Trichloroethane				ND	ND
Benzene	1	19	370	9	837
Tetrachloroethylene	9	19	1 - 280	15	850
Toluene	8	19	2 - 350,000	29	84,994
Ethylbenzene	1	19	4,800	11	1,263
Xylenes	2	19	16,000 - 30,000	15	6,095
		No. Of	Range Of	Average	Geometric A
	Detects	Samples	Levels	Level	Level
olynuclear Aromatic Hydrocarbons(1)					
Benzoic acid	0	6	ND	ND	ND
Napthalene	0	7	ND	ND	ND
2-Methylnapthalene	0	7	ND	ND	ND
Acenaphthene	o l	6	ND	ND	ND
Acenaphthylene	0	7	ND	ND	ND
Dibenzofuran	0	6	ND	ND	ND
Fluorene	0	6	ND	ND	ND
Phenanthrene	0	6	ND	ND	ND
Anthracene	0	6	ND	ND	ND
Fluoranthene	0	7	ND	ND	ND
Pyrene	0	6	ND	ND	ND
Benzo(a)anthracene	0	7	ND	ND	ND
	0	7	1		
Chrysene	0	7	ND	ND	ND
Benzo(b)fluoranthene	1	-	ND	ND	ND
Benzo(k)fluoranthene	0	7	ND	ND	ND
Benzo(a)pyrene	0	7	ND	ND	ND
Indeno(1,2,3-cd)pyrene	0	7	ND	ND	ND
Dibenzo(a,h)anthracene	0	7	ND	ND	ND
Benzo(g,h,i)perylene	0	7	ND	ND	ND
ththalates(1)					
Di-n-butylphthalate	4	7	7 - 73	24	15
Butylbenzylphthalate	0	6	ND	ND	ND
bis(2-Ethylhexyl)phthalate	4	7	8 - 20	14	13
Di-n-octylphthalate	0	7	8 - 20 ND	ND	ND
z cotypinment			1,10		1
henols(1)					
2-Methylphenol	0	6	ND	ND	ND
4-Methylphenol	0	6	ND	ND	ND
1 / D' 1 L. L.			N.D.		

ND

ND

ND

ND

Notes

Range and average values reported in this table include all constituents detected in samples including estimated values and constituents detected in method blanks.

(I) - Polynuclear Aromatic Hydrocarbons, Phthalates, and Phenols statistics are based on RI data only.

ND - Not Detected

2,4-Dimethylphenol

2,4-Dinitrophenol

NA - Not Applicable

UCL - Upper 95% Confidence Limit

ND

ND

TABLE 5-4
Summary of Compounds Identified in Zone B Groundwater Samples during the RI and PDI (ug/l)

Shore Realty Superfund Site

Glenwood Landing, New York

VICTOR OF THE WILL REPORT USE OF THE	1. 1. 1.74		Groundwater	3 S. S. W S.	
CHEMICAL NAME	No. Of	No. Of	Range Of	Geometric.	
	Detects	Samples	Levels	UCL □	UCL
Volatile Organic Compounds					
Vinyl Chloride	0	43	ND	ND	ND
Methylene Chloride	14	43	1 - 120	5	363
Acetone	1	43	66	9	722
1,1-Dichloroethane	12	43	1 - 8	5	360
trans-1,2-Dichloroethylene	6	30	1 - 77	6	520
1,1-Dichloroethylene	3	43	1 - 2	5	360
1,1,1-Trichloroethane	16	43	1 - 20	6	361
1,2-Dichloropropane	0	43	ND	ND	ND
Trichloroethylene	16	43	1 - 13	5	361
1,1,2-Trichloroethane	0	43	ND	ND	ND
Benzene	1	43	370	5	363
Tetrachloroethylene	18	43	1 - 280	8	371
Toluene	9	43	2 - 350,000	9	37,073
Ethylbenzene	2	43	1,500 - 4,800	6	592
Xylenes	3	_ 43	4,500 - 30,000	7	2,766
	No. Of	No. Of	Range Of	Average	Geometric Avg.
	Detects	Samples	Levels	Level	Level
Polynuclear Aromatic Hydrocarbons(1)					
Benzoic acid	0	6	ND	ND	ND
Napthalene	0	7	ND	ND	ND
2-Methylnapthalene	0	7	ND	ND	ND
Acenaphthene	0	6	ND	ND	ND
Acenaphthylene	0	7	ND	ND	ND
Dibenzofuran	0	6	ND	ND	ND
Fluorene	0	6	ND	ND	ND
Phenanthrene	0	6	ND	ND	ND
Anthracene	0	6	ND	ND	ND
Fluoranthene	0	7	ND	ND	ND
Pyrene	0	6	ND	ND	ND
Benzo(a)anthracene	0	7	ND	ND	ND
Chrysene	0	7	ND	ND	ND
Benzo(b)fluoranthene	0	7	ND	ND	ND
Benzo(k)fluoranthene	0	7	ND	ND	ND
Benzo(a)pyrene	0	7	ND	ND	ND
Indeno(1,2,3-cd)pyrene	0	7	ND	ND	ND
Dibenzo(a,h)anthracene	0	7	ND	ND	ND
Benzo(g,h,i)perylene	0	7	ND	ND	ND
DLat -1-4(f)					
Phthalates(1)	4	7	7 72	24	1.6
Di-n-butylphthalate	4	7	7 - 73 ND	24 ND	15 ND
Butylbenzylphthalate	0 4	6 7	8 - 20	ND	13
bis(2-Ethylhexyl)phthalate Di-n-octylphthalate	0	7	8 - 20 ND	14 ND	ND
		·	1.0		1,0
Phenols(1)					1
2-Methylphenol	0	6	ND	ND	ND
4-Methylphenol	0	6	ND	ND	ND
2,4-Dimethylphenol	0	6	ND	ND	ND
2,4-Dinitrophenol	0	6	ND	ND	ND

Notes:

Range and average values reported in this table include all constituents detected in samples including estimated values and constituents detected in method blanks.

- (1) Polynuclear Aromatic Hydrocarbons, Phthalates, and Phenols statistics are based on RI data only.
- ND Not Detected
- NA Not Applicable

UCL - Upper 95% Confidence Limit

TAB5-4.WK1 15-Jul-93

identified in Zone B groundwater and water table samples during the RI & PDI. Table 5-5 presents the Zone A water table results within the organic sheen from the RI & PDI. Figure 5-4 presents a summary of CVOCs, BTEX, and SVOCs data from both the RI and PDI.

5.4.1 Water Table Wells

CVOCs were detected in seven of the nine WT wells during the RI. Eight different CVOCs were detected in all. Detections ranged from 2 ppb of methylene chloride, PCE, and trans-1,2-dichloroethylene to 970 ppb of methylene chloride. Five non-chlorinated volatile compounds were detected in six wells. Ethylbenzene, toluene, and xylenes were the most predominant, but there were also detections of benzene and acetone. Ethlybenzene, toluene and xylenes ranged from 806 ppb to 384,800 ppb. Two wells had PAHs detected ranging from 9 ppb of 2-methylnaphthalene to 120 ppb of benzoic acid. Four of the wells had two phthalates ranging from 6 ppb of butylbenzylphthalate to 130 ppb of di-n-butylphthalate. Two of the wells contained three phenolic compounds ranging from 47 ppb of 4-methylphenol to 490 ppb of 2-methylphenol. Twelve unfiltered total metals were detected in all of the wells ranging from less than 1 ppb of mercury to 76,200 ppb of iron.

The PDI detected lower levels of CVOCs, non-chlorinated volatile compounds and SVOCs in all but one of the wells. WT-14 had 60 ppb of non-chlorinated volatile compounds, no non-chlorinated volatile compounds were detected during the RI in this well. Samples from wells WT-6, WT-13 and WT-14 were analyzed for SVOCs during the PDI although they were not analyzed for SVOCs during the RI. PAHs were detected in WT-6 at 97 ppb. Phthalates were detected in WT-13 at 10 ppb. Phenols were detected in WT-6 at 444 ppb. A well was installed and screened across the water table, in the center of the tank farm. This well detected 180 ppb of CVOCs, 11,500 ppb of non-chlorinated volatile compounds, 75 ppb of PAHs, 1 ppb of phthalates and 27 ppb of phenols. No wells were present in this area prior to this well installation. The off-site well, WT-93-4, results will be reported in this July's monthly progress report.

5.4.2 Shallow Groundwater Wells

There are six SW wells at the Site. Seven SVOCs were detected in the six wells ranging from 1 ppb of methylene chloride, PCE, TCE, and 1,1-DCA to 280 ppb of PCE. Of the non-chlorinated volatile compounds, only toluene was detected at 2 ppb in the wells. No PAHs or phenolic compounds were detected. Two phthalates were detected in one well, 8 ppb of bis(2-ethylhexyl)phthalate and 73 ppb of di-n-butylphthalate. Twelve unfiltered total metals were

TABLE 5-5

Summary of Compounds Identified in Zone A Water Table Samples during the RI and PDI (ug/l) (Basis for Pump & Treat)

Shore Realty Superfund Site

Glenwood Landing, New York

14 V 2 3 V			Zone A Water Tab	ole :	
CHEMICAL NAME	No. Of	No. Of	Range Of	Geometric	L. Carajy.
	Detecs	Samples	Levels	UCL -	UCL
Volatile Organic Compounds				,	
Vinyl Chloride	1	12	12	113	3,575
Methylene Chloride	6	12	2 - 970	81	536
Acetone	6	12	15.0 - 2,000	121	744
1,1-Dichloroethane	3	12	6 - 19	80	1,867
trans-1,2-Dichloroethylene	3	9	25 - 31	176	2,539
1,1-Dichloroethylene	0	12	ND	ND	ND
1,1,1-Trichloroethane	3	12	6 - 23	79	1,867
1,2-Dichloropropane	1	12	5	73	1,866
Trichloroethylene	2	12	4 - 29	85	1,868
1,1,2-Trichloroethane	0	12	ND	ND	ND
Benzene	7	12	6 - 270	48	250
Tetrachloroethylene	2	12	10 - 430	95	1,894
Toluene	10	12	330 - 270,000	2,354	95,458
Ethylbenzene	9	12	160 - 5,600	444	4,153
Xylenes	_11	12	450 - 45,000	5,532	28,077
	No. Of	No. Of	Range Of	Average	Geometric Avg.
	Detecs	Samples	Levels	Level	Level
Polynuclear Aromatic Hydrocarbons(1)		•			
Benzoic acid	1	4	120	NA	NA
Napthalene	2	4	29 - 40	35	34
2-Methylnapthalene	2	4	9 - 11	10	10
Acenaphthene	0	4	ND	ND	ND
Acenaphthylenc	0	4	ND	ND	ND
Dibenzofuran	0	4	ND	ND	ND
Fluorene	0	4	ND	ND	ND
Phenanthrene	0	4	ND	ND	ND
Anthracene	0	4	ND	ND	ND
Fluoranthene	0	4	ND	ND	ND
Pyrene	0	4	ND	ND	ND
Benzo(a)anthracene	0	4	ND	ND	ND
Chrysene	0	4	ND	ND	ND
Benzo(b)fluoranthene	0	4	ND	ND	ND
Benzo(k)fluoranthene	0	4	ND	ND	ND
Benzo(a)pyrene	0	4	ND	ND	ND
Indeno(1,2,3-cd)pyrene	0	4	ND	ND	ND
Dibenzo(a,h)anthracene	0	4	ND	ND	ND
Benzo(g,h,i)perylene	0	4	ND	ND	ND
n d 1 d /d)					
Phthalates(1)	4	4	7 - 130	45	80
Di-n-butylphthalate Butylbenzylphthalate		4	6	NA NA	NA
bis(2-Ethylhexyl)phthalate	0	4	ND	NA ND	ND
Di-n-octylphthalate	0	4	ND ND	ND	ND
					1
Phenols(1)	2		100 400	205	221
2-Methylphenol	2	4	100 - 490	295	221
4-Methylphenol	2	4	47 - 130	89	78
2,4-Dimethylphenol	3	4	120 - 390	247	221
2,4-Dinitrophenol	0	4	ND	ND	ND

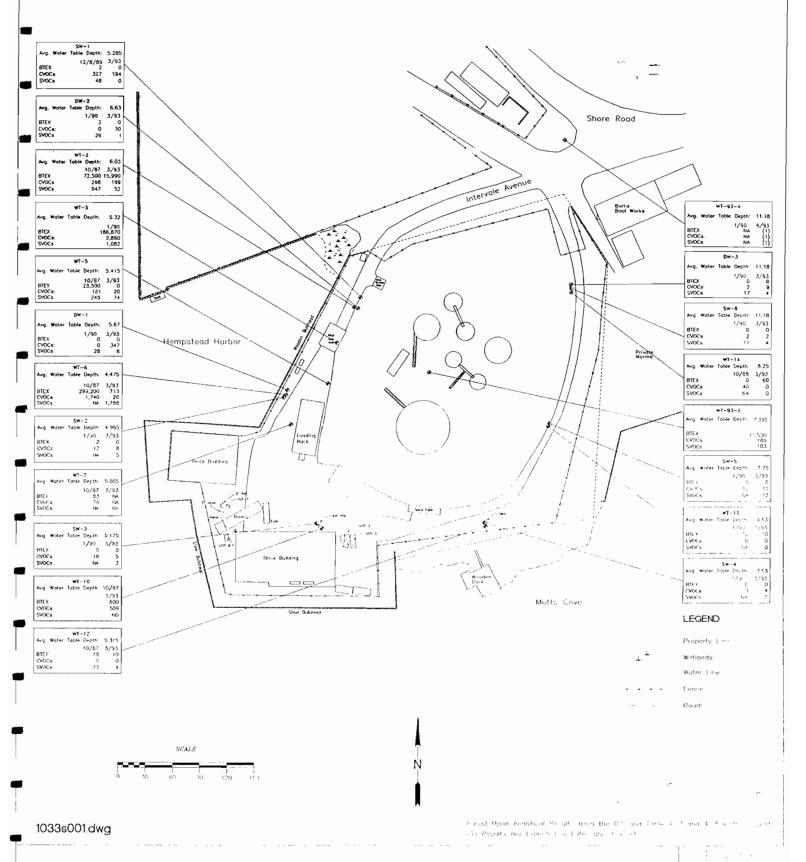
Notes

Range and average values reported in this table include all constituents detected in samples including estimated values and constituents detected in method blanks.

- (I) Polynuclear Aromatic Hydrocarbons, Phthalates, and Phenols statistics are based on RI data only.
- ND Not Detected
- NA Not Applicable

UCL - Upper 95% Confidence Limit

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Ground Water Data

detected in the six wells ranging from less than one ppb of mercury to 644 ppb of iron.

The shallow groundwater samples analyzed during the PDI detected similar or lower concentrations in all of the wells for each of the six compound categories. The samples from SW-2, SW-3, SW-4 and SW-5 were analyzed for SVOCs, whereas the samples from the same wells during the RI were not analyzed for SVOCs. Samples from SW-2, SW-3, and SW-4 detected only phthalates ranging from 1 ppb to 5 ppb.

5.4.3 Deep Groundwater Wells

There are three DW wells at the site. Only bis(2-ethylhexyl)phthalate and di-n-butyl phthalate were above state MCLs, 5 ppb. They were detected in the range of 7 ppb to 20 ppb. Six total metals were detected in the three wells ranging from 6.1 ppb of cadmium to 414 ppb of iron.

The deep groundwater samples analyzed during the PDI detected similar or lower concentrations in all of the wells for all of the six compound categories except CVOCs. Only DW-3 detected CVOCs during the RI and the PDI detected a similar concentration at this location. Wells DW-1 and DW-2, along the western portion of the site, detected CVOCs at 347 ppb and 30 ppb, respectively.

5.5 REPRESENTATIVE CONTAMINANTS

The RI identified twelve VOCs, fifteen PAHS, four phthalates, and one phenol. In soils, the PDI identified five VOCs, twelve PAHs, two phthalates and two phenols. Most of the CVOCs detections during the RI were single instances, and the PDI was limited and focused on filling in data gaps associated with engineering design parameters; therefore, the differences are minimal between these two investigations in this regard. The highest PAH detection was during the RI, and was 13 ppm of 2-methylnaphthalene and acenaphthene. These concentrations are minimal, and are susceptible to further reduction through the implementation of the *in situ* biodegradation component of the Remedy. (Ryan and Loehr, 1991)

The RI identified fourteen VOCs, three PAHs, two phthalates and three phenolic compounds in the groundwater samples from the water table wells. The PDI identified twelve VOCs, ten PAHs, three phthalates and six phenolic compounds. Generally, the PDI found the concentrations of these compounds were lower. All of the PAHs and phthalates identified during the RI and PDI are below their respective clean-up standard. The phenols as well as the other

constituents are degradable, and the remedy for the site should be effective at reducing their respective concentrations.

The RI identified eight VOCs and two phthalates in the groundwater at the site. The PDI identified seven of the same eight VOCs. The concentrations of these were generally the same as those identified during the RI. The PDI also identified one of the same phthalates as the RI at generally the same concentrations. Neither investigation identified PAHs or phenols in the groundwater. The discrepancies between the two investigations are negligible and do not effect the design of the remedy.

The ROD identified six VOCs as the primary contaminants of interest. Generally, VOC concentrations were found to be somewhat lower during the PDI than the RI. In addition, the ROD identifies PAHs and phthalates as potential contaminants of concern. The PDI suggests these compounds are not primary constituents of concern. The PAHs and phthalates have been removed from the Table 5-6 since their concentrations are below clean-up criteria in the groundwater and were found in such minute amounts in the soil during both the RI and PDI. These primary COIs are presented in Table 5-7 along with some of their chemical properties. These chemical properties are pertinent to the design of the remedy. These contaminants are presented in Table 5-6 with their maximum concentrations detected by media during the RI and PDI.

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TABLE 5-6

Physical Properties of Primary COIs

Representative	Class	æ	Boiling Point	Henry's	Vapor	Water	Race of
Contaminants	and the special section of the special sectin	:	(0,	(atm)	(mmHG)	(mg/l)	Stripping
Benzene	Aromatics	78.11	78.00	315.00	95.04	1750	Easily Stripped
Ethylbenzene	Aromatics	106.16	136.20	357.00	06.6	152	Easily Stripped
Toluene	Aromatics	92.13	110.70	360.00	28.40	535	Easily Stripped
Xylenes	Aromatics	106.16	~ 140	338.00	8.50	198	Easily Stripped
Tetrachloroethylene	Halogenated Hydrocarbons	166.00	121.00	1,457.00	133.35	150	Easily Stripped
1.1,1-Trichloroethane	Halogenated Hydrocarbons	133.42	75.00	251.00	17.80	1500	Easily Stripped

Notes: The Boiling Temperature of Xylenes is not available. However, Xylene isoniers boiling points are:

139 °C for m-Xylene (meta)
144 °C for o-Xylene (ortho)
138 °C for p-Xylene (para)

TABLE 5-7

Primary COIs Maximum Concentration Detected by Media

Contaminants	Soil (ppm)	Groundwater (ppb)	Sediments (ppb)	Air (ppb)
Benzene	28	270	-	1.00
Ethylbenzene	1,300	6,200	150	0.36
Toluene	2,600	350,000	13	0.84
Xylenes	8,400	45,000	1,400	-
Tetrachloroethylene	0.004	430	3	-
1,1,1-trichloroethane	7.6	11		

⁽¹⁾ This table was presented in the ROD and modified after the PDI.

6.0 TREATABILITY STUDIES

6.1 IRON PRECIPITATION STUDY

The iron precipitation study was conducted using two samples collected with the 4" split spoon and plastic liners. The samples collected were two columns of soil inside clear 2½"-by-2' sleeves. The two samples were connected using a rubber coupler and silicone. After connecting the two tubes, soil samples were collected from each end of the column to determine initial iron concentrations. The original design of the column was to pump water from well DW-2 to the top of the soil column and allow it to pass by gravity through the soil column while diffusing air from the bottom of the column into the sample. Refer to Figure 6-1.

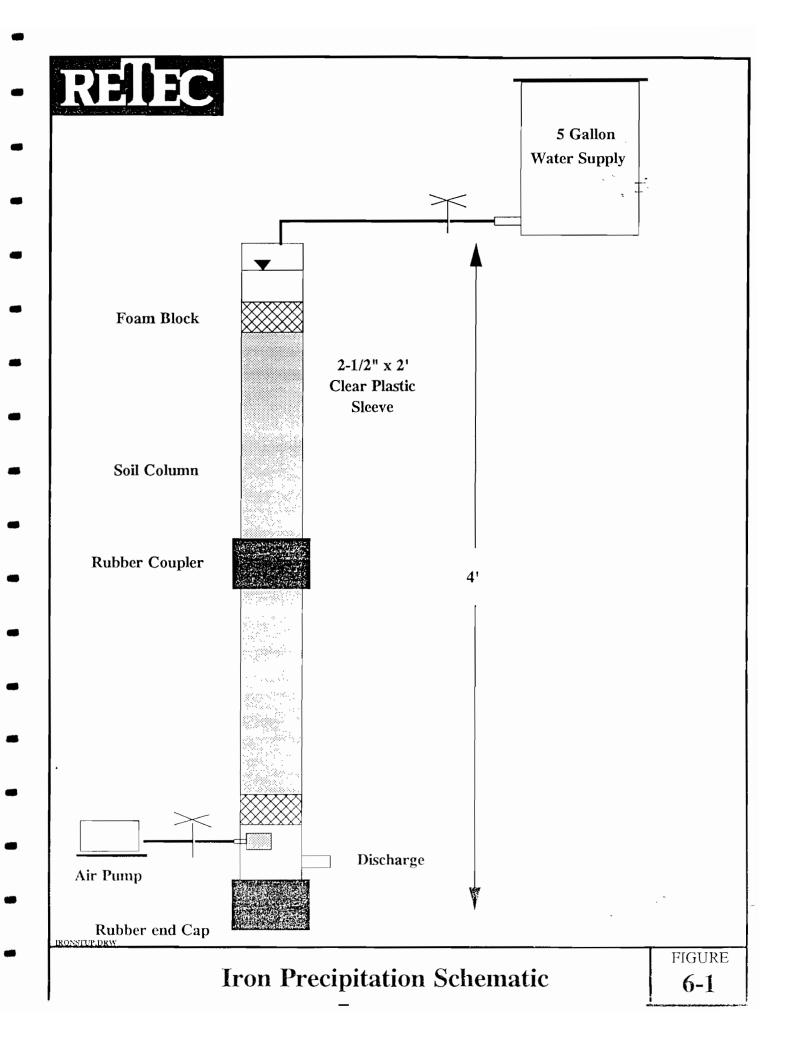
After initial setup of the column, flow was negligible due to sample disturbance resulting in compaction of the loose sands which will not occur in the field during implementation of the remedy. To try and increase the flow rate, the sample was remolded, and the apparatus reconnected. The flow in the column increased to approximately 0.2 gpd. The air flow was maintained by keeping a constant pressure of air below the soil column. Initially, the soil column needed daily water additions above the soil. As the study progressed, significant settlement of the soil column was noted, resulting in lower flow rates through the column. The system operated from March 8 until June 3, 1993. The material in the column has changed appearance modestly since it was initially remolded; the soil now has a slightly darker color in the lower half of the soil column, and a few very small spots of iron staining have developed. Samples were collected from the top, bottom, and middle of the column and sent for iron analysis.

It appears that some iron precipitation occurred during the three months of the study; however, it does not appear that the precipitation has significantly impacted air or water flow. Reduction in permeability has been caused by settlement rather than iron precipitation. Once the sample results are received from the laboratory, a final determination will be made.

6.2 BIODEGRADATION STUDIES

6.2.1 Microbial Characterization

Tests to characterize the microbial populations in the soils and groundwater at the site were performed on samples collected from soil borings B-93-1, B-93-2, and B-93-3, and from monitoring wells SW-2, WT-93-2, and SW-6 (wells in the vicinity of the soil sampling



locations). The procedures used for microbial enumeration of soil and groundwater samples are described in RETEC Standard Operating Procedure (SOP) #510.

The locations of the soil borings were chosen to reflect the range of soil-phase contaminants and constituent concentrations that exist at the site. B-93-1 was selected to characterize the areas of highest observed contaminant concentrations along the western bulkhead. B-93-2 was located to characterize the soils within the tank farm, which is suspected as a potential source area. B-93-3 was chosen to characterize the soils at the upgradient portion of the site, which the data from the RI suggests may have been impacted from an off-site source of CVOCs.

The depths of the soil samples were chosen to facilitate evaluation of physical, chemical, and microbiological characteristics of three specific zones within the B level:

- near the top (unsaturated) portion of the B level (approximately four feet above the static groundwater elevation);
- at the water table interface;
- approximately three feet below the water table interface (near the bottom of the B level).

The shallow soil sample represents conditions in the vadose zone at a depth that would be addressed by venting/bioventing. The sample collected at the water table interface is most likely to reflect conditions resulting from exposure to LNAPL, and to exhibit the highest contaminant concentrations, and is potentially subject to treatment by *in situ* bio-remediation through the addition of nutrients and appropriate electron acceptor. The sample collected from within the saturated zone was expected to contain lower concentrations of the more soluble contaminants at the site. Remediation of contamination within this level would occur primarily through aqueous flushing, stripping through groundwater sparging, and *in situ* bio-degradation.

6.2.1.1 Initial Microbial Enumerations

The results of the microbial enumeration of the soil samples are presented in Table 6-1. These results confirm the presence of relatively high numbers of total heterotrophic bacteria, ranging from 7.2 x 10^5 colony forming units (CFU) per gram, to over 1 x 10^8 CFU/gm. Numbers of bacteria capable of growth on volatile organic compounds (VOC) were somewhat lower, ranging from 7.6 x 10^3 CFU/gm to 2.6 x 10^6 CFU/gm; however, these are very

TABLE 6-1

Results of Initial Microbial Counts: Soil Samples

Soil Samples	Initial Counts (CFU/gm)
B-93-1 (2-4') Heterotrophs VOC Degraders	1.1 x 10 ⁷ 4.5 x 10 ⁵
B-93-1 (4-6') Heterotrophs VOC Degraders	1.02 x 10 ⁸ 4.9 x 10 ⁵
B-93-1 (8-10') Heterotrophs VOC Degraders	6.4 x 10 ⁶ 1.1 x 10 ⁴
B-93-2 (12-14') Heterotrophs VOC Degraders	1.4 x 10 ⁷ 7.1 x 10 ⁵
B-93-2 (15-17') Heterotrophs VOC Degraders	1.5 x 10 ⁷ 2.56 x 10 ⁶
B-93-2 (19-21') Heterotrophs VOC Degraders	9.1×10^6 3.8×10^5
B-93-3 (3-5') Heterotrophs VOC Degraders	8.5 x 10 ⁶ 1.7 x 10 ⁴
B-93-3 (7-9') Heterotrophs VOC Degraders	7.2×10^5 7.6×10^3
B-93-3 (11-13') Heterotrophs VOC Degraders	7.8×10^6 3.0×10^3

Notes:

CFU: Colony-forming units VOC: Volatile organic compound

Microbial counts performed according to RETEC SOP #510.

typical of sites containing volatile organic contaminants and represent a healthy microbial population. It should be noted that the two soil samples exhibiting the highest contaminant concentrations (B-93-1, 4-6'; and B-93-2, 15-17') also exhibited the highest densities of total heterotrophic and VOC-degrading bacteria. These samples were collected from the vicinity of the groundwater table interface and, therefore, high concentrations of contaminants having a density less than that of water were expected. Exposure of the micro-organisms to these elevated contaminant concentrations has not produced any bio-toxicity effects.

The results of the microbial enumeration of the three groundwater samples are presented in Table 6-2. These results are generally consistent with the data obtained from the microbial enumeration of the soil samples, and show essentially equal numbers of total heterotrophic bacteria in each of the groundwater samples (approximately 3 x 10⁵ CFU/Ml). Numbers of VOC-degrading bacteria were somewhat more variable, ranging from 5 x 10² CFU/mL to 7 x 10⁴ CFU/mL. Like the results of the soil analyses, these data are suggestive of normal levels of microbial activity that may be stimulated to increase rates of organic constituent biodegradation.

6.2.1.2 <u>Nutrient Stimulation Testing</u>

In order to evaluate the response of the soil microbial populations to enhanced oxygen and nutrient conditions, slurries of water table and saturated zone soil samples were prepared in shake-flasks using deionized water to achieve a solids content of approximately 20%. Duplicate slurries from each sample were prepared: one slurry received inorganic nutrients (100 mg/L nitrogen, 10 mg/L phosphorus), while the other slurry received no additional nutrients and served as an aeration control. All of the flasks were then placed on a shaker table for approximately 48 hours, after which samples of each slurry were collected for microbial enumeration of total heterotrophic and VOC-degrading bacteria. Details of this nutrient stimulation procedure are provided in RETEC SOP #545.

The results of the nutrient stimulation study are shown in Table 6-3. Significant increases (greater than 10x) in total heterotrophs and VOC-degraders resulting from aeration alone were observed only in the saturated zone (8-10' deep) sample from B-93-1. A significant increase in VOC-degraders only, as a result of aeration, was also noted in the water table (4-6' deep) sample from B-93-1. Insignificant additional increases in microbial numbers in these samples were recorded when nutrients were also provided. Essentially no changes in microbial numbers were observed in the samples collected from the northeast portion of the site (boring B-93-3); however, organic analyses of these samples showed very low contaminant

TABLE 6-2

Results of Initial Microbial Counts: Groundwater Samples

Groundwater Samples	Initial Counts (CFU/mL)
WT-93-2 Heterotrophs	2.7 x 10 ⁵
VOC Degraders	5.0×10^{2}
SW-2 Heterotrophs VOC Degraders	3.8×10^5 7.0×10^4
SW-6 Heterotrophs VOC Degraders	3.2×10^5 1.3×10^4

Notes:

VOC: Volatile organic compound

CFU: Colony-forming units

Microbial counts performed according to RETEC SOP #510

concentrations, suggesting that microbial growth in these samples may be limited by carbon availability, rather than oxygen or nutrient availability.

6.2.2 Chemical Characterization

6.2.2.1 Soil and Groundwater Chemical Analyses

Each of the soil samples were analyzed by GC/MS for volatile organic compounds (EPA Method 8240) with a library search for the 15 highest non-target peaks and for semi-volatile organics (EPA 8270) with a library search for the 15 highest non-target peaks. The soil samples were also analyzed for total nitrogen, total phosphorus, and total iron. Groundwater samples were analyzed for the same parameters as the soil samples, with the exception of TPH and hydrocarbon characterization. The objective of these chemical analyses is to provide an indication of the chemical environment to which indigenous microbial populations have been exposed.

The results of the chemical characterization of the soil and groundwater samples are presented in Table 6-4. These data show that soil contamination is primarily composed of volatile organic compounds, which exist at highest concentrations near the water table interface near the western bulkhead and in the tank farm area. Significant levels of soil contaminants were not found in samples collected from B-93-3, located near the northeastern property boundary. Contaminant concentrations in groundwater samples from these areas of the site were relatively low; only the sample from the tank farm area (WT-93-2) exceeded 1.0 mg/L total organics. As with the soil samples, organic constituents were predominantly volatile organics.

Concentrations of total iron in soils were shown to be relatively low, and were fairly consistent among the samples, ranging from 1.34 mg/kg to 5.77 mg/kg. Iron concentrations in the groundwater samples were higher and were considerably more variable, ranging from 8.62 mg/L (SW-2) to 101 mg/L (WT-93-2). These data do not indicate the form of the iron, and it is not possible to determine from these data the propensity for iron to form precipitates upon exposure to oxygen. Additional laboratory tests, discussed in later sections of this report have been performed to empirically evaluate the potential for the precipitation of iron and other groundwater minerals.

Table 6-4 also shows concentrations of nitrogen and phosphorus, which constitute the principal nutrients required by heterotrophic micro-organisms. Low concentrations of nitrogen and phosphorus, relative to the total mass of bio-degradable organic carbon, may limit microbial growth and, correspondingly, limit rates of contaminant bio-degradation. The optimal ratio of

TABLE 6-3
Microbial Testing

Sample I.D.	Initial Counts (CFU/mL)	O ₂ Only (CFU/mL)	O ₂ & Nutrients (CFU/mL)
B-93-1 (4-6')			
Total Bacteria	2.04×10^7	4.8×10^7	6.5×10^7
VOC-Degraders	1.0 x 10 ⁵	1 x 10 ⁶	3.8×10^{5}
B-93-1 (8-10')			
Total Bacteria	1.28×10^6	1.1 x 10 ⁷	1.4×10^7
VOC-Degraders	2×10^{3}	4 x 10 ⁴	3 x 10 ⁴
B-93-2 (15-17')			_
Total Bacteria	3.0×10^6	5.7 x 10 ⁶	9.4×10^6
VOC-Degraders	5.1 x 10 ⁵	1.4 x 10 ⁶	2.9×10^{5}
B-93-2 (19-21')			
Total Bacteria	1.82 x 10 ⁶	2.2 x 10 ⁶	2.5×10^6
VOC-Degraders	8 x 10 ⁴	3 x 10 ⁴	7 x 10 ⁴
B-93-3 (7-9')			
Total Bacteria	7.2 x 10 ⁵	2.9 x 10 ⁵	3.2×10^5
VOC-Degraders	2×10^{3}	1 x 10 ³	1×10^{3}
B-93-3 (11-13')			
Total Bacteria	1.56 x 10 ⁶	1.9 x 10 ⁶	2.5×10^6
VOC-Degraders	1 x 10 ³	2 x 10 ³	3×10^3

TABLE 6-4

Summary of Soil and Groundwater Chemical Data For Bioremediation Evaluation

										_		
C:N:P	100:40:33	100:0:03:5	100:1:25	100:10,000:320	100:66:71	100:0:04:4	100:98:251	100:8:0.9	100:3,080:13,300	100:4,300:12,250	100:3,580:4,000	100:24,300:1,000
TOTAL	2.78	5.77	3.81	8.620	3.23	2.21	1.34	101.000	4.4	4.05	5.27	59.000
TOTAL TOTAL NITROGEN PHOSPHATE	59.3	160	19.5	0.160	46.4	101	47.2	0.120	160	147	124	0.290
TOTAL NITROGEN	72.5	<2	<2	3.730	42.5	<2	18.5	1.100	37	22	111	7.27
тен	<10.0	NA	< 10.0	AN	< 10.0	550.0	<10.0	NA	<10.0	< 10.0	< 10.0	NA
TOTAL TOTAL 8270 ORGANICS	179.0	2,924.8	78.0	0.050	64.9	2,595.1	18.8	13.095	1.2	1.2	3.1	0.030
H	29.1	42.8	4.9	0.042	64.1	51.1	17.3	0.589	1.2	1.2	3.1	0.013
8270 TIC	27.8	33.0	4.0	0.042	63.7	47.0	17.2	0.485	0.4	0.7	8.0	0.004
8270 TCL	1.3	8.6	0.9	0.005	0.4	4.1	0.1	0.104	8.0	0.5	2.3	0.009
TOTAL 8240	149.9	2,882.0	73.1	0.008	8.0	2,544.0	1.5	12.506	<0.1	< 0.1	<0.1	0.017
8240 TTC	2.0	286.0	N	ND	8.0	1,735.0	0.7	0.886	ND	ND	ND	ΩN
8240 TCL	147.9	2,596.0	73.1	0.008	0	809.0	8.0	11.620	< 0.1	< 0.1	< 0.1	0.017
Matrix	Soil	Soil	Soil	GW	Soil	Soil	Soil	GW	Soil	Soil	Soil	GW
Location	Bulkhead	Bulkhead	Bulkhead	Bulkhead	Tank Farm	Tank Farm	Tank Farm	Tank Farm	Northeast	Northeast	Northeast	Northeast
Number	B-93-1, 2-4'	B-93-1, 4-6	B-93-1, 8-10'	SW-2	B-93-2, 12-14' Tank Farm	B-93-2, 15-17, Tank Farm	B-93-2, 19-21' Tank Farm	WT-93-2	B-93-3.3-5	B-93-3,7-9	B-93-3, 11-13	SW – 6

· ` ; =:

NOTES: Soil data reported in units of mg/Kg.

Groundwater data reported in units of mg/L

NA = Not Analyzed

8240 := EPA Method 8240 - Volatile Organics

8270 = EPA Method 8270 - Semivolatile Organics

TCC = Target Compound List Constituents

TIC = Tentatively Identified Compounds (Non-Target Compounds)

(carbon:nitrogen:phosphorus for bio-degradation processes is typically about 100:2:0.5. RETEC, 1991). As the table indicates, nitrogen and phosphorus were present at concentrations exceeding the theoretical requirement in all but three of the samples that were evaluated. While all of the nitrogen and phosphorus in these samples may not be biologically available, these data nevertheless suggest that significant concentrations of inorganic nutrients, relative to organic compounds, exist at the site, and that limitation of bio-degradation rates due to ambient nutrient levels may not be a significant problem at the site.

6.2.2.2 Nutrient Adsorption Testing

A simple testing program was used to evaluate the degree to which inorganic nutrients of the type typically added during in situ bioremediation programs may be adsorbed by the soil matrix of the site. This testing consisted of preparing soil slurries using sample material collected from the deep (saturated zone) soil levels. The slurries were spiked with a known concentration of a nitrogen/phosphorus nutrient blend (ammonium chloride, monopotassium and dipotassium phosphates), used for bioremediation applications, which was mixed for several hours after which the nutrient concentrations were re-measured (refer to RETEC SOP #730).

The results of the nutrient adsorption testing are shown in Table 6-5. These data indicate that nutrient adsorption onto the saturated zone soils was generally insignificant. Phosphorus was observed to decrease by 9 percent upon exposure to the saturated zone soil sample collected from boring B-93-1. This degree of nutrient loss does not suggest however, that the transport of nutrients through the site soils will be compromised, in the event that nutrient addition is found to be required to enhance contaminant bio-degradation rates.

6.2.2.3 Nutrient Precipitation Testing

The potential for precipitation due to reaction of introduced nutrients with dissolved minerals (primarily calcium and magnesium) in the groundwater was measured by first filtering two groundwater samples to remove suspended solids, and then spiking the samples with known concentrations of nitrogen (1,000 mg/L nitrogen as NH₄Cl) and phosphorus (500 mg/L phosphorus as KH₂PO₄). The samples were then placed on a rotary shaker, and aliquots collected after 24 and 48 hours for measurement os total suspended solids (TSS), amonia, nitrogen and phosphorous. The results of the nutrient precipitation tests are shown in Table 6-6. Slight increases in TSS were noted at the two sampling times, suggesting the formation of small amounts of precipitate. At the conclusion of the study, analysis of ammonia-nitrogen and phosphorus in filtered samples, however, demonstrated that more than 80% of the original

TABLE 6-5

Results of Nutrient Adsorption Testing

		Meası	Percent		
Sample Description	Сотроило	Sample A	Sample B	Mean	Adsorbed
Flasks 1 & 2	Ammonia—N	571	335 *	571	2 9
B-93-1 (8-10')	Phosphorus	326	317	322	
Flasks 3 & 4	Ammonia – N	594	615	605	0
B-93-2 (19-21')	Phosphorus	349	349	349	
Flasks 5 & 6	Ammonia-N	582	575	579	1 4
B-93-3 (11-13')	Phosphorus	341	331	336	
Flasks 7 & 8	Ammonia – N	578	592	585	
Water Controls	Phosphorus	349	354	352	

Notes:

Nutrient adsorption testing performed according to RETEC SOP #730. Samples incubated 24 hours on a rotary shaker, then centrifuged and filtered. Filtrate preserved with H_2SO_4 and then analyzed for total Kjeldahl Nitrogen (EPA 353.2) and total phosphorus (EPA 365.2). Samples could not be analyzed for ammonia and ortho-phosphate directly because sample preservative results in precipitation of phosphate.

^{*} This sample not included in calculation of sample mean due to disparity between duplicate samples.

TABLE 6-6

Results of Precipitation Testing

Sample Description	Compound	Measured Concentration . (mg/L)			
		TSS (mg/L)	Ammonia	Phosphate	
Initial Groundwater Composite	Sample A Sample B	5,040 5,000			
(Before Filtering)	Mean	5,020			
Initial Groundwater	Sample A	4			
Composite (After Filtering)	Sample B Mean	8			
8 Hour Samples	Sample A	32			
	Sample B Mean	8 20			
24 Hour Samples	Sample A	40	820	531	
	Sample B Mean	28 34	816 818	502 516. 5	
			(82 % Recovery)	(103 % Recovery)	

Notes:

Composite GW sample filtered, spiked with 1,000 mg/L nitrogen (as NH_4Cl) and 500 mg/L phosphorus (as potassium phosphate). Samples incubated for 24 hours on a rotary shaker, then centrifuged and filtered. Filtrate preserved with H_2SO_4 , analyzed for total Kjeldahl Nitrogen (EPA 353.2) and total phosphorus (EPA 365.2). Samples could not be analyzed for ammonia and ortho-phosphate directly because sample preservative results in precipitation of phosphate.

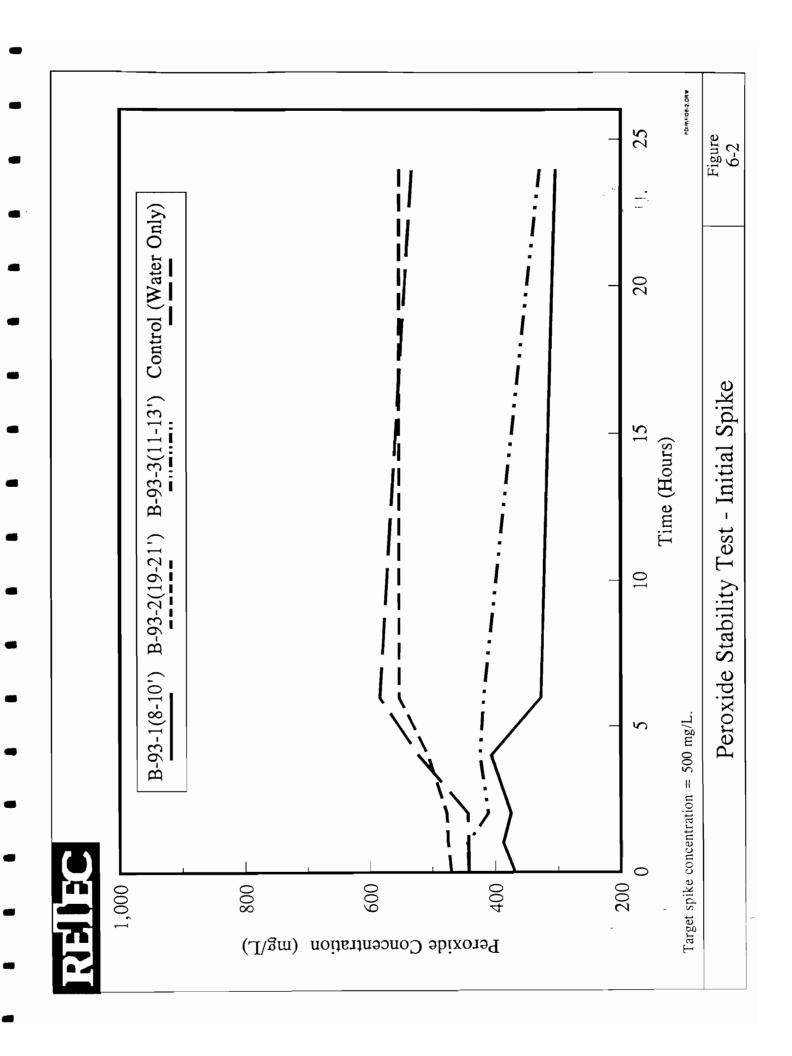
ammonia concentration, and 100% of the original phosphorus concentration, were recovered after 24 hours of shaking. These results indicate that significant losses of added nutrients due to reaction with groundwater constituents, and subsequent precipitation, are not expected and should not interfere with a bioremediation process at the site.

6.2.2.4 Peroxide Stability Testing

The stability of hydrogen peroxide, which is often considered as an oxygen source for in situ bio-remediation, was evaluated by spiking the slurries used for the nutrient adsorption study with dilution-grade, 50% hydrogen peroxide (Interox-America) to achieve an initial target peroxide concentration of 500 mg/L. Each of the slurry flasks was placed on a rotary shaker, and samples of the aqueous phase collected at regular intervals for measurement of residual peroxide concentrations. After approximately 24 hours, the slurries were spiked with a second peroxide addition, and the test repeated.

The results of the hydrogen peroxide stability test are presented in Table 6-7, and illustrated graphically in Figures 6-2 and 6-3. Figure 6-2 shows the observed concentration of hydrogen peroxide in the test flask as a function of time. The data suggest that peroxide stability increases with depth, which is fairly common, since the shallow soils tend to contain more organic compounds and therefore more chelated metals than deeper soils. The stability in the presence of soils from the 19-21 foot horizon were very good, indistinguishable from water alone. Figure 6-3 illustrates the second 24 hours of the study. The initial concentrations in this graph are approximately 500 mg/L greater than the concentrations present at the end of the first segment of the test. During this second portion of the test, all of the samples reached a final peroxide concentration of about 500 mg/L, suggesting that some of the losses in the first phase were from consumption rather than catalytic decomposition. This trend suggests that peroxide stability and therefore utilization at this site would improve over time.

These results are consistent with the data obtained from the chemical analyses of the soils, which showed very low concentrations of iron. High iron concentrations in soils are known to catalyze the rapid decomposition of peroxide, which often leads to de-gassing, i.e., the formation of bubbles of oxygen gas. When de-gassing occurs within the subsurface, it can result in poor utilization efficiency of oxygen by the sub-surface microorganisms and, in severe instances, can reduce the permeability of the formation due to plugging of the soil pore spaces with oxygen gas. The results obtained from this test suggest that hydrogen peroxide constitutes a potentially effective oxygen source for *in situ* bio-remediation at the Shore Realty site.



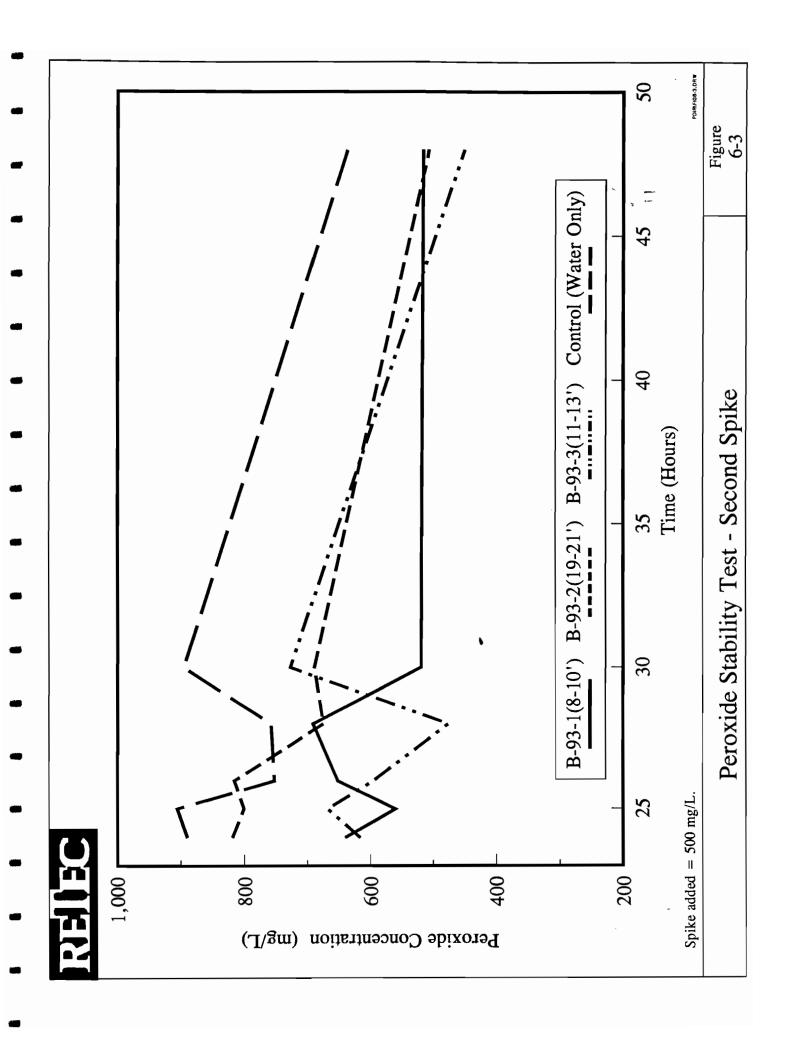


TABLE 6-7

Results of Peroxide Stability Testing

		\$\$\$\$\$\$\$\$\$\$\$\$\$\$\;\`\\.\`	Sample Peroxide				
Sample Description		Time (hr)	Concentration (mg/L) Sample A Sample B Mean				
	Spike						
Flasks 1 & 2	Initial	0	384	354	369		
B-93-1 (8-10')		1	417	358	387		
		2	359	391	375		
		4	322	490	406		
		6	319	336	327		
		24	330	279	304		
	Repeat	0	624	656	640		
		1	644	476	560		
		2	645	660	652		
		4	761	620	691		
		6	483	556	519		
		24	548	483	516		
			-				
Flasks 3 & 4	Initial	0			0		
B-93-2 (19-21')		1	489	460	475		
,		2	467	486	477		
		4	525	490	507		
		6	532	574	553		
		24	605	502	554		
	Repeat	0	814	823	819		
		1	711	891	801		
		2	823	809	816		
		4	568	782	675		
		6	677	700	689		
		24	538	475	507		
Flasks 5 & 6	Initial	0					
B-93-3 (11-13')		1	442	446	444		
		2	417	405	411		
		4	446	403	424		
		6	400	435	418		
		24	335	323	329		
	Repeat	0	613	620	616		
		1	705	632	668		
		2	559	631	595		
		4	415	537	476		
		6	707	748	727		
		24	461	439	450		
			422	460	442		
Flasks 7 & 8	Initial	0	422	462	442 354		
Water Controls		1	353	354	443		
		2	417	469			
		4	446	597	522 584		
		6	488	679			
		24	503	563	533		
	Repeat	0	851	930	890		
		1	907	906	906		
		2	721	785	753		
	1				7.50		
		4	855	660	758		
		4 6 24	855 855 610	935 662	895 636		

Notes:

Peroxide testing performed according to RETEC SOP #735 using slurries with a 10 percent solids loading. Slurries were used for nutrient adsorption testing (RETEC SOP #730) prior to initial spiking with hydrogen peroxide. Target spike concentration was 500 mg/L hydrogen peroxide.

6.2.3 Evaluation of Contaminant Bio-degradation - Slurry Respirometry

The bio-degradation of contaminants in the site soils, and the effect of nutrient additions in enhancing the rate and extent of constituent bio-degradation, was evaluated using a slurry respirometry system. A composite soil sample was prepared from the samples collected from the water table interface and from the saturated zone sampling depths. Using the composited sample and groundwater from the site, ten identical slurry mixtures were prepared; five slurries were supplemented with inorganic nutrients (100 ppm nitrogen as KNO₃ and 20 ppm phosphorus as an equimolar mixture of KH₂PO₄ and K₂HPO₄) and the remaining five slurries received no The slurries were be prepared in electrolytic respirometer flasks additional nutrients. (Bioscience Management, Inc., Bethlehem, PA). After preparation, the slurries were allowed to equilibrate and two slurries were sampled for analysis of initial concentrations of aromatic hydrocarbons (Method 8020) and TPH (Method 8015). Two flasks (one receiving nutrients and one without nutrients) were sterilized by addition of mercuric chloride to eliminate biological activity. All of the flasks were then sealed and connected to the electrolytic respirometer instrumentation to continuously monitor oxygen uptake (refer to RETEC SOP #526). As oxygen is used in the biologically active flasks, it is simultaneously replaced by an electrolytic reaction, thus maintaining aerobic conditions within the sealed reactor. Carbon dioxide is removed from each flask using an alkaline trapping agent, which maintains a neutral pH. The sterilized flasks serve as controls to monitor non-biological consumption of oxygen and volatile losses of the contaminants.

The results of the oxygen uptake monitoring during the respirometry study are summarized in Table 6-8, and are presented graphically in Figure 6-4. This figure shows the cumulative consumption of oxygen over a 300-hour respirometry study. The data shows that the slurry reactor containing nutrients consumed approximately 40% more oxygen than the slurry without nutrients. Additionally, the flasks which did not receive nutrients appeared to reach a plateau, at which point significant additional oxygen uptake was not observed, sooner than the slurries that were supplemented with nutrients. The figure shows that sterile conditions were maintained in the control reactors, as evidenced by the insignificant amount of oxygen uptake.

The results of the analyses of contaminant concentrations at the initiation and completion of the slurry respirometry studies are presented in Tables 6-9 and 6-10, respectively. These data show low levels of total xylenes present in the soil fraction of both the nutrient-amended and nonutrient slurries at the initiation of the study, while significant concentrations of toluene and

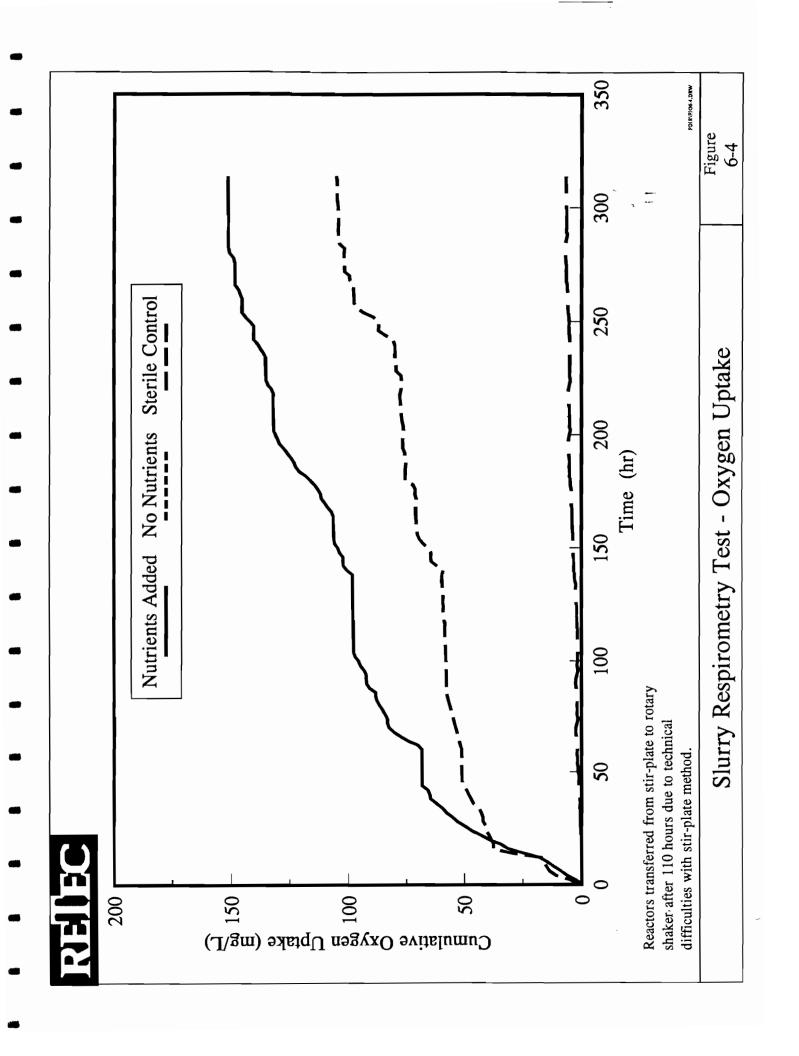


TABLE 6-8

		Dioactive Academia	e lo	Average		Bioactive Reactors		Average	Sterile Control Reactors		Average
l me (br)	Nitrogen a	Nitrogen and Phosphorus Added Cell 1 Cell 2 Cell 3	rus Added Cell 3	of Cells	Cell S	No Nutrients Added	Cell 7	of Cells	Cell 4 Ce		of Cells
0	0.0	0.0	0.0	0.0	0.0	00	00			San II mile	4 and 6
7	3.2	1.5	2.1	2.3	3.6	4.1	0.0	0.0		0.0	0.0
4	8.9	3.0	6.0	0.9	8.0	10.8	12.3	10.6	0.0	0.0	0.0
9	13,3	3.0	10.0	8	12.0	14.2	14.7	13.6		0.0	0.0
00	18.2	3.0	13.8	11.7	12.1	2.7.1	17.7	17.0		0.0	0.0
10	22.3	3.5	173	17.7	12.1	15.3	17.1	V.4.1		0.0	0.0
2 5	27.0		C. 7.	1.1	1.71	10./	18.8	15.9		0:0	0.0
71	0.77	4.0	21.3	17.8	12.1	18.2	20.7	17.0	0:0	0.0	0.0
14	35.1	13.0	26.9	25.0	Reactor	30.0	30.4	30.2	0.0	0.0	0.0
16	42.2	19.1	32.8	31.4	Destroyed	37.9	37.3	37.6		0.0	00
18	45.4	20.7	37.8	34.6	. !	37.9	37.4	37.6		0.0	200
20	48.7	27.8	42.1	39.5		38.7	38.5	38.6		2.0	3 6
77	51.8	32.5	45.4	43.2	!	0.04	30.0	30.5		7.0	7 6
24	55.1	36.5	49.2	46.9	!	42.4	30.1	300		+ (5)	† (
56	60.5	36.6	523	40.8		\$ C P	20.1	0.00			ر د د
28	65.4	37.4	56.4	52.1		42.4	39.1	5.04		0.3	0.3
£ 6	7.60	37.4	50.1	17.7	l 1	4.74	4I./	47.1		0.3	0.4
3 8	1.60	4.10	19.1	55.3	1	47.5	41.7	42.1		0.4	0.5
75	8.4/	3/.5	61.1	57.8	!	42.5	43.2	42.9	9.0	0.5	9.0
34	80.0	37.5	61.1	59.5	!	44.6	43.3	43.9		9.0	0.7
36	85.9	38.3	62.0	62.1		46.3	45.9	46.1		0.5	0 6
38	89.0	40.8	64.1	64.6	!	47.3	46.9	47.1		9.0	0.0
9	89.0	41.6	64.1	64.9	1	48.3	47.9	48.1		80	1
42	88.9	44.5	04.0	65.8	1	49.3	48.9	49.1		00	- 1
4	6.06	47.3	9.99	68.3		50.3	49.9	50.1	1.2	0.0	1
46	6.06	47.3	9.99	68.3	!	51.3	50.9	51.1		10	1 1
48	6.06	47.3	9.99	68.3	!!	51.3	50.9	51.1		80	10
20	6.06	47.3	9.99	68.3	!	51.3	50.9	51.1		1.2	1 2
25	6.06	47.3	9.99	68.3	!!	51.3	50.9	51.1		1.2	
2 4	90.9	47.3	9.99	68.3	!	51.3	50.9	51.1		1.5	8
20	6.06	47.3	66.7	68.3	-	51.3	50.9	51.1		4.1	1.7
28	91.0	47.4	66.7	68.4	!	51.4	51.0	51.2		1.7	20
09	91.0	47.4	66.7	68.4	!	51.4	51.0	51.2		1.4	1.7
62	92.2	49.9	0.89	70.0		51.9	51.5	51.7		1.7	2.1
64	97.0	53.7	72.1	74.3	!	52.4	52.0	52.2			2.1
99	100.9	56.7	75.0	77.5		52.9	52.5	52.7	,	1.9	2.2
89	104.5	59.7	77.4	80.5	!	53.4	53.0	53.2	1	2.1	2 6
20	106.8	61.9	78.9	82.5	!	53.9	53,5	53.7	-	1 0	
72	107.4	62.7	79.4	83.2	l I	54.4	54.0	54.2	-	1.5	÷ -
74	107.4	62.7	79.4	83.2	1	54.9	54.5	54.7		1.5	1.0
92	108.7	63.5	80.3	84.2		55.4	55.0	55.2		1.5	10
78	109.9	64.3	81.4	85.2		55.9	55.5	55.7		4.	1 7
80	111.2	62.9	82.4	86.5	-	56.4	56.0	56.2		1.7	2.0
82	112.7	66.5	83.2	87.5		56.9	56.5	2,95		1.5	- 1
•											

TABLE 6-8

			Cumulative Oxygen Consumption (mg/L)			
	Bioactive Reactors	Average	Bioactive Reactors	: 6	Sterile Control Reactors	Average
Time	en and Phosphorus	of Cells	No Nutrients Added	of Cells		of Cells
(hr)	Cell 2	1, 2 and 3	9 [5, 6 and 7	No Nut's. With Nut's	4 and 8
98	113.2 67.2 84.1	88.2	57.9 57.5	57.7	2.5 1.8	2.1
88	Transferred to single shake-flask	6.06	Transferred to single shake—flask	57.7	Transferred to single	2.2
8	reactor due to risk of reactor	92.2	reactor due to risk of reactor	57.7	shake-flask reactor	2.5
92	failure.	92.3	failure.	57.6	due to risk of reactor	2.3
94		92.3	1 1	57.7		1.8
96	1 1	93.3	11	57.8	!!	1.5
86		94.8		57.8		1.9
100		95.6	1 1	57.9	-	1.7
102	1 1	97.1	!!	58.1		1.7
104		7.76	1 1	57.9	-	1.6
106		97.8		58.1	!!	1.6
108		97.8		58.3	:	. 8
110	!!	97.8		58.4		1.7
112	1 1 1 1 1	97.8	1 1	58.4	!	1.7
114	1 1 1 1 1	97.8		58.4	1	· ~
116	1 1 1 1 1	97.9	1	58.3		0.0
118		97.9	;	888	1	1.0
120	!	98.1	;	888	!!	2.7
122	!	080		50.5		; c
124	!!	0.8.0		2.60		0.7
127		70.1		0.80	11	7.7
120	111	98.1		59.3	!!!	2.4
128	: ! ! ! ! ! !	98.2		59.1		2.5
130	1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1	98.2	1 1 1 1	59.3		2.5
132	1 1	98.2		59.5		2.5
134	1 !	98.3	1 1 1 1	59.5		2.4
136		98.3		59.8		2.9
138	!!!	98.1	1	59.6		2.9
140	!!!	100.5		59.2		3.2
142	1	102.1	!!	61.7	!!	3.2
144		102.3		64.3		3.2
146	!	102.3	1	64.5		3.5
148		103.8	1 1	64.5	!!	3.5
150		104.5		65.6	11 .	3.7
152	!!	105.9	1	62.9		3.6
154	1 1	106.2	!!	69.5	, ,	3.6
156	!!!	106.3	!!!	70.3		3.6
158		106.4	! !	70.6		3.8
160	1 1	106.4	;	70.9	1 1	3.8
162		106.5	:	70.8		4.0
164		106.5		71.0		3.9
166		107.2	:	71.1		4.0
168	1 1	108.9	!!	71.1		4.2
170		110.1	1 1	70.9		4.4

TABLE 6-8

Time Nitrogen and Picophicus Added O'Cali Substitute Added O'Cali Substitu					Cumulativ	Cumulative Oxygen Consumption (mg/L)	on (mg/L)				
Nitrogen and Mode OriCials		Bioa	ctive Reactors	Average		Bioactive Reactors		Average	Sterile Control R		Average
1116 1116 1117 1118 118	Time (hr)	Nitrogen an	nd Phosphorus Added Cell 2 Cell 3	of Cells		No Nutrients Added	C 11-2	of Cells	Cell 4 C		of Cells
111.9	172	1		111.6	-			708			1 all 0
113.0 113.0 114.4 115.1	174		!	111.9	;	!] 	71.4		- - 	4 4
11144	176	1	1	113.0		!	1	71.2	1	-	4
1181	178	!!	;	114.4	1	!	!	72.1	!	!	8
1210	180	1	:	116.2	!	!	 	75.5	!	-	4.7
121.0	182	1	!!	118.1		*	-	75.6	!	-	5.1
122.5	184	1		121.0	!	ţ	1	75.6	!		2.0
124.5 124.5	186	!	1	122.5	1	1	;	75.4	!	1	5.5
124.5	188	1	1 1	123.2		!!	!	75.2	;	-	1 C
1260	190	1	!	124.5	:	!	!	75.0	;) v
1975 1975	192	1	11	126.0	i	1	;	7.57	!		. v
1994	194	1		127.5	ł	;	!	76.5	!		t v
130.2	196	1	!!!	129.4	1	!	;	7.97	!		
13.11	198	!	1	130.2	-	!	!	76.6	1		
1320	200	!	11	131.1	!	;	!	76.4	!		
1320	202	1		132.0	;	!	ļ	76.3		 	7.7
132.0	204	!		132.0				50.0		 	4 , ,
132.1	202	1		132.0	!	i I	i I	4.0,		t I	5.1
13.21	200	l I	-	132.0	! I	:	 	76.9		1	2.0
1321	200	 		132.1	I I	!!	! !	77.2		1	5.3
1322	210	!		132.1	!	!!	1	77.3		1	2.0
1322	212			132.2	!		!!	77.5		1	5.3
13.2	214	1		132.2	-	1 1	!	77.6		-	5.4
1321	216	i i		132.2	!	!	!	77.9		-	5.5
132.9	218	!		132.1	!	i	!	77.8		Į.	5.8
1344	220	!		132.9	1	!	1	77.3			5.5
135.2	222	1		134.4	!	[77.3		- !	5.1
135.2	224	!		135.2	1	1	!	77.2		-	4.7
135.3	226	1		135.2	!	1		77.4	1	!	5.1
135.3	228	1	1	135.3	!	I I	1	79.5	1		5.0
135.4	230		1 1	135.3	!	1	-	79.5	-	!	5.0
135.4	232	1	:	135.4	!	!	1	79.9	;	- -	4.9
136.3 80.0 79.9 79.9 137.6 79.9 137.6	234	1	1	135.4	1	[80.0	!	-	4.9
137.6	236	ŀ	1	136.3	!	1	1	80.0	!	-	5.1
138.8	238	1	1	137.6	1	I I	!	79.9		 	5.0
140.6	240	ł ł	1 1	138.8	!	!	1	80.2	2		5.0
140.4 84.5 140.5 87.2 140.5 87.2 142.1 87.2 144.0 89.5 145.5 96.4	242	1		140.6	1	!	!	81.0			5.0
140.5 87.2 140.5 87.0 142.1 87.2 142.1 87.2 144.0 89.5 145.5 145.5 145.5	244	1	1	140.4	!	!	!	84.5	!		5.2
	246	1	1	140.5	1	;	1	87.2	!	!	5.2
142.1 87.2 144.0 89.5 93.7 93.7 96.4 96.4	248	1	1	140.5	!	1	1	87.0	!	!!	4.5
144.0 89.5 93.7 93.7 96.4 96.4	250	1	!!	142.1	1	1	!	87.2	ŀ	-	5.2
145.5 93.7 93.7 96.4 96.4	252	!	1	144.0	!	:	!	89.5	!	_	5.4
	254	1	1	145.5	1	!	!	93.7	!	1	5.6
	256	1		145.5		ł	!	96.4		1	×

TABLE 6-8

	Average of Cells	4 and 8	5.7	5.8	5.6	6.1	6.1	5.9	6.4	9.9	6.4	6.7	8.9	6.9	7.1	6.9	6.5	1.9	6.5	6.4	6.4	6.2	6.3	6.5	6.4	6.4	6.4	9.9	9.9	6.8	9.9
	ol Reactor Cell 8	With Nut's	;	1		1	l	1	1	!	1	1	!	1	1	1	1	1	1] 	1	!	t I	1	1	!	1	1	!	!	1
	Sterile Control Reactor Cell 4 Cell 8	No Nut's.	1	!	!	1	1	1		1	1	!		1	1	1	1	1	I I	!	!	1	1	1	1	!	1	 	1	!	
	Average of Cells	5,6 and 7	9.76	2.76	2.7.6	97.9	2.76	99.4	93.6	101.8	101.8	101.8	102.0	102.0	101.8	103.9	104.3	104.4	104.5	104.4	104.3	104.4	104.4	104.6	104.7	105.0	105.0	105.0	105.0	105.1	105.3
ion (mg/L)		C=11.7		!	1			1		1	1	1	1	:		1		1	-	l f	1	I	!	!			1	!	!		I I
Cumulative Oxygen Consumption (mg/L)	Bioactive Reactors No Nutrients Added	Cell 6	!!	: 1	!!		1	l i	1	1]	!	1	1	!	i i		1	1	ŀ	1	ļ	!!	1	!	1	1 1	!	11	!	!!!
Cumula		Cell 5	1	1	!	1	!	1	ļ	! !	!	1	1	1	1	1	i	;	!	1	1	1	!	1	1	1	1	1	1	!	
	Average of Cells	9	145.5	145.5	146.3	147.0	148.5	148.5	148.5	148.5	148.5	148.7	149.2	150.8	151.2	151.4	151.4	151.3	151.3	151.3	151.2	151.3	151.5	151.5	151.5	151.5	151.5	151.6	151.6	151.6	151.4
	ors us Added	Cell 3	1	!	!	1	1	1	1	1	1	!	[]	[]	! 1	1	1	1	1	!	1	!	1	!	!	1	1	1	 	! !	1
	Bioactive Reactors en and Phosphorus	Cell 2	!	1	1	 	!	!	!		i I	<u> </u>	1	!	!	 	1	 	1	i	}	1			1	1	1	!	1	1	1
	Bioactive Reactors Nitrogen and Phosphorus Added	Cell 1	! !	:	 	1	 	 	1 -	1	!	1	!!	i	1	! !	1	i 	i	1	! ;	1	 	1	1	!		!	1	!	1
	Time	(br)	258	260	262	264	506	268	270	272	274	276	278	280	282	284	286	288	290	292	294	296	298	300	302	304	306	308	310	312	314

Notes:

1) Set 2 amended with 100 mg/L nitrate – nitrogen and 20 mg/L phosphorus.

Sterile controls amended with 2 percent mercuric chloride (20 g/L).
 Heavy stir-bar wear observed in all reactors. Iron oxidation visible in sterile control reactors.
 Reactor operation terminated after 86 hours to prevent reactor destruction and sample compromise. Samples transferred to shake—flask reactors (4-Liter serum bottles).

Table 6-9

Results of Initial Contaminant Analyses Shore Realty Site Glenwood Landing, New York

Analysis		Nutrie	Soil
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ons	nple	Vate ug/L	10000
Cons	ample	Wate (ug/L	<10000
Cons	Sample	Wate (ug/L	<10000
Cons	Sample	Wate (ug/L	<10000
Cons	d Sample	Wate (ug/L	<10000
Cons	led Sample	Wate (ug/L	<10000
Cons	ided Sample	Wate (ug/L	<10000
Cons	anded Sample	Wate (ug/L	<10000
Cons	nended Sample		<10000
Cons	mended Sample		<10000
Cons	Amended Sample		<10000
Cons	t Amended Sample		<10000
Cons	nt Amended Sample		<10000
Cons	ent Amended Sample		
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Cons	trient Amended Sample		
Cons	Iutrient Amended Sample		<10.0 <10000
Cons	Nutrient Amended Sample		
Cons	Nutrient Amended Sample		
Cons	Nutrient A		
Cons	Nutrient A		
Cons	Nutrient Amended Sample		
Cons	Nutrient A		
Cons	Nutrient A		

TPH (EPA 8015) (1) < 10.0 BTEX (EPA 8020)	<10000		<1.0	<1.0	<1.0	V10
TPH (EPA 8015) (1) BTEX (EPA 8020) Benzene Toluene Ethylbenzene Total Xvienes	< 10.0		< 0.0012	0.0042	< 0.0012	0.001
	TPH (EPA 8015) (1)	BTEX (EPA 8020)	Benzene	Toluene	Ethylbenzene	Total Xvlenes

<5.0<5.0<5.0<0.0

<0.0012 <0.0012 <0.0012 0.011

Notes:

1) TPH analysis performed by NEI (Port Washington, NY) using hexane as an extraction solvent. J - Concentration quantified below detection limit.

xylenes (0.59 mg/L and 2.00 mg/L, respectively) were noted in the aqueous phase of the nonutrient treatment condition only.

At the conclusion of the respirometry study, all analytes were reduced to below detection limits in the biologically active reators; however, low concentrations of xylenes were detected in both the soil and aqueous phases of the sterile control reactors. Additionally, diesel-range petroleum hydrocarbons were shown to persist in the sterile controls, while these hydrocarbons were not observed in any of the biologically active reactors. These results confirm that biological processes are capable of reducing concentrations of the constituents of concern at the site.

It should be noted that initial constituent concentrations in the soils and groundwater were lower than desirable for this type of study. The soil sample used for the respirometry studies was a composite sample consisting of equal portions of the water table and saturated zone samples from each of the soil borings (a total of six individual samples). Only three of these samples were subsequently found to have significant concentrations of purgeable aromatic compounds detectable by Method 8020, and none of the samples exhibited TPH concentrations detectable by NEI's analyses (Table 6-10). Thus, the concentrations of the target compounds for the respirometry study were reduced through a combination of dilution and volatilization as a result of the formation of the composite sample. Therefore, these data do not allow direct evaluation of the effect of nutrient supplementation on enhanced organic constituent biodegradation. Nevertheless, the increased oxygen utilization resulting from nutrient addition suggests that the rate of organic constituent biodegradation is likely to be increased by supplementing ambient levels of inorganic nutrients. The Basis of Design for the remedy at the site will therefore incorporate nutrient addition as a component of the *in situ* bioremediation process.

6.3 WATER TREATMENT

The objective of the water treatment evaluation was to assess the various treatment processes for the effective removal of the potential constituents-of-interest, specifically iron and various volatile and semi-volatile organic compounds. Many chemicals could be used for the precipitation of iron. However, the principal considerations are the implementability and economics of the process. The study below evaluated iron removal methods for optimum removal of iron. In groundwater, iron is present in its reduced form as ferrous hydroxide. Oxygen oxidizes it and precipitates it as flocculent ferric hydroxide. Under natural conditions, atmospheric oxygen can oxidize ferrous hydroxide, but its oxidation and precipitation are

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Table 6-10

Results of Week 3 Contaminant Analyses Shore Realty Site

Glenwood Landing, New York

					Cor	stituent C	Constituent Concentrations	SUO				
Parameters	Nu	trient Ame	Nutrient Amended Samp	oles	Samples	Samples Not Amended W		ith Nutrients	S	Sterile Cont	trol Samples	w
	Sam	Sample 1	Ѕашр	ple 2	Sample	le 1	Samı	Sample 2	Sample 1	ple 1	Sample 2	le 2
	Soil (mg/kg)	Water (ug/L)	Soil (mg/kg)	Water (ug/L)	Soil (mg/kg)	Water (ug/L)	Soil (mg/kg)	Water (ug/L)	Soil (mg/kg)	Water (ug/L)	Soil (mg/kg)	Water (ug/L)
											10 8	
NEI Data TPH (EPA 8015) (1)	<10.0	<10000	<10.0	<10000	<10.0	<10000	<10.0	<10000	<10.0	<10000	<10.0	<10000
BTEX (EPA 8020)												
Benzene	<0.002	<1.0	<0.0017	<1 .0	<0.0012	<1.0	<0.0018	<1.0	< 0.0015	<1.0	<0.0018	<1.0
Toluene	<0.002	<1.0	<0.0017	<1.0	<0.0012	<1.0	<0.0018	<1.0	<0.0015	<1.0	<0.0038 B	<1.0
Ethylbenzene	<0.002	<1.0	<0.0017	<1.0	<0.0012	<1.0	<0.0018	<1.0	< 0.0015	<1.0	<0.0018	<1.0
Total Xylenes	<0.002	<1.0	<0.0017	<1.0	<0.0012	<1.0	<0.0018	<1.0	<0.0023	2.2	<0.0093	2.2

Notes:

NS NS

2 2 3 3 3 3 3

SN SN

12 J 100

NS NS

< 20 < 25 < 25

SS SS

8 8

NS NS

<20 <25

SZ SZ

<20 <25

ARI Data TPH (EPA 8015 mod.)(2) Gasoline Range

Diesel Range

- TPH analysis performed by NEI (Port Washington, NY) using hexane as an extraction solvent.
 TPH analysis performed by Analytical Resources (Seattle, WA) using methylene chloride as an extraction solvent. Gasoline range TPH includes C12 to C24 hydrocarbons. Diesel range TPH includes C24 to C36 hydrocarbons.

 - NS No sample.

 J Concentration quantified below the detection limit.

 B Detected in method blank.

protracted. Conventional treatment involves the addition of an oxidizing chemical to enhance ferric hydroxide formation. In addition to the oxidants, coagulant aids are often added to improve the settling characteristics of the hydroxide flocs in the solid/liquid separation. Co-oxidation of organics is an indirect benefit achieved in the chemical oxidation. Many chemicals can be used for the precipitation of iron. However, implementability and economy dictate the final choice of chemicals for the removal of iron by chemical precipitation.

Oxygen present in air, ozone (O₃), chlorine (Cl₂ or NaOCl), hydrogen peroxide (H₂O₂), and potassium permanganate (KMnO₄) could achieve oxidation of the ferrous iron (Fe²⁺) to the ferric iron (Fe³⁺) and its subsequent precipitation as Fe(OH)₃(s). Potassium permanganate (KMnO₄) was selected as the oxidant of choice for the following reasons:

- (1) Chlorine addition could potentially result in the partial breakdown of specific organics as well as the formation of chlorinated organics (e.g., chloroform). Additionally, chlorine supplied as a gas would require a separate storage building. Also, special handling provisions would be needed to insure against the rupture and subsequent explosion of the gas cylinders.
- (2) Air oxidation is relatively slower than the other oxidants near a neutral pH range (i.e., hours as opposed to minutes). Thus, to achieve faster reaction kinetics, pH adjustment of a value greater than 9.0 would be required. After such oxidation of the iron, the groundwater pH would have to be adjusted down to a more neutral range prior to treatment.
- (3) Ozone requires significantly more capital equipment than the other oxidants, and would entail use of a contractor, an ozone generator, and cooling water and thus it would be more costly.
- (4) Hydrogen peroxide (H₂O₂) is most reactive with iron within a pH range of 4-6; thus, acid addition and pH control would be required to adjust and control the groundwater within this range. Also, H₂O₂ addition could result in partial breakdown products of the specific organics in the Site groundwater as well as result in an increase in the H₂O₂ requirement above the stoichiometric amount required for iron oxidation due to potential reactions with organics in the groundwater.

(5) Potassium permanganate (KMnO₄) is a proven and widely-accepted oxidant for iron removal. Additionally, the oxidation potential of KMnO₄ is similar to that of Cl₂, O₃, and H₂O₂ and KMnO₄ are most reactive within a neutral pH range. Residual levels of manganese are typically less than one mg/L and must be monitored where an effluent manganese limit is applicable. While the reaction generates small amounts of acid, pH adjustment may be necessary when significant levels of iron are present.

Potassium permanganate (KMnO₄, FW- 158.0) is a granular crystal having a specific gravity of 2.703 g/cm³ and a bulk density of 1.607 g/cm³. Equation 6-1 is the iron oxidation reaction between KMnO₄ and Fe²⁺ when they are both at a weight ratio of 1:1. Iron oxidation results in the precipitation of both reduced Mn as MnO₂ and oxidized iron as Fe(OH)₃.

$$KMnO_4 + 3Fe^{2+} + 7H_2O - MnO_2(s) + 3Fe(OH)_3(s) + 5H^+ + K^+$$
 (6-1)

This reaction is rapid and independent of pH between a range of 6 to 10. It is noteworthy that since this reaction generates an acid, pH adjustment, via caustic addition, is usually necessary if influent iron levels are significant relative to the influent alkalinity.

6.3.1 Iron Removal Evaluation (Jar Testing)

In order to evaluate the oxidant and coagulant dosages, a series of jar tests was performed using samples of groundwater and various dosages of KMnO₄ and coagulant aids. The following equipment and materials were used in the jar test.

- six-paddle stirrer;
- potassium permanganate;
- cationic and anionic polymers;
- acid (usually sulfuric);
- base (usually NaOH);
- pH meter;
- stopwatch;
- glassware (pipettes, beakers, and graduated cylinders);
- jars (plastic and glass); and
- test strips for iron and manganese determination.

The evaluation proceeded in three (3) steps to determine the optimum dosage of chemical required.

Step 1 - Optimization of KMnO₄ Dosage

- A 0.2M KMnO₄ solution was prepared dissolving 3.16 grams of KMnO₄ in 100 ml. Five doses, namely, 50, 75, 100, 125, and 150% of stoichiometric requirements, were evaluated.
- To 500 ml groundwater water samples, a calculated quantity of 0.2 M KMnO₄ solution was added. The sample was rapidly mixed at a rate of 100 rpm for 1 minute, then slowly mixed at 20-30 rpm for 5 minutes, and then allowed to settle for 10 minutes. The pH was adjusted to within the range of 7.5 to 8.5, with a 10% by weight solution of NaOH. The volume of solution required was then recorded. The initial and final pH, along with soluble iron and manganese concentrations were measured. Test strips were used to determine the concentration of iron and manganese in the water and to select the best dose. The volume of the supernatant and the sludge were noted. TSS, VSS, and FSS in the supernatant and the sludge were measured, and the sludge production computed. Samples of treated and untreated water were sent to an approved lab for the analysis of soluble iron and manganese.

Step 2 - Optimization of Single Cationic or Anionic or Non-Ionic Polymer

The potential polymers that were evaluated in the jar test are:

- Drewfloc,
- Clearwater,
- Amerfloc, and
- Betz.

In this test, the performance of each polymer was evaluated with the combination of the previously determined optimal KMnO₄ dose. The optimum dosage of a polymer was selected from its ability to produce clear supernatant and good settling flocs.

Cationic polymers were first tried at 1, 2, 5, 7.5, 10, and 15 mg/l, and the optimal $KMnO_4$ dose and the best combination dose were selected. Then the anionic polymers were tried at 0.5, 1.0, 1.5, 2.0, and 5.0 mg/L, and the optimal $KMnO_4$ dosage and the best combination were selected. Finally, the non-ionic polymer was tried at 0.5, 1.0, 1.5, 2.0, and 5 mg/L with the optimum $KMnO_4$ dosage.

The test was repeated with the best dosage of each polymer. If a single polymer produced clear supernatant after 20 minutes settling, that polymer was selected for the final use. TSS, FSS, and VSS in the supernatant and the sludge were measured.

Step 3 - Optimization of Polymer Combination

Sometimes a single polymer produced only pin flocs which did not settle within any reasonable time. In such cases, a combination of cationic and anionic polymers was required.

Then as before, an optimum cationic polymer dose was first selected in conjunction with the optimal KMnO₄ dose. Various dosages of anionic polymer were tried with the best combined dose of KMnO₄ plus cationic polymer. The anionic dose which produced the best settling floc and clear supernatant was selected for the final design. TSS, VSS, and FSS were measured in the supernatant and sludge.

6.3.2 Jar Test Results

The sample as received exhibited a pH of approximately 6.4 units, and contained 25 mg/L soluble iron as determined by the test strips. Six treatability runs were conducted to evaluate pre-treatment of the groundwater for removal of soluble iron. Run #1 was conducted using only pH adjustment to oxidize the soluble iron, with cationic polymer added to enhance Runs #2 through #6 were conducted as previously described. Run #2 was performed to determine the optimal dosage for potassium permanganate (KMnO₄) addition in terms of the stoichiometric requirements based on the quantity of soluble iron present. This dosage was identified as 25% of the stoichiometric requirement. Run #3 was performed to determine the optimal cationic dosage to enhance separation of the oxidized iron after KMnO₄. This dosage was determined to be 5 mg/L. Runs #4 and #5 were performed to determine the optimal combination of cationic and anionic dosage to enhance separation of the oxidized iron after KMnO₄. This combination dosage was determined to be KMnO₄ at 25%, cationic polymer at 1 mg/L and anionic polymer at 0.8 mg/L. Run #6 was performed using each of the best performing treatment scenarios identified to compare the results. The three treatment scenarios included pH adjustment to 8.4 with NaOH and cationic polymer addition at 1 mg/L, 25% KMnO₄ pH adjustment to >7.5 and 5 mg/L cationic polymer, and 25% KMnO₄ with pH adjustment to >7.5, and 1 mg/L cationic polymer followed by 0.8 mg/L anionic polymer. Each of these three scenarios yielded adequate results with no detectable levels of soluble iron present in the supernatant as determined by the test strips. Copies of the laboratory data sheets and results summary sheets are provided as Appendix F. Results of confirming laboratory analysis will be available within two weeks. Since three treatment scenarios appeared to treat the

groundwater equally in terms of soluble iron removal, selection will be based upon economic considerations and confirmation of the removal efficiency. Both the technical and cost-benefit evaluations will be incorporated into the Basis of Design Memorandum. Selection of the specific treatment process will be incorporated into the final design process.

7.0 REFERENCES

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APPENDIX A

Boring Logs

SOIL BORING LOG B-93-1

9 Pond Lane Concord, MA 01742 (508)371-1422

								(000/371 142)	_
PROJE	CT NO): <i>3-10</i>	33-210	2		DRILLING CO.: Warren George	MP ELE	V.: ' (MSL)	
		formin			rust	DRILLER: Vincent Gandolgo	TOTAL	DEPTH: 15'	
					anding, NY	BORING ID: 6 3/4		CE ELEV.: ' (MSL)	
		E: <i>3-9</i>			E: 9:30	CASING ID:	WATER	LEVEL DURING DRILLING: 5'	
					TIME: 10:35	METHOD: Hollow Stem			
BORIN	G LOC	ATION			alty	LOGGED BY: Dieter Geithner			
DEPTH (feet)	RECOVERY (%)	SAMPLE DEPTH	PID HEADSPACE (ppm)		LITHOLOGY	DESCRI			
	83		О	10 10 10 12		Brown fine to medium SAND, trace f trace shells	ine grave	el	
	33	93-1A	12	6 7		Grey fine SAND			‡
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REMEDIATION TECHNOLOGIES, Inc.

SOIL BORING LOG

9 Pond Lane Concord, MA 01742 (508)371-1422

						B-93-3		(508)371-	1422
PROJE	CT NO	· 3-10	33-21	2		DRILLING CO.: Warren George	MP FI F	V.: ' (MSL)	
		forming		_	rust	DRILLER: Vincent Gandolgo		DEPTH: 49'	
					anding, NY	BORING ID: 6 3/4		CE ELEV.: ' (MSL)	
STAR					ME: 10:45	CASING ID:		LEVEL DURING DRILLING: 7	· -
					TIME: 12:10	METHOD: Hollow Stem			
		ATION				LOGGED BY: Mike Devir		3	-
DEPTH (feet)	RECOVERY (%)	SAMPLE DEPTH	PID HEADSPACE (ppm)	BLOW CTS. / 6"	LITHOLOGY	DESCRIPT	ION		
5		93-3A		4 2 2 2 6 5 3		Dark brown loam Damp brown fine to medium SAND Tan to light brown fine to medium SAN Saturated dark brown medium to fine S			5
				1		Tan silty fine to medium SAND, little fi	ine to	medium gravel	
1				1		1 on sirty time to medium SAND, little II		medium gravei	
10-									-10
15—	17	93-36	N/A	1 0 1 0		Saturated brown fine to medium SAND,	-		15
				9		Brown fine to medium SAND, trace fine	to me	dium gravel	
20-	83		4	8 8		Redish fine to medium SAND and fine t Tan to reddish fine to medium SAND	o medi	um GRAVEL	-20
1	Ì			10		Tan fine to medium SAND			
-	100	93-31	N/A	17 17					
				16		Yellow fine to medium SAND			
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SOIL BORING LOG B-93-3

9 Pond Lane Concord, MA 01742 (508)371-1422

Page 2 of .

						В 93-3		(508)371-142	2
PROJE	CT NO	. 3-10	33-01	0		DRILLING CO.: Warren George	MP FI F	.V.: ' <i>(MSL)</i>	
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					TIME: 12:10	METHOD: Hollow Stem			
BORIN						LOGGED BY: Mike Devir			
DEPTH (feet)	RECOVERY (%)	SAMPLE DEPTH	PID HEADSPACE (ppm)	BLOW CTS. / 6"	LITHOLOGY	DESCRIP			
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APPENDIX B
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Address & Pand Land

CONCORD MA 01742

SHIP TO: Nytest Environmental Inc.

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Port Washington, NY 11050

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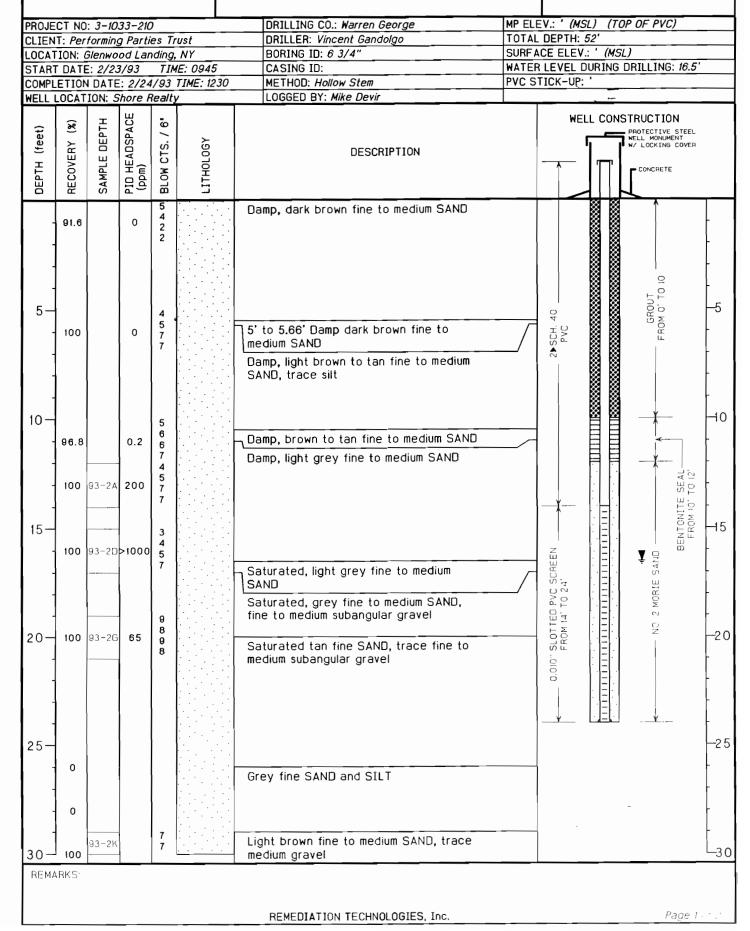
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APPENDIX C

Well Point and Monitoring Well Log

WELL INSTALLATION LOG Soil Boring WT-93-2

9 Pond Lane Concord, MA 01742 (508)371-1422



WELL INSTALLATION LOG Soil Boring WT-93-2

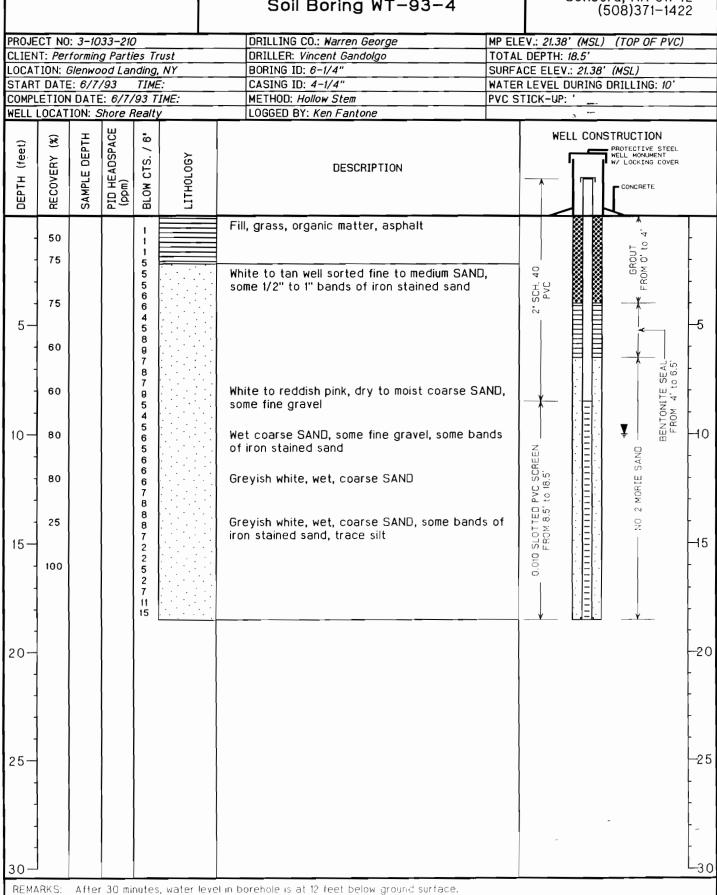
9 Pond Lane Concord, MA 01742 (508)371-1422

PROJE	CT NO	3-10	33-210)		DRILLING CO.: Warren George		EV.: (MSL) (TOP OF PVC)
CLIEN	T: Per	forming	g Parti	es Ti		DRILLER: Vincent Gandolgo		DEPTH: 52'
LOCAT						BORING ID: 6 3/4"		ACE ELEV.: ' (MSL)
START					1E: 0945	CASING ID:		R LEVEL DURING DRILLING: 16.5'
					TIME: 1230	METHOD: Hollow Stem	PVCS	TICK-UP: '
WELL I	LOCAT	ION: 5		Realt	<u>/</u>	LOGGED BY: Mike Devir		-
DEPTH (feet)	RECOVERY (%)	SAMPLE DEPTH	PID HEADSPACE (ppm)	BLOW CTS. / 6"	LITHOLOGY	DESCRIPTION	_	WELL CONSTRUCTION
35-	91.6	93-2K	>10	6 6 3 3 3 4 2 3 3 3		White to light grey fine SAND, some silt		- - -35 -
40-	100		5	6 5 5 6 5		Saturated light brown to tan medium to fine SAND White to light grey fine to medium SAND, some silt		-40 -40 45
50-	100		0.5	7 8 7		Light brown fine to medium running SAND, some silt Grey fine to medium SAND, little silt, trace medium gravel Orange fine to medium SAND, little silt Grey fine to medium SAND		- - -50
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REMA	RKS:					REMEDIATION TECHNOLOGIES, Inc.	_	Page 1 - 13

WELL INSTALLATION LOG Soil Boring WT-93-4

9 Pond Lane Concord, MA 01742 (508)371-1422

Page Lot 1



REMEDIATION TECHNOLOGIES, Inc.

APPENDIX D

Soil Gas Survey Field Data Sheets



to in the Rain" - A unique All-Weather till paper created to shed water and subset the written image. It is widely used bughout the world for recording critical field a in all kinds of weather.

pule in a variety of standard and custom ted case-bound field books, loose leaf, all and stapled notebooks, multi-copy sets mputer papers.

ten the Rain" All-Weather Writing Papers also available in a wide selection of rolls sheets for printing and photocopying.

a product of

.. DARLING CORPORATION COMA, WA 98421-3696 USA "Rite in the Rain"

ALL-WEATHER LINE RULE

Notebook No. 391

SHORE ZEALITY
Long Island N.Y
3-1033-210

1. PATTOCK My Calhoon te in the Rain_ EATHER WRITING PAPER ®

ne Protective Slipcovers (Item #31) are available for this style of notebook.

notebook from wear & tear. Contact your dealer or the J. L. Darling Corporation.

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	BH 5	0-2" CYSC:-4
-	35" Benzene > 50 ppm	2"-38" Derk Braun Son. 35" 44" Chareo! Sand
· · · · · · · · · · · · · · · · · · ·	Percharethylan ND	38 44" Charcot - 5:07
	2/9	
-	11:24 10-	* water 31"
	346	7
<u> </u>	30"	0-2" organics
	Benzene 750ppn Perchanthylan ND	2" -15" grave sond Brun
(Trichoratherlan ND	15"-44" Blank sand som
	methylan charden 2 100 pp	n
	12.03 pm	# 30" water
	BH 7 Near felephone p	·lų
	Benzene 750 ppm	0-2 organics
	motifien aborde toge	Brown
•	Tank Pad	2) "- 42 sand a grave!
	348 50 EAST BHZ	Dank discolored Howay occur
·	22	Orlshorn
	Benzene 7 50 pm	* 30" water
<u> </u>	Mathaghan chanda 6100 ppm	0-4" Black 70P
		4-18" Dark Brown to Bisc
		18-42 dark Groy Brown &
	u u	Frist oler x 22" water
· · · · · · · · · · · · · · · · · · ·		- C Mahar
The state of the s	المريب المراب المستعدمة المستملة مستكار المراجاة كالمراجية والمتعالية المتعالية المتعالية المتعالية المتعالية	the state of the s

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	307pm tank Feb
· · · · · · · · · · · · · · · · · · ·	BH9 25 Fast BH8 0-6" Black TOE
	45 Benzene TRACE 6"-20" Light Brown 20"-33" Dark color 5000
	20"-33" Dark call
	30' Boncon TRACE
	83 -44 chapter
	methyla dronda ND 45" 67" Grayani,
	See 60th with table
	in hole water 30"
	3154
	BH 10 25 south BHS OH" Black Top
	4"-24" No 1/ Ribur
-	motherefly chande ND 24"-42" Dark charcol
	motherefor chards ND 24"-42" Dark charcol sand 32" TARKLE
	* water 20" hohe
	<i>i</i>
	BHIL 25' West BHIO
	75" Benzene 15 mm O-4" Black Tox
	motherfor abounds 4'-2" Black 54
	- 3 revel
	21'-JE Dark same
	4 Sravel
	36-43 Dark som
-	to 27" work

	Near I case Door
	BH 12 25' 60th BH 11
	36" Danzone < 5 ppm O4" B/44 Tap
	motherwhen charicle TRAKE 4- C4 Tan said
	24"-38 Bat Boms
	38"- 43" crey sand
	* water hole 35"
	5,75
	3/2/83
	8,00 IN Garger Floor
_	BH 13 25 500th BH1 0-6" COMICA
	41" Benzons ND 6-19 Sands
	19-30 Brown 50
	44 Benzene TRACE 30"- 35" Black som
	Perchovethyles ND 392-41 Brown son
	mathematica dorula NE 41-58 Brown 50.
1	58"- 27" Grey Son
<u>:</u>	# worder taked 67"
* (IN FROM IS pump statum & oble water 44"
	BH 14 25' west BH &
	Revolvably has No 10"-18" Back Top Revolvably has NA 10"-18" Sound a conn
	Revolvablyhow ND 10"-18" Band & cren
	Methergula drondy > 2000/8"-29" Dark Opole Here Hat Hole 29"-42" Tand
	Very Hot Hole 29"- 42" Thank

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• ;	•
	9.'46 Am
	BH 15 25' work BH14 10-14" Black top
	12 15 25 west 8144 11 15 25 west 8144 11 15 25 west 8144 11 15 15 15 ppm 0-4" Black top 11 15 15 mentantyne chande > 2000 4"-17" I zwn 50m2 12 12 13 14 bille 17 13 2" Gray 23-00
	of to mentalyne chande > 2000 4"- 17" I awa Sount
	very Hot hate 17"-36" Gray 23-on
	6 9200
	Beside Trailor Orders 31"-42" light Ton 50
	BH 16 25' South BH 18 Life Bank Bank Top 10 - 42" Dark Brun - 5
	Soppa 0-6" Bfack Top
	1 - 43" Dark Brun - 5
	the above
· ·	10:24
l	BH 17 25 50 th BH 16
	42" Ben an TRACE 0-10" Back Top
<u> </u>	methogh dronder NO 19"-36" sond - In
	sine stames
	Between Trales and others 36-42" Brown 500
	53" 67 arey semy
	to water to be 52"
	# weder hate 45"
· · · · · · · · · · · · · · · · · · ·	
	26. Zin

a service de la companya de la comp	Base of Hill
	RH 18 25' North BH8
	33" Bonzan TRAGE 0-2' organic wests
	2" - 42" Brown send
	- som # whe table 35"
	12.5 BASE OF HILL 10 water hole 33"
	B14 19 25' FLOOL BH3
	HZ" Barreren ND 0-2" Organic everly
<u> </u>	35" Bareau PACF 2"-42" Brown send
	42"-67 Brown sand
<u> </u>	& under table 50"
·	1:02 on hill side A hole noter 30"
	BH 20 25' EAST BHY
	72" Benezure ND 0-72" Tam Gans
	9'6" Benene 8 ppm 0-144 Tou sand
	10' Benen 15ppm 144-128" Fory 5>
	mother TRACE * wester type 12
·	ZIZpm # 10' hole mater
	BH ZI Z5' EAST BH5 ON HUM Side 2-72" Bornon Some
<u> </u>	ON HIM SICC 3-12. Sond
· · · · · · · · · · · · · · · · · · ·	72" Benozer > 50ppn 72"-120" Brian
·	8 Benon > 50 ppm 170"-144" 6my Sent
	8' Potrolair Hydro aubn 260 ppn to water 120"
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	1

	Du 22 25/4
	BH 22 25 East 346-
	ZH" Benzen Boppm 0-2" Object
	z"- 41" Dark &
	of water hole 24"
•	3/3/93 on top before pière
	8:00
	RH 23 56' F. L BH 20
	72" Benezona ND 0-6" Dark or = 1
·	144" Benezen > 50 ppm 6"-36 5400 nson
	216" Pertstown hydrocombon 1200 pag 36" -72" Sand
	72: 148" sand
ù.	140"-144° cape
	144-216" Grove
	9:30 corn of dike Heave odo-
	D11, 24 25' south B23
	72" Banazan 750ppm 6-6" Dark again
	Potrolun hydrombo Trace ("-60" sand mile of orey o
	of orey o
•	144" Benow > 50 ppm 72 - 189 grey occ
-	1, 1, 1, 1, 1, 1, 1, 1, 1, 1, 1, 1, 1, 1
	109 139 1.6 1000
	139 - 144 tanih Bours
-	(FIRST # 202 %
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e de la companya de l	
	Near Light 10,30
	BH 25 25 south BXZ4
	77" Beneza ND 0-68" sand as 50
<u> </u>	(8"-72" while is
	72"-118 Sancing
	Inside Dike Aren 118"-126" concret-
·	11:19
-	BH 26 25 Ext of 8425
	72" Benerame ND 0-6" Black sind
- <u>_</u>	144" Benera 750ppm 6"-72" Brown Somi Black tube Very Hot 72"-144 Brown Som
<u> </u>	
-	144" Pokum Hydrocarkon 200 pm a atok
<u> </u>	
- <u> </u>	1/36 pm on dike Top
<u> </u>	BH 27 50' East of BH 26
	144" Bearing ND " 7" 2"
-	6-12 8000
- 	27 Store 1 50ppm 12 -149 Bruse 1 150 - 216 Dark Bounts 15191 5911
	Dark Bount
	on Road to tank Howay sour
	B28 50 East of BAZ7
	72" Benezen ND 0-3" organi scuis
	3"-67" Brown End
	# 5 how water 12"-138 Dark James water N' 138"-144 Dark James
	water N 12-138 Dark 3mm
	water N 138 1440 Darking
•	

 BH 29 25 Enot 8127
194" Benzon ND 0-3" cozon
216" Benown 750ppm J"-72" Trongs
 72 -144" Brown
 144"-180" Brown
180"-216" Bind
12 Hear
 a Bittle at Tank Farm side Dike last above water
 BH 30 100' NORTH BHZS
 BH 30 100' NORTH BHZS 144" Renezem ND 0-3" Organi
216" Burn ND 3"-72" Barn
22' Bonaron 13 ppm 72"-120" Do-1230
120"-144" Dark
1447 - 180" Dark
5:45 180"- 214" Bixix
216"-264" any 5=
508) 329-9279 264-288 Grov Fa
 Chare / suteritz
 314193 7:45 3/4/93
 3 H 31 25' West BH30
144" Bonezona > 50pm 6-3" ergonec
3"-72 Brown 61
 72" -120" Brown sa
120-14411 +47 5

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······································	Inside Dike on wall Northwest come
	BH 32 180' North 3423
	144" Benezame ND 0-6" Organics som? " 15' Brown 750pm 6"- 72" Brown From
	12' Potralus Mydrus 100ppn 72"-120" Branson.
	120 - 144" light sons
	120 - 144" light sond
	on bike not se when table 17'
	BH 33 50' East from BH 32
	"
<u>-</u>	74411 Benevane > 50pm 6-6" 010 and 60.
	P 129 Brown 5 - 729 Brown 5 - 7
	72"-107 tom Gard
	109-120 Red sang
<u> </u>	120-144 for 6and
	·
	BH 34 50' North 3433
	18 Bonneau NB 0-6" disancesan
	24' Brave ND 6-72' Brave out
•	72"-144 Brown sam
	144" - 216 Brown ou
	71/ - 298 Born Sen
<u> </u>	
· · · · · · · · · · · · · · · · · · ·	<u> </u>

APPENDIX E

Bio-treatability Study Data Sheets

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	1

RI	JEC	, L	ABO	RAT	ORY	•	
	ON TECHNOLOG		TRAV	ELER N	o. 93	<u> 303-</u>	027
DISTRIB	• • • • • • • • • • • • • • • • • • • •	☐ Microbiology				7	<u>.</u>
		,					
	,	ズ Treatability	•				
		☐ Chemical Analy	sis				
PROJECT	(CLIENT): SI	nore-Real	ty	DATE	RECEI	VED: 3	-18-93
PROJECT			/	DATE	COLLE	CTED:	3-109
PROJECT	CONTACT: +	e:d: Ander	SON	FX	eld	Som	ples
PHONE:	RETEC	, Seattle	2			eat ting	
DASH	SAMPLE	DESCRIPTION	MATE	RIX	SIZE-	QUANT	LOC.
01	WT-93-	2 Water	wa	ter	1-	1	Q-
02			1			 	· .
<u>03</u>							
09	<u> </u>			<u> </u>	<u> </u>		:
05	SW-2	Sample	Was	er_			:
06			: 				
<u>0.7</u>							
09	SW-6	Sample	Wate				
13			: 1				:
11							
12	<u> </u>		- للد	-	7	_	
	C == 1 S a					cal	



TABLE

Number of Total and VOA Degrading Microorganisms in Shore Realty Water Sample.

Sample ID ^a	Total Microorganisms CFU/ml of water ^b (10 ⁴)	VOA Degraders ^c CFU/ml of water (10 ⁴)
9303-027-02		
Mean +/- Std. Dev.	27 +/-11	0.05
9303-027-06		
Mean +/- Std. Dev.	38 +/-9	7.0
9303-027-10		
Mean +/- Std. Dev.	32 +/-5	1.3

^a Results represent the mean value and standard deviation of triplicate platings.

^c VOA degraders represent the cell growth in the presence of a VOA substrate minus the cell growth in the absence of any carbon substrate.

Released by

Heidi Anderson Project Scientist

^b Results represent colony forming units (CFU)/ml of sample.



LABORATORY

TRAVELER NO. 9304-016

REMEDIATION TECHNOLOGIES

DISTRIBUTION

☐ Microbiology

□ Treatability

☐ Chemical Analysis

PROJECT (CLIENT): Short Realty

DATE RECEIVED: 4-14-93

PROJECT NUMBER:

3-1033-777

DATE COLLECTED: 2-23-93

PROJECT CONTACT: Mark Lorsen

Shore Realty Samples from

PHONE: RETEC, Seattle

NYTest Labo

DASH	SAMPLE DES	CRIPTION	MATRIX	SIZE-QUANT	LOC.
01	13-93-1A	(2'-4')	5071	402	R-4
02	8-93-18	(2'-4')		802	. 1
03	· B- 93-1D	(4'-6')		402	<u>.</u>
04	B-93-1E	(4'-6')		800	:
05	B-93-16	(8'-10')		404	:
06	B-93-1H	(8'-10')	•	8oz	:
07	B-93-2A	(12'-14)		402	
08	B-93-2B	(12'-14')		804	
09	B-93-2D	(15'-17')		402	
10	B-93-2E	(15'-17')		802	
1)	B-93-26	(19'-21)		402	
12	B-93-2H	(19'-21')		802	:
13_	13-93-3A	(3'-5')		402	
14	B-93-3B	(3'-5')		802	
1.5	B-93-3D	(7'-9')		400	
16	B-93-3E	(7'-9')		802	
17	B-93-36	(11'-13')		4 02	·
18	B-93-3H	(11'-13')	<u> </u>	802	



9 Pond Lane Concord, MA 01742 Telephone: (508) 371-1422 Facsimile: (508) 369-9279

April 12, 1993

Mr. Mike Brenen NYTEST 60 Seaview Blvd, P.O. Box 1518 Port Washington, NY 11050

Dear Mike:

Please reference our telephone conversation on March 25, 1993. The following is a list of soil samples that you said you could ship to our Seattle lab:

B-93-1A (2'-4') 15977-0(B-93-2E (15'-17') 15897-05 15977-06 B-93-1B (2'-4') B-93-2G (19'-21') 15197-03 15977 -02 B-93-1D (4'-6') B-93-2H (19'-21) 15897-06 B-93-1E (4'-6') 15977-07 B-93-3A (3'-5') 15922-0[B-93-1G (8'-10') 15477-03 B-93-3B (3'-5') 15922-06 B-93-1H (8'-10') 15977-08 B-93-3D (7'-9') 15922-02 B-93-2A (12'-14') [5897-0] B-93-3E (7'-9') 15922-07 B-93-2B (12'-14') 15897-04 B-93-3G (11'-13') 15922-05 B-93-2D (15'-17') 15897-02 B-93-3H (11'-13') 15922-10

RETEC's Federal Express account number is 1025-2239-7. Please mark the number 3-1033-210 in the "Your Internal Billing Reference Information" section of the Airbill. These samples should be sent Priority One for delivery on a weekday. The samples should be shipped to:

RETEC

1011 SW Klickitat Way, Suite 207

Seattle, WA 98134

ATTN: Heidi Anderson

Thank you very much for forwarding these samples for me. If you have any questions, please do not hesitate to call me at (508) 371-1422.

Sincerely,

Duter Geithner

Dieter Geithner

DG/mkg

File: DG-001/3-1033-210

Concord, MA • Pittsburgh, PA • Fort Collins, CO • Seattle, WA • Austin, TX • Chape! Hill, NC St. Paul, MN • Mandeville, LA • Billings, MT



nytest environment inc

CHAIN	OF	CUSTODY	RECORD
-------	----	---------	--------

	TOM P			Address 1011 SW FISS Seattle, WA 9813 Phone Attn. Heidi Anderson	
Project No. 93-19853	Project Nam	ne Note Realty.		Date Shipped 4/13/93 Air Bill No.	Carrier FED EX Prior
Sampler: (Signature	•)	NOTE Realty - Analytical Protocol		Air Bill No.	Cooler No.
Sample I.D.	Date/Time Sampled	Sample Description	No. Of Con- tainers	ANALYSIS R	EQUESTED
15977-01		Soil	1	Returning Sampl	les to Client
к -06)	1)	
11 -02			ì		
11 -07					
11 -03					
11 -08			1		
15897 - 61			i		
11 -04			1		
11 -02					
11 - 05			1		
03			1		
11 - 06			1		
Relinguished by (Signatur	•, ()	Date / T	ime Rec'd	By (Signature)	Date / Til
Print Name	000	4/3/2	Print 1	Name	
Relinquished by (Signature	eripes		ime Rec'd	by (Signature)	Date / Tir
Print Name			Print N	lame	
Relinquished by (Signature	:1	Date / T	ime Receive	ed for Laboratory by (Signature)	Date / Tu
Print Name			Print N	lame	
Special Instruction	ons/Comment	s Please Return	NEI	coolert 11/2 And 19	ie packs
	O	Λ (i	1.00		-



CHAIN OF CUSTODY RECORD Report to
SHIPTO: Nytest Environmental Inc. SHIPTO
REPORT O: Client Name. Address 1011 Kick 1-9- Luque suite 60 Seaview Blvd. Port Washington, NY 11050 (516) 625-5500 Attn. TOM Petrollo Project No. 93-19853 Date Shipped Project Name Shore FEDER Priority Air Bill No. Sampler: (Signature) Analytical Protoco Cooler No No. Of Sample Date/Time Sample **ANALYSIS REQUESTED** Con-I.D. Sampled Description tainers Seturning Sandes to client -01 - O6 1 -02 1 -07 ~ のブ -10 Rec'd By (Signature) Time Relinguished by (Signature Rec'd by (Signature) Time Print Name Print Name Relinquished by (Signature) Received for Laboratory by (Signature) Time Print Name Print Name Return NEI cooler # 412 Find ice packs Special Instructions/Comments Flease to Above Address via UPS grand service



TABLE 1

Number of Total and VOA Degrading Microorganisms in Shore Realty Soil Samples.

9304-016-09 B-93-2D (15-17')			3	
B-93-1A (2-4') Mean +/- Std. Dev. 9304-016-03 B-93-1D (4-6') Mean +/- Std. Dev. 102 +/- 45 9304-016-05 B-93-1G (8-10') Mean +/- Std. Dev. 6.4 +/- 1.0 1.1 9304-016-07 B-93-2A (12-14') Mean +/- Std. Dev. 14 +/- 0.3 71 9304-016-09 B-93-2D (15-17')	-	CFU/g of soil ^b	CFU/g of soil	
Mean +/- Std. Dev. 11 +/- 3.2 45 9304-016-03 B-93-1D (4-6') Mean +/- Std. Dev. 102 +/- 45 49 9304-016-05 B-93-1G (8-10') Mean +/- Std. Dev. 6.4 +/- 1.0 1.1 9304-016-07 B-93-2A (12-14') Mean +/- Std. Dev. 14 +/- 0.3 71 9304-016-09 B-93-2D (15-17')	9304-016-01			
Mean +/- Std. Dev. 11 +/- 3.2 45 9304-016-03 B-93-1D (4-6') Mean +/- Std. Dev. 102 +/- 45 49 9304-016-05 B-93-1G (8-10') Mean +/- Std. Dev. 6.4 +/- 1.0 1.1 9304-016-07 B-93-2A (12-14') Mean +/- Std. Dev. 14 +/- 0.3 71 9304-016-09 B-93-2D (15-17')	B-93-1A (2-4')			
B-93-1D (4-6') Mean +/- Std. Dev. 102 +/- 45 49 9304-016-05 B-93-1G (8-10') Mean +/- Std. Dev. 6.4 +/- 1.0 1.1 9304-016-07 B-93-2A (12-14') Mean +/- Std. Dev. 14 +/- 0.3 71 9304-016-09 B-93-2D (15-17')		11 +/- 3.2	45	
B-93-1D (4-6') Mean +/- Std. Dev. 102 +/- 45 49 9304-016-05 B-93-1G (8-10') Mean +/- Std. Dev. 6.4 +/- 1.0 1.1 9304-016-07 B-93-2A (12-14') Mean +/- Std. Dev. 14 +/- 0.3 71 9304-016-09 B-93-2D (15-17')	9304-016-03			
Mean +/- Std. Dev. 102 +/- 45 49 9304-016-05 B-93-1G (8-10') Mean +/- Std. Dev. 6.4 +/- 1.0 1.1 9304-016-07 B-93-2A (12-14') Mean +/- Std. Dev. 14 +/- 0.3 71 9304-016-09 B-93-2D (15-17')	B-93-1D (4-6')			
9304-016-05 B-93-1G (8-10') Mean +/- Std. Dev. 6.4 +/- 1.0 1.1 9304-016-07 B-93-2A (12-14') Mean +/- Std. Dev. 14 +/- 0.3 71 9304-016-09 B-93-2D (15-17')		102 +/- 45	49	
B-93-2A (12-14') Mean +/- Std. Dev. 14 +/- 0.3 71 9304-016-09 B-93-2D (15-17')	9304-016-05 B-93-1G (8-10')			
B-93-2D (15-17')		14 +/- 0.3	71	
		15 +/- 1.6	256	

^a Results represent the mean value and standard deviation of triplicate platings.

b Results represent colony forming units (CFU)/g of soil on a wet weight basis.

^c VOA degraders represent the cell growth in the presence of a VOA substrate minus the cell growth in the absence of any carbon substrate.

Released by

Heidi Anderson Project Scientist



TABLE 2

Number of Total and VOA Degrading Microorganisms in Shore Realty Soil Samples.

Sample ID ^a	Total Microorganisms CFU/g of soil ^b (10 ⁶)	VOA Degradets ^c CFU/g of soil (10 ⁴)
9304-016-11		
B-93-2G (19-21')		
Mean +/- Std. Dev.	9.1 +/- 3.1	38
9304-016-13 B-93-3A (3-5')		
Mean +/- Std. Dev.	8.5 +/- 0.4	1.7
9304-016-15 B-93-3D (7-9')		
Mean +/- Std. Dev.	0.72 +/- 0.09	0.76
9304-016-17		
B-93-3G (11-13')	70.405	0.20
Mean +/- Std. Dev.	7.8 +/- 0.5	0.30

^a Results represent the mean value and standard deviation of triplicate platings.

^c VOA degraders represent the cell growth in the presence of a VOA substrate minus the cell growth in the absence of any carbon substrate.

Released by:

Heidi Anderson Project Scientist

^b Results represent colony forming units (CFU)/g of soil on a wet weight basis.

		<u> </u>				
R	HIE	L	ABO	RAT	ORY	
	ION TECHNOLOGIES		TRAV	ELER N	10. 9304.	-033
DISTRIB		icrobiology				,
		reatability				- · .
	□с	hemical Analys	sls			
PROJECT	(CLIENT): Shore	-Realty		DAT	E RECEIVED:	
	NUMBER: 3-1C	,		DAT	E COLLECTED:	7-23
PROJECT	CONTACT: MON	tharsen		F.V St	nal (7-00 imulation	7) Tes
PHONE:	RETER,	seattle	-		Samples.	, 0,
DASH	SAMPLE DESC	RIPTION	MATE	ıx	SIZE-QUANT	LOC.
01	Stim Test	03/04 A	Sluce	7	200 mL	
02	_	03/04 B				
0.3		05/06 A			-	
04		05/06/3			<u> </u>	
05		09/10 A				<u> </u>
06_		09/108				
07		11/124				
08		11/12B		_		
09		15/16 A			2	
10		15/16B			-	
		17/18A		_		1
12	-\t-	17/18B	- H		*	
						1
	<u>.</u>					
		1				1

-()



TABLE 1

Number of Total and VOA Degrading Microorganisms in Shore Realty Slurry Samples

Sample ID ^a	Total Microorganisms CFU/mL slurry ^b (10 ⁶)	VOA Degraders ^c CFU/mL slurry (10 ⁵)
9304-033-01		
Mean +/- Std. Dev.	48 +/- 19	10
9304-033-02		
Mean +/- Std. Dev.	65 +/- 10	38
9304-033-03		
Mean +/- Std. Dev.	11 +/- 4.7	0.44
9304-033-04		
Mean +/- Std. Dev.	14 +/- 2.9	0.29
9304-033-05		
Mean +/- Std. Dev.	5.7 +/- 1.1	14
9304-033-06		
Mean +/- Std. Dev.	9.4 +/- 0.8	2.9

^a Results represent the mean value and standard deviation of triplicate platings.

b Results represent colony forming units (CFU)/mL of slurry.

^c VOA degraders represent the cell growth in the presence of a VOA substrate minus the cell growth in the absence of any carbon substrate.

Released by.

Heidi Anderson Project Scientist



TABLE 2

Number of Total and VOA Degrading Microorganisms in Shore Realty Soil Samples.

Sample ID ^a	Total Microorganisms CFU/mL slurry ^b (10 ⁶)	VOA Degraders ^c CFU/mL slurry (10 ⁵)
9304-033-07		
Mean +/- Std. Dev.	2.2 +/- 0.6	0.30
9304-033-08		
Mean +/- Std. Dev.	2.5 +/- 0.8	0.69
9304-033-09		
Mean +/- Std. Dev.	0.29 +/- 0.07	0.007
9304-033-10		
Mean +/- Std. Dev.	0.32 +/- 0.02	0.014
9304-033-11		
Mean +/- Std. Dev.	1.9 +/- 0.6	0.015
9304-033-12		
Mean +/- Std. Dev.	2.5 +/- 0.6	0.023

^a Results represent the mean value and standard deviation of triplicate platings.

Besults represent colony forming units (CFU)/mL of slurry.

^c VOA degraders represent the cell growth in the presence of a VOA substrate minus the cell growth in the absence of any carbon substrate.

Released by:

Heidi Anderson Project Scientist

R		
•		741

REMEDIATION TECHNOLOGIES

LABORATORY

TRAVELER NO. 9304-026

DISTRIBUTION	0	Ν	
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☐ Microbiology

☐ Treatability

MChemical Analysis Sub to ARI

PROJECT (CLIENT):

Shore-Realty

DATE RECEIVED:

PROJECT NUMBER:

3+033-777

DATE COLLECTED: 4-16

PROJECT CONTACT:

Mark Larsen

Nutrient Adsorption + 24 hour Nutrient Precipitation Samples

PHONE:

RETEL, Seattle

DASH	SAMPLE DESCRIPTION	MATRIX	SIZE-QUANT	LOC.
01	Nut. Adsorp. 5/6 Rct 1	water	108 mL	ARI
02	1 1		1	R-4
03	Rct 2			AZI
04	- L		TO THE PARTY OF TH	R-4
05	11/12 Rct/		111111111111111111111111111111111111111	ARI
<u> </u>				R-4
07 08 09	Rcf 2			ARI
08	\ \\ \\ \\ \\ \\ \\ \\ \\ \\ \\ \\ \\ \			R-4
09	17/18 Ret/			ARI
10				R-4
11	Rct 2			ARI
12	* * *			R-4
13	Control Ret 1			ARI
14	1 1			2-4
15	Rcf 2			ARI
16	A A	:		R-4
17	Precip 24 hir-Samples Ret	1		AP 1 R-4
18	1			R-4
19.	Ret	2	,	ARI
_ 20	1		9	R-4

03 May 1993

ANALYTICAL RESOURCES INCORPORATED

Analytical Chemists & Consultants

333 Ninth Ave. North
Seattle, WA 98109-5187
(206) 621-6490
(206) 621-7523 (FAX)

Mark Larsen Remediation Technologies Inc. 1011 S.W. Klickitat Way Suite 207 Seattle, WA 98134

RE: Client Project: 3-1033-777 Shore Realty; ARI Project: #D581

Dear Mr. Larsen:

Please find enclosed the original chain-of-custody record (COC) and results for the above referenced project. Ten water samples were received on 4/21/93, in good condition. The COC specified analysis for ammonia-N, however the precipitate that was present in the samples prohibited this analysis from being performed without refiltering the samples. Total Kjeldahl Nitrogen analysis was performed instead, as agreed upon on 4/22. The analyses proceeded without incident of note, and these results were faxed to you earlier today.

A duplicate and spike analysis were performed for both parameters on your sample ...-01. Results of these QC analyses have been included on the report, as you requested.

A copy of this package will be kept on file with ARI should you require any further information or copies of additional documentation. If you have any questions please feel free to call any time.

Sincerely,

ANALYTICAL RESOURCES, INC.

Kate Stegemoeller Project Coordinator 206-340-2866, ext. 117

Enclosures cc: file #D581

REMEDIATION TECHNOLOGIES precio tate North Lasser 1011 S.W. Klickitat Way Suite 207 SEND RESULTS TO: C7 40 Received by: Samples Preserved PC, 2,779 andrag OCCIES Do not P. Herro Syche H2504 O.e.r Date / Time ンくめ Sofe REMEDIATION TECHNOLOGIES INC Relinquished by: 1. 1. 1. 1. Date / Time CHAIN OF CUSTODY RECORD 15CC) CB/ NO. OF CONTAINERS Beceived for Laboratory by: Expected 03 Received by: Sanatano 0 40 0 ٢ Q 9304-026-01 CONTRO SAMPLE NO. RESOURCES Shor- Rea 82 Date / Time - Date / Time 2-2 Proked & By ENA 8 10x 1xx TIME PROJECT NAME 4-16 17/2 DATE RECEIVING LABORATORY:

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REMARKS

3-1033-777

PROJ. NO.

SAMPLERS:

Anal LAB I.D.

5778

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ART # 1358

Seattle, WA 98134 (206) 624-9349

PINK CODY Sampler

Relinquished by: (Signature

Shipper Information

Relinquished by: (Sign 10.44



Final Report Laboratory Analysis of Selected Parameters

Matrix: WATER

Data Release Authorized: Modellar

Report Prepared: 05/03/93 DWN

Project No: 3-1033-777 Shore Realty

QC Report No: Retec-D581 Date Received: 04/21/93

ANALYTICAL RESOURCES INCORPORATED

Analytical Chemists & Consultants

333 Ninth Ave. North
Seattle, WA 98109-5187
(206) 621-6490
(206) 621-7523 (FAX)

			DATE OF	ANALYSIS
Sample	Dat	ta:	04/26/93	04/23/93
Method		EPA 353.2	EPA 365.2	
		Number	SM 4500-NO3 F	SM 4500-P
			TKN	TOTAL-P
Lab	ID	Sample Number	(mg-N/L)	(mg-P/L)
D581	Α	9304-026-01	571	326
D581	В	9304-026-03	335	317
D581	С	9304-026-05	594	349
D581	D	9304-026-07	615	349
D581	Е	9304-026-09	582	341
D581	F	9304-026-11	575	331
D581	G	9304-026-13	578	349
D581	Н	9304-026-15	592	354
D581	I	9304-026-17	820	531
D581	J	9304-026-19	816	502

Method Blank Analysis:

		TKN	TOTAL-P
		(mg-N/L)	(mg-P/L)
Metl	iod Blank 1	0.2	< 0.010
Dete	ection Limit:	0.1	0.010

Check Standard Analysis:

	_	(mg-N/L)	(mg-P/L)
	Measured Value	496	0.084
	"True" Value	500	0.080
1	% Recovery	99.2%	105%

Duplicate Analysis:

	(mg-N/L)	(mg-P/L)
Sample ID	D581A	D581A
Original	571	32 6
Duplicate	5 7 7	323
RPD	1.05%	0.92%

Spike Analysis:

1		(mg-N/L)	(mg-P/L)
	Sample ID	D581A	D581A
	Original	571	326
	Spike	6 7 5	336
	Spike Level	100	10.0
	% Recovery	104%	100%

Comments:

R	16	7	3
DEMEDIATIO	MITEOU	INIOL OC	150

LABORATORY

		J.=	MDOI	MIC	/111	
REMEDIAT	ION TECHNOLOGIES		TRAV	ELER NO.	. 9305-0	203
DISTRIB	BUTION N	licrobiology				
	□ ⊤	reatability			3	
	,	•		_		
	 	hemical Analys	sis Suk	, to N	y Test L	cbs
PROJECT	(CLIENT): Shor	e Realty		DATE F	RECEIVED:	
PROJECT	NUMBER: 3-1	DATE COLLECTED: 5-4-93				
PROJECT	CONTACT: MC	rkLorse	\bigcirc	•) Shore	1
PHONE:	RETEC,	Seattle		, Sc	sp:rome mples	
DASH	SAMPLE DESC	RIPTION	MATR	IIX :	SIZE-QUANT	LOC.
01	T=O Reac	ter # \$ BTE	x Wate	er	2 × VOA	NYTE
02	1	TPA	, Wat	er	2 × VOA	<u>i</u>
03		BIEXTP	1 505/	1	1×60mL	
04	<u> </u>	Arch.iv	s War	ler	1 ×500ml	
<u></u>	T-OReada	-NO.BBTEX	Wafe	20	2 × VOA	NYTE
06		TPH	wat	es	2 × VOA	1
07		BTEXTA	4 501	,	1 × 60mL	1
08	4	Archive			1 x 500mL	<u>V</u>
09	composite '	T+0 50[]	Soil	<u> </u>	8×250N	AR
						:
-	·					
		<u>;</u>		-		
<u>i</u>						
:				:		ſ



07 May 1993

ANALYTICAL RESOURCES INCORPORATED

Analytical Chemists & Consultants

333 Ninth Ave. North
 Seattle, WA 98109-5187
 (206) 621-6490
 (206) 621-7523 (FAX)

Mark Larsen Remediation Technologies Inc. 1011 S.W. Klickitat Way Suite 207 Seattle, WA 98134

RE: Client Project: 3-1033-777 Shore Realty;

ARI Project: #D702

Dear Mr. Larsen:

Please find enclosed the original chain-of-custody record (COC) and results for the above referenced project. One soil sample was received on 5/5/93, in good condition. The analysis proceeded without incident of note, and these results were reported to you verbally yesterday.

As we discussed in our telephone conversation, the hydrocarbons in this sample appear to match the diesel pattern. The reported gas-range concentration is actually diesel, eluting at the early end of the pattern. I've enclosed the method blank and sample chromatograms to assist in your evaluation of the results. Also enclosed is a blank spike recovery report to provide QC documentation for the analysis.

A copy of this package will be kept on file with ARI should you require any further information or copies of additional documentation. If you have any questions please feel free to call any time.

Sincerely,

ANALYTICAL RESOURCES, INC.

Kate Stegemoeller Project Coordinator 206-340-2866, ext. 117

Enclosures

cc: file #D702

5376 Š

CHAIN OF CUSTODY RECORD

PROJ. NO. PROJECT NAME	PROJECT NAME	NAME		5	(7)		SEND RESULTS TO:
SAMPLERS:	7 0 2	9 7 X	really	PINERS	2 ₁ / ₁ / ₁		140-14-10-130-1
RECEIVING LABORATORY:	ATORY: 17	Res	Resources	т соит	FID		
LABI.D. NO.	DATE	TIME	SAMPLE NO.	NO. C	\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\		REMARKS
	2-4	1900	9305-003-09	_	7	Nee	ed ItaD
			-			day	La #54P (by Thur.
							1
							(200)
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						0.0	41-1-1-1-1-1-1-1-1-1-1-1-1-1-1-1-1-1-1-
						(/) //	
Relinquished by: Signature:	ignature.	2	Date / Time Received by: \(\mathcal{Synature} \)		Relinquished by: उद्गाराज्य	Date /	Time Received by: 5 : 7 :
Relinquished by: (Signatura)	ignature	Δ	Date / Time Received Tor Laboratory By	lei.	5/ Date / Time	13 M 13 M	REMEDIATION TECHNOLOGIES 1011 S.W. Klickitat Way
Shipper Information Probability	e e	4	P. Courier			R E M E D I A T I O N TECHNOLOGIES INC	Surte 207 Seattle, WA 98134 (206) 624-9349
	PINK COP	PINK COPY - Sampler	YELLOW COPY - Laboratory	- Laboratory		WHITE COPY - ReTeC	



ANALYTICAL RESOURCES INCORPORATED

Analytical Chemists & Consultants

233 Ninth Ave. North Seattle, WA 98109-5187 (206) 621-6490 (206) 621-7523 (FAX)

WA HCID Method by GC/FID

TOTAL PETROLEUM HYDROCARBONS

Matrix: Soils/Sediments

ARI LAB ID: D702

Data Release Authorized

Data Prepared: 05/06/93 MAC: Gat

Client: RETEC

Project: 3-1033-777

Shore Realty

VTSR: 05/05/93

Date Extracted: 05/05/93

		Date	Dilution	Gas	Diesel	Oil	Surrogate
Lab ID	Client Sample ID	Analyzed	Factor	Range†	Range*	Range°	Recovery
D702 MB0505	Method Blank	05/06/93	-	20 U	25 U	50 U	96.7%
D702 A	9305-003-09	05/06/93	-	41	150	50 U	117%

Surrogate is Me-Arachidate. Values reported in ppm (mg/kg).

GC Data Reporting Qualifiers

- U Indicates compound was analyzed for but not detected at the given detection limit.
- X Indicates a value above the linear range of the detector. Dilution required.
- J Indicates an estimated value when the value is less than the calculated detection limit.
- S Indicates no value reported due to saturation of the detector. Dilution required.
- D Indicates that surrogate was not detected because of dilution of the extract.
- C Indicates a probable value which is unable to be confirmed due to matrix interference.
- NR Indicates no recovery due to matrix interference and/or dilution.
- t Value based on total peaks in the range from Toluene to C12.
- Value based on total peaks in the range from C12 to C24.
- Value based on total peaks in the range from C24 to C32.



ANALYTICAL RESOURCES INCORPORATED

Analytical Chemists & Consultants

TOTAL RANGE HYDROCARBON SPIKE BLANK RECOVERY

333 Ninth Ave. North Seattle, WA 98109-5187 (206) 621-6490 (206) 621-7523 (FAX)

Matrix: Soils/Sediments

ARI Job No: D702

Client: RETEC

Project: 3-1033-777

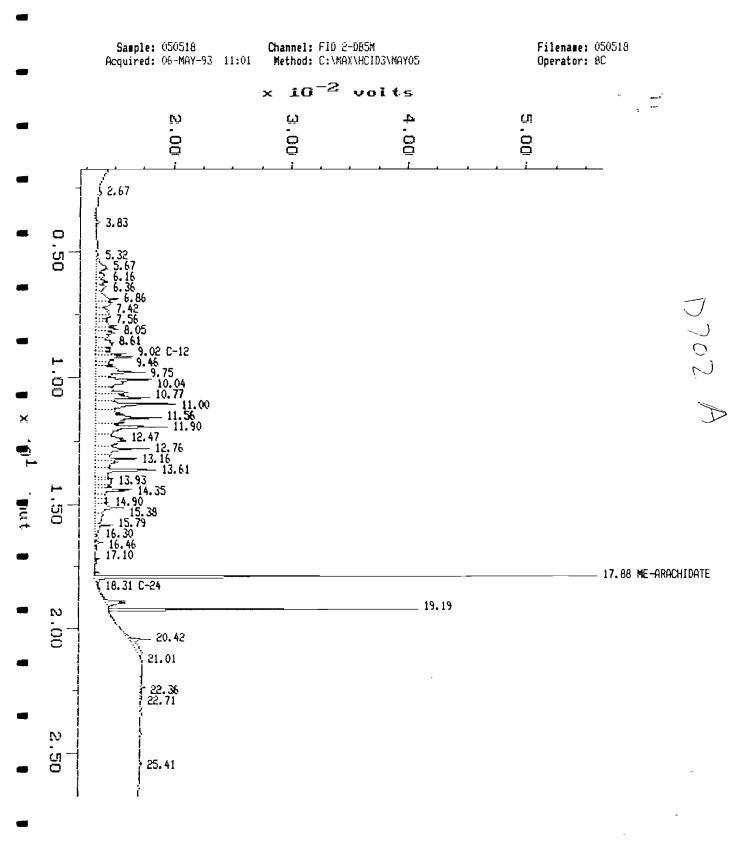
Shore Realty

Date Extracted: 05/05/93 Date Analyzed: 05/06/93

	SPIKE	SB	SB
	ADDED	CONC.	%
COMPOUND	(mg/kg)	(mg/kg)	REC
Diesel	574	659	115%

	Surrogate % rec.
Methyl Arachidate	111

Comments:



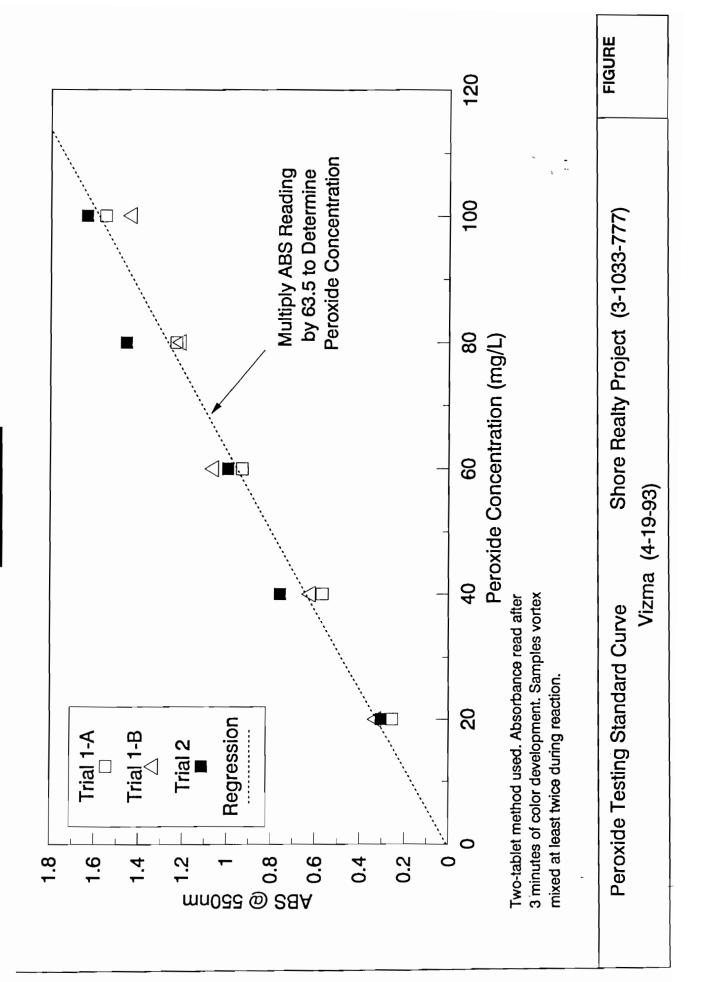


Results of Peroxide Stability Testing

Sample Description	Spike	Time		ample Peroxic	
		(hr)	Sample A	centration (m Sample B	g/L) Mean
			Sample A	Sample D	Wiean
Flasks 1 & 2	Initial	0	384	354	3 69、
B-93-1 (8-10')		1	417	358	387
_		2	359	391	375
]	4	322	490	406
		6	319	336	327
		24	330	279	304
	Repeat	0	624	656	640
	1 - 1	1	644	476	560
		2	645	660	652
		4	761	620	691
		6	483	556	519
		24	548	483	516
Flasks 3 & 4	Initial	0			0
B-93-2 (19-21')		1	489	460	475
2 30 2 (23 22)		2	467	486	477
		4	525	490	507
		6	532	574	553
		24	605	502	554
	Repeat	0	814	823	819
	Repeat	1	711	891	801
		2	823	809	816
		4	568	782	675
]	6	677	700	689
		24	538	475	507
Flasks 5 & 6	Initial	o			
B-93-3 (11-13')		1	442	446	444
•	1 1	2	417	405	411
		4	446	403	424
		6	400	435	418
		24	335	323	329
	Repeat	0	613	620	616
	1 1	1	705	632	668
		2	559	631	595
		4	415	537	476
		6	707	748	727
		24	461	439	45 0
Flasks 7 & 8	Initial	0	422	462	442
Water Controls		1	353	354	354
		2	417	469	443
		4	446	597	522
		6	488	679	584
		24	503	563	533
	Repeat	0	851	930	890
	Repeat	1	907	906	906
		2	721	785	753
		4	855	660	758
			8 5 5	935	895
		6 24	610	662	6 3 6
	II .	7.4	010	OOZ	0.30

Notes:

Peroxide testing performed according to RETEC SOP #735 using slurries with a 10 percent solids loading. Slurries were used for nutrient adsorption testing (RETEC SOP #730 prior to initial spiking with hydrogen peroxide. Target spike concentration was 500 mg/L hydrogen peroxide.



PEROXIDE STABILITY TESTING: SUMMARY

SAMPLE	SAMPLE	Actual	TIME	ABS	Dilution	SAMPLE PEROXIDE
NUMBER		Time	(Hours)	@ 3 min.	Factor	CONC. (ppm)
T=0 Data	Flasks Spiked	12:50 PM	0			
#1	Sample 5/6 A	1:10 PM	0	0.605	10	384.175
	•	2:05 PM	1	0.656	10	416.56
1		3:05 PM	2	0.565	10	358.775
l		5:07 PM	4	0.507	10	321.945
		7:05 PM	6	0.502	10	318.77
		1:10 PM	24	0.519	10	329.565
T=0 Data	Flasks Spiked	2:40 PM	0			
l				 		
#1	Sample 5/6 A	2:55 PM	0	0.983	10	624.205
		3:57 PM	1	1.014	10	643.89
		4:57 PM	2	1.015	10	644.525
1		6:55 PM	4	1.199	10	761.365
		8:57 PM	6	0.76	10	482.6
		3:00 PM	24	0.863	10	548.005
T=0 Data	Flasks Spiked	12:50 PM	0			
						25.22
#2	Sample 5/6 B	1:10 PM	0	0.558	10	354.33
		2:05 PM	1	0.563	10	357.505
		3:05 PM	2	0.616	10	391.16
		5:07 PM	4	0.771	10	489.585
1		7:05 PM	6	0.529	10	335.915
		1:10 PM	24	0.44	10	279.4
T=0 Data	Flasks Spiked	2:40 PM	0			
	2 1 5 5 5	0.55 D) (0	1.022	10	655.955
#2	Sample 5/6 B	2:55 PM	0	1.033	10	
1		3:57 PM	1	0.75	10	476.25 659.765
1		4:57 PM	2	1.039	10 10	620.395
1		6:55 PM	4	0.977 0.876	10	556.26
1		8:57 PM	6	0.876	10	483.235
		3:00 PM	24	0.701	10	403.233
T=0 Data	Flasks Spiked	12:50 PM	0			
1 =0 Data	riasks Spiked	12:30 FM	U			
#3	Sample 11/12 A	1:10 PM	0	0.628	10	398.78
#3	Sample 11/12 A	2:05 PM	1	0.020	10	488.95
		3:05 PM	2	0.736	10	467.36
		5:07 PM	4	0.827	10	525.145
		7:05 PM	6	0.838	10	532.13
		1:10 PM	24	0.953	10	605.155
		1.101.11	24	0.250		000000
T=0 Data	Flasks Spiked	2:40 PM	0			
#3	Sample 11/12 A	2:55 PM	0	1.282	10	814.07
#.3	Sample 11/12 A	2:33 PM 3:57 PM	1	1.12	10	711.2
				1.12	10	822.96
		4:57 PM	2 4	0.894	10	567.69
		6:55 PM 8:57 PM	6	1.066	10	676.91
		3:00 PM	24	0.848	10	538.48
		3.00 PM	24	0.040		5.70.40
L						

T=0 Data	Flasks Spiked	12:50 PM	0			
1 -0 Data	riasks spiked	12:30 FM	U			
#4	Sample 11/12 B	1:10 PM	0	1.64	10	1041.4
	22,222	2:05 PM	1	0.725	10	460.375
		3:05 PM	2	0.765	10	485.775
		5:07 PM	4	0.771	10	489.585
		7:05 PM	6	0.904	10	574.04
		1:10 PM	24	0.791	10	502.285
						· .
T=0 Data	Flasks Spiked	2:40 PM	0			•
#4	Sample 11/12 B	2:55 PM	0	1.296	10	822.96
		3:57 PM	1	1.403	10	890.905
		4:57 PM	2	1.274	10	808.99
		6:55 PM	4	1.232	10	782.32
		8:57 PM	6	1.103	10	700.405
		3:00 PM	24	0.748	10	474.98
T=0 Data	Flasks Spiked	12:50 PM	0			
#5	Sample 17/18 A	1:10 PM	0	0.053	10	33.655
		2:05 PM	1	0.696	10	441.96
		3:05 PM	2	0.657	10	417.195
		5:07 PM	4	0.703	10	446.405
		7:05 PM	6	0.63	10	400.05
		1:10 PM	24	0.528	10	335.28
T=0 Data	Flasks Spiked	2:40 PM	0			
#5	Sample 17/18 A	2:55 PM	0	0.965	10	612.775
"3		3:57 PM	1	1.11	10	704.85
		4:57 PM	2	0.881	10	559.435
		6:55 PM	4	0.654	10	415.29
		8:57 PM	6	1.113	10	706.755
		3:00 PM	24	0.726	10	461.01
T=0 Data	Flasks Spiked	12:50 PM	0			
#6	Sample 17/18 B	1:10 PM	0	0.439	10	
	_	2:05 PM	1	0.703	10	446.405
		3:05 PM	2	0.638	10	405.13
		5:07 PM	4	0.634	10	402.59
		7:05 PM	6	0.685	10	434.975
		1:10 PM	24	0.509	10	323.215
T=0 Data	Flasks Spiked	2:40 PM	0			
#6	Sample 17/18 B	2:55 PM	0	0.976	10	619.76
		3:57 PM	1	0.995	10	631.825
		4:57 PM	2	0.994	10	631.19
		6:55 PM	4	0.845	10	536.575
		8:57 PM	6	1.178	10	748.03
		3:00 PM	24	0.691	10	438.785

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APPENDIX F

Water Pretreatment Study Data Sheets

REMEDIATION TECHNOLOGIES, INC.

MEMORANDUM

TO:

D. Miller

DATE:

July 12, 1993

FROM:

B. Ross 865

RE:

Jar Tests - Shore

Realty (3-1033)

Sample Receipt:

A 5 gallon water sample from Shore Realty was received in RETEC's Pittsburgh Laboratory on June 7, 1993. Upon arrival it was inspected for integrity and then it was logged in. The water sample was used for the laboratory experiments on the same day. Care was taken to keep the water container closed all the time to minimize atmospheric oxidation of the soluble iron in the water sample.

Procedure:

In order to evaluate the oxidant and coagulant dosages, a series of jar tests was performed using samples of the groundwater and various dosages of $KMnO_4$ and coagulant aids. The experimental set up is shown in Figure 1. The apparatus and materials used in the jar test were:

- · six paddle stirrer,
- potassium permanganate (0.1 M solution),
- · cationic polymer (Klar Aid 2400),
- · anionic polymer (Aqua Floc 800),
- base (10% w/w NaOH solution),
- · pH meter,
- stop watch,
- · glassware (pipets, beakers, and graduated cylinders), and
- · test strips for dissolved iron and manganese determination.

The water sample was first analyzed for iron, manganese (using test strips) and pH. The concentration of iron was approximately 25 mg/L and the level of manganese non-detectable. The pH of the water was 6.4 standard units. Two water samples were preserved with acid and sent to Wadsworth/Alert Laboratory (W/A Lab) for the analyses of iron and manganese, an additional sample (as received) was sent for determination of alkalinity. The results indicated approximately 25 to 40 mg/L iron, 2.5 mg/L manganese, and 290 mg/L alkalinity reported as calcium carbonate.

The evaluation of the physical/chemical process proceeded in three steps to determine the optimum dosage of chemicals required for the oxidation and removal of iron in the groundwater.

Step 1 - pH Adjust with Polymer Enhancement

To provide a basis for comparison, a sample of groundwater was first adjusted to pH 8.5 with 10% w/w NaOH solution and various cationic polymer (Klar Aid 2400) dosages were added to enhance gravity separation. Visual observation and test strips measurements for iron determined the optimal dosage to be 1 mg/L cationic polymer. A confirming sample was sent to the W/A Lab for total metals analysis. Results indicated iron at 3.5 mg/L and manganese at 4 mg/L.

Step 2 - Optimization of Potassium Permanganate (KMnO₄) Dosage with Polymer Enhancement

Since the concentration of iron in the water was approximately 25 mg/L, a $0.1~M~KMnO_4$ solution was used. The $0.1~M~KMnO_4$ solution was prepared by dissolving 1.58~g of $KMnO_4$ in 100~mL. Five doses namely, 25, 50, 75, 100, and 125% of stoichiometric requirements were evaluated.

To 500 mL groundwater samples, the calculated quantity of 0.1 M KMnO₄ solution (ranging from 0.4 to 1.0 mL) was added. The sample was rapid mixed (100 rpm) for one minute, slow mixed (20-30 rpm) for ten minutes and then allowed to gravity settle for twenty minutes. The pH was adjusted to within the range of 7.5 to 8.5 with a 10% by weight solution of NaOH. The volumes of KMnO₄ and NaOH added were recorded. The initial and final pH along with iron concentration were measured. The residual concentration of iron after KMnO₄ addition and prior to pH adjustment was also recorded. Test strips were used to determine the concentration of dissolved iron in the water.

The optimal dosage was the one which utilized the minimum chemicals yet oxidized the iron. A dosage of 25% stoichiometric KMnO₄ met these requirements. It was observed that the KMnO₄ reduced the iron concentration from 25 to 10 mg/L and the subsequent pH adjustment reduced the remaining 10 mg/L to non-detectable levels via the test strips. One unfiltered treated sample was sent to the W/A Lab for the analysis of total iron and manganese.

Optimization of Cationic Polymer

In this test, the performance of Klar Aid 2400, a cationic polymer, was evaluated with the combination of previously determined optimal $KMnO_4$ dose. The cationic polymer for this test was tried at 1, 2, 5, 10, and 15 mg/L. The optimal dosage was visually determined by the ability of the chemicals to produce clear supernatant and good settling flocs.

Since the supernatant was very clear, no attempt was made to evaluate the total suspended solids (TSS) in the supernatant. The sample treated with 25% KMnO $_4$ plus 1.0 mL/L weight 10% NaOH plus 5 mg/L Klar Aid 2400 cationic polymer gave the best results in terms of iron oxidation and sludge settling. The volume of dry sludge produced was 691 mg/L. The treated was sent to the W/A Lab for the analysis of total iron and manganese, results indicated approximately 3 mg/L for both metals.

Optimization of a Combination Cationic/Anionic Polymer Enhancement

Quite often, anionic polymers produce only pin flocs when they are used alone without the combination of any other cationic polymer. In order to find whether a lower dose of cationic polymer would give similar result with a combination of an anionic polymer, namely Aqua Floc 408, a test run was made with two samples. Each sample had 25% KMnO₄ in common. Additionally, the first sample had 1.0 mg/L cationic polymer and 1.0 mg/L anionic polymer whereas the second polymer had 2.0 mg/L cationic polymer and 2 mg/L anionic polymer. Both samples had no residual iron and the sludge exhibited good settling characteristics. Therefore it was decided to perform another Jar Test with 25% KMnO₄ and 1.0 mg/L cationic polymer in common with varying dosages of the anionic polymer. In this Jar Test, the optimal anionic dosage was found to be 0.8 mg/L. The volume of dry sludge produced was 902 mg/L. A sample of the unfiltered treated water was sent to W/A Lab for the analysis of total iron and manganese, results were approximately 1 mg/L iron and 3 mg/L manganese.

Comparison of Individual Runs:

From the above three runs, three optimal dosages, one from each run, were selected. Since the three runs were run independently, it was not possible to select one as the best of the three optimal dosages. Therefore, in the final Jar Test, three water samples were taken and they were treated with the three optimal dosages identified previously. At the end of the experiment each sample was analyzed for residual iron. In addition, visual observations were made on floating and settled sludges, turbidity and color. In this final Jar Test, 25% $\rm KMnO_4$ plus 1.0 mL/L 10% w/w NaOH plus 5.0 mg/L cationic polymer gave the best results, in terms of visual observation and test strips.

Summary:

The optimal dosage identified by each run is as follows:

- ph Adjust to 8.5 (approx. 1.4 mL/L) 10% w/w NaOH plus 1 mg/L cationic polymer (Klar Aid 2400),
- 25% stoichiometric KMnO₄ plus 1.0 mL/L 10% w/w NaOH plus 5.0 mg/L cationic polymer (Klar Aid),
- 25% stoichiometric KMnO₄ plus 1.0 mL/L 10% w/w NaOH plus 1.0 mg/L cationic polymer (Klar Aid 2400) plus 0.8 mg/L anionic polymer (Aqua Floc 408).

It was observed that all the optimal dosages performed in an equivalent manner regarding the ability to oxidize dissolved iron present in the water samples. The sludge formation in all the three samples was good. However, the pH adjusted sample followed by cationic polymer addition produced the most suspension with a dry sludge production of 71 mg/L. The sample treated with 25% stoichiometric KMnO₄ followed by 1.0 mL/L 10% w/w NaOH plus 5.0 mg/L cationic polymer exhibited minimal floaters. The dry sludge produced was 691 mg/L. The 25% KMnO₄, NaOH, cationic polymer, and anionic polymer produced the most floaters with a dry sludge production of 902 mg/L.

Discussion of Results:

The analytical results are summarized in Table 1, copies of the W/A Lab reports are provided as Attachment A, copies of the treatability laboratory data sheets are provided as Attachment B. The results for the untreated water show total iron ranged from 23 to 38 mg/L. The variation in iron concentration could be due to the non homogeneity of the water samples. Such variations occur whenever iron is present in both soluble and insoluble forms. These results, however, compare fairly well with the test strip result of 25 mg/L. The concentration of manganese was 2.4 mg/L for the two samples and increased slightly in the treated samples, indicating no change in manganese levels. The test strips did not identify the presence of manganese. The results show that the treated water small amounts (less than 5 mg/L) of iron and manganese. The differences in the concentration among the samples is not appreciable. The results indicate that the a dosage slightly higher than the 25% of stoichiometric may be required to fully oxidize the iron and manganese, however, since only total metals analysis was performed, it can not be conclusively stated whether the chemical dosage was insufficient or the physical gravity separation was responsible for residual iron and manganese concentration (i.e. whether the metals are present in soluble or insoluble form).

Figure 1

Test Apparatus

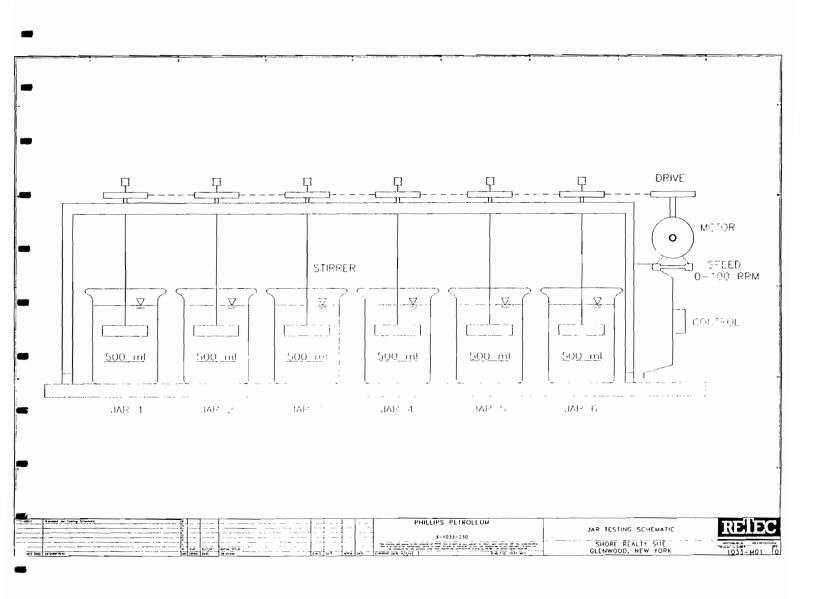


Table 1

Summary of Analytical Results

Description	Total Iron (mg/L)	Total Manganese (mg/L)	Alkalinity as mg/L CaCO ₃	
Initial				
Sample 1	38.0	2.4	Not Applicable	
Sample 2	23.0	2.4	Not Applicable	
As Recieved	Not Analyzed	Not Analyzed	290	
pH Adjusted with Cationic Polymer Enhancement	3.4	4.1	Not Applicable	
25% KMnO ₄ with 5 mg/L Cationic Polymer	2.8	3.6	Not Applicable	
25% KMnO ₄ with 1 mg/L Cationic and 0.8 mg/L Anionic Polymer	0.9	2.7	Not Applicable	

ATTACHMENT A

Wadsworth/Alert Laboratory Reports



WADSWORTH/ALERT Laboratories

Division of Enseco Incorporated

450 William Pitt Way Pittsburgh, PA 15238 412-826-5477 FAX 412-826-5571

ANALYTICAL REPORT

PROJECT NUMBER 3-1033-250

ENSECO-WADSWORTH/ALERT PROJECT NUMBER 1252

Presented to :

Robin Weightman

REMEDIATION TECHNOLOGIES INC.

ENSECO-WADSWORTH/ALERT LABORATORIES

Thomas Tomayko

Project Manager

Renee' Gigliotti

Quality Assessment Group Leader - Pittsburgh

June 30, 1993

NARRATIVE

The following report contains the analytical results for samples submitted to ENSECO-Wadsworth/ALERT Laboratories. The samples were received into the laboratory in accordance with documented sample acceptance procedures.

ENSECO-Wadsworth/ALERT Laboratories utilizes USEPA approved methods and instrumentation in all analytical work. The methods used for the analyses presented in this study can be found on the following pages.

The following codes are utilized in various analyses of this report:

- ND (None Detected)
 - J (Detected, but below quantitation limit; estimated value)
 - B (Compounds detected in method blank associated with this sample)
- DIL (Diluted Out)
 - MI (Matrix Interference)

ANALYTICAL METHODS

ENSECO-Wadsworth/ALERT Laboratories utilizes only USEPA approved analytical methods and instrumentation. The analytical methods used in the analyses of these samples are listed below.

<u>Parameters</u>	<u>Methods</u>
Matala.	
Metals:	
Iron	EPA 200.7
Manganese	EPA 200.7

EPA: Methods for Chemical Analysis of Water and Wastes EPA 600/4-79-020, March 1983.

DATE RECEIVED: 6/10/93

LAB #: 1252-81873 MATRIX : WATER

SAMPLE ID : INTIAL A 6-9-93

METALS ANALYTICAL REPORT SELECTED LIST Torse Merres

Total metals analysis results - as received

	PREPARATION -		DETECTION		
ELEMENT	ANALYSIS DATE	RESULT	LIMIT		
Iron	6/17- 6/21/93	38000	100	ug/l	
Manganese	6/17- 6/21/93	2400	15	ug/l	

DATE RECEIVED: 6/10/93

LAB #: 1252-81874 MATRIX : WATER

SAMPLE ID : INTIAL B 6-9-93

Torac Merais

METALS ANALYTICAL REPORT SELECTED LIST

Total metals analysis results - as received

ELEMENT	PREPARATION - ANALYSIS DATE	RESULT	DETECTION LIMIT
Iron	6/17- 6/21/93	23000	100 ug/l
Manganese	6/17- 6/21/93	2400	15 ug/l

DATE RECEIVED: 6/10/93

LAB #: 1252-81877 MATRIX : WATER

SAMPLE ID : NAOH 1MG/L 6-9-93

PH adjust to E.4 and Imgle Cationic Total Meracs

METALS ANALYTICAL REPORT SELECTED LIST

Total metals analysis results - as received

ELEMENT	PREPARATION - ANALYSIS DATE	RESULT	DETECTION LIMIT
Iron	6/17- 6/21/93	3400	100 ug/l
Manganese	6/17- 6/21/93	4100	15 ug/l

DATE RECEIVED: 6/10/93

LAB #: 1252-81876 MATRIX : WATER

SAMPLE ID: 25% LU 04 5 MG/L 6-9-93 25% KMmO4 5mg/L cationic

SELECTED LIST

METALS ANALYTICAL REPORT TOTAL PROTECTION

Total metals analysis results - as received

	PREPARATION -		DETECTION	
ELEMENT	ANALYSIS DATE	RESULT	LIMIT	
Iron	6/17- 6/21/93	2800	100	ug/l
Manganese	6/17- 6/21/93	3600	15	ug/l

DATE RECEIVED: 6/10/93

LAB #: 1252-81875 MATRIX : WATER

SAMPLE ID: 25% LU 04 1 MG/L 6-9-93

25% KM_O4 Color long/Locations

METALS ANALYTICAL REPORT

SELECTED LIST

Total Metals

Total metals analysis results - as received

	PREPARATION -		DETECTION	
ELEMENT	ANALYSIS DATE	RESULT	LIMIT	
Iron	6/17- 6/21/93	900	100	ug/l
Manganese	6/17- 6/21/93	2700	15	ug/l

Nº 0003

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WADSWORTH/ALERT Laboratories

Division of Enseco Incorporated

450 William Pitt Way Pittsburgh, PA 15238 412-826-5477 FAX 412-826-5571

ANALYTICAL REPORT

PROJECT NUMBER 3-1033-250

ENSECO-WADSWORTH/ALERT PROJECT NUMBER 1262

Presented to :

Robin Weightman

REMEDIATION TECHNOLOGIES INC.

ENSECO-WADSWORTH/ALERT LABORATORIES

Thomas Tomayko
Project Manager

Renee' Gigliotti
Quality Assessment Group Leader - Pittsburgh

June 17, 1993

NARRATIVE

The following report contains the analytical results for samples submitted to ENSECO-Wadsworth/ALERT Laboratories. The samples were received into the laboratory in accordance with documented sample acceptance procedures.

ENSECO-Wadsworth/ALERT Laboratories utilizes USEPA approved methods and instrumentation in all analytical work. The methods used for the analyses presented in this study can be found on the following pages.

The following codes are utilized in various analyses of this report:

- ND (None Detected)
 - J (Detected, but below quantitation limit; estimated value)
 - B (Compounds detected in method blank associated with this sample)
- DIL (Diluted Out)
 - MI (Matrix Interference)

ANALYTICAL METHODS

ENSECO-Wadsworth/ALERT Laboratories utilizes only USEPA approved analytical methods and instrumentation. The analytical methods used in the analyses of these samples are listed below.

<u>Parameters</u> <u>Methods</u>

Inorganics:

Alkalinity EPA 310.1

EPA: Methods for Chemical Analysis of Water and Wastes EPA 600/4-79-020, March 1983.

COMPANY: REMEDIATION TECHNOLOGIES INC. DATE RECEIVED: 6/11/93

LAB #: 1262-81993 MATRIX : WATER

SAMPLE ID : INITIAL 6-11-93 10:00

ANALYTICAL REPORT

PREPARATION - DETECTION
ANALYSIS DATE RESULT LIMIT

Alkalinity (CaCO3 to pH 4.5) 6/14/93 290 1.0 mg/l

0004 Ž

CHAIN OF CUSTODY RECORD

Destination: WADS

830028

P.O. No._

PROJ. NO.	PROJEC	PROJECT NAME: (, ,) ,											Γ
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3-1022-250 PROJECTMANAGER:	PROJEC	CALSO T RODIL WEIGHT											
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Relinqueshed by: (Signature)		Date / Time Received for Laboratory by: (Engnature)	gnature)	6-11-83	00:01			<u> </u>	<u>α</u>	EMEDIA 3040 V	TION TI	REMEDIATION TECHNOLOGIES	GIES
REMARKS:						r	EMED	1 A 7 1 0	■ z	PITT	SBURGH	PITTSBURGH, PA 15238	•
						Ĭ	CHNOL	TECHNOLOGIES INC	: <u>-</u> ;	Ā	(412) 826-3340 FAX (412) 826-3409	-3340	

PINK - RETEC

ATTACHMENT B

Treatability Laboratory Data Sheets

1 2 8

JAR TESTING CATIONIC POLYMER SCREENING

PH adjusted to 8.4 No KM1009 + Cationic Polymer

	ionio i ocimen donce inica	t Cationic Iclym
Project Name: Shore Reglity	Date:	6 9-93
Project Name: Shope Reglity Project No.: 3-1033-220	Operator:	ALK/BSR
Project Mgr.: REK / DLM		/
Cationic Polymer: KhARA	toichiometric dose = mg/500ml =	=O ml KMnO4
Best Dose Rivino4=	tolemometric dose = mg/300im =	= mi
1. Initial Screening		
1. Addml KMnO4 to one 50	Oml sample, rapid mix 1 minute.	
2. Add cationic polymer in incremental dose	es as indicated below, noting visual observations w	//each addition.
3. Continue rapid mixing.		
1	20 - 0 - 1 - 1 - 1 - 1 - 1 - 1 - 1 - 1 -	Tarkishanada
1 mg/L = 0.05 ml polymer/500ml (0.05ml) 5 mg/L = 0.25 ml polymer/500ml (+0.2ml)	20mg/L = 1 ml polymer/500ml (+0.5ml)	Try higher or in
10 mg/L = 0.5 ml polymer/500ml (+0.25 ml)	50 mg/L = 2.5 ml polymer/ 500 ml (+1.5ml) 100 mg/L = 5 ml polymer/ 500 ml (+2.5ml)	between doses if necessary
10mg E = 0.2 m polymoly300m (10.23m)	room gr = 5 mr potymen, secum (· 25 mr)	n necessary
Best cationic polymer dose:		
	 _	
2. Standard Cationic Screening		
Repeat steps 1-3 above using separate beak	ers for each polymer dose (0X to 2X the best dos	****
		ual Observations
Bkr 1: 0 ml cationic polymer/500ml	,	
Bkr 2: ml cationic polymer/500ml	= 1 mg/L very clock no float.	is solids
	= 2 mg/L 1% of pin flow still so	
Bkr 4:	= 5 mg/L yerry werry classe no so = 10 mg/L ge % settled in a min no	13 Styles Solies
Bkr 6: ml cationic polymer/500ml		
- Dai o un cacionic polymen/200m -	mgr /y o seriected King	No Sunsacer Solicio
Best cationic polymer dose:		
Best cationic polymer dose:		
NOTES: No Khnoy used.	small floc Indiately After entration After ph adjustmen	ph Adjustment to
- Inin Rapid mix 10min slow	nix 20 min. settling	
avel & che called in	and a leaty After 100 in store	
SKI 40- FICE SELLING	of all	·
BKQ-4- WERY Small Pint	floc suspended After 4mins	setting.
		.•
BKZ = Best dosAge	visually , orm of catonie ((my/L)
	10% Nath added to	•

JAR TESTING

JAR TESTING KMnO4 ADDITION

Toject Hanie.	3 Yoke	Reglity		- Date.	6-9-9	S	_
Project No.:		- 1033 -26	0	Operator:	ALK /BS	R	_
Project Mgr.:	REK/D	LM		_	/		
		_(if < 6, increas	e to ~7 using 1 mls of 10% Na	.0% NaOH) OH/10% H2SC	Adjuste.	s soon	then ndivi
. Fe/Mn Content			-			Deskers.	
Fe content of compos	ite:	25	t As	u.g/500ml			
Mn content of compo	site:			mg/500ml			
6. KMnO4 Addition 6.1 ml of 0.1 molar Stoichiometric Dose of the stoichiometric dose	– KMnO4 soluti of 0.1M KMnC)4 equals the sto	ichiometric dos	e for Fe (1mg F	Fe: 1mg KMnC)4) plus	. •
					,		
3kr 1: 0% of stoichio 3kr 2: 15% of stoichi			mg/500ml= mg/500ml=		ml KMnO4 ml KMnO4		
3kr 3: 25% of stoichi			mg/500ml=		ml KMnO4		- 11
Bkr 4: 50% of stoichi			mg/500ml=		ml KMnO4	4.8 mL AK	HOs
Bkr 5: 100% of stoich	iometric dose	=	mg/500ml=		ml KMnO4		
01 (1050 5							
Bkr 6: 125% of stoich After addition, rapid i solution used to adjus	mix for 5 minu		mg/500ml= <6, note amou		ml KMnO4 ic		
After addition, rapid is solution used to adjus	mix for 5 minu t pH.	tes, adjust pH if	<6, note amou	nt of acid/caust	ic	Beaker 6]
After addition, rapid is colution used to adjust the After Rapid Mixing	mix for 5 minu t pH.		<6, note amou			Beaker 6 125%	
After addition, rapid is colution used to adjust the After Rapid Mixing	mix for 5 minu t pH. : Beaker 1	tes, adjust pH if	<6, note amou	nt of acid/caust	Beaker 5		
After addition, rapid is obtained to adjust the Rapid Mixing Check for:	mix for 5 minu t pH. : Beaker 1	tes, adjust pH if	<6, note amou	nt of acid/caust	Beaker 5		
After addition, rapid rolution used to adjust After Rapid Mixing Check for: pH Fe Mn	mix for 5 minut t pH. Beaker 1 Dose: 0%	tes, adjust pH if	<6, note amou	nt of acid/caust	Beaker 5		
After addition, rapid roll to adjust the following of the	mix for 5 minute t pH. Beaker 1 Dose: 0%	Beaker 2	<6, note amou Beaker 3 25%	Beaker 4	Beaker 5		
After addition, rapid resolution used to adjust to adjust the After Rapid Mixing Check for: PH Fe Mn Visual Observations	mix for 5 minute t pH. Beaker 1 Dose: 0%	tes, adjust pH if	<6, note amou Beaker 3 25%	Beaker 4	Beaker 5		
After addition, rapid resolution used to adjust to adjust the After Rapid Mixing Check for: PH Fe Mn Visual Observations Bkr 1	mix for 5 minut t pH. Beaker 1 Dose: 0%	Beaker 2	<6, note amou Beaker 3 25%	Beaker 4	Beaker 5		
After addition, rapid resolution used to adjust to adjust the After Rapid Mixing Check for: PH Fe Mn Visual Observations Bkr 1	mix for 5 minut t pH. Beaker 1 Dose: 0%	Beaker 2	<6, note amou Beaker 3 25%	Beaker 4	Beaker 5		
After addition, rapid rolution used to adjust After Rapid Mixing Check for: PH Fe Mn Visual Observations Bkr 1 Bkr 2	mix for 5 minut t pH. Beaker 1 Dose: 0%	Beaker 2	<6, note amou Beaker 3 25%	Beaker 4 50%	Beaker 5		
After addition, rapid rolution used to adjus After Rapid Mixing Check for: pH Fe Mn Visual Observations Bkr 1 Bkr 2 Bkr 3	mix for 5 minut t pH. Beaker 1 Dose: 0%	Beaker 2	Seaker 3	Beaker 4 50%	Beaker 5 100%		
After addition, rapid resolution used to adjust the After Rapid Mixing Check for: PH Fe Mn Visual Observations Bkr 1 Bkr 2 Bkr 3 Bkr 4	mix for 5 minut t pH. Beaker 1 Dose: 0%	Beaker 2	Seaker 3 25%	Beaker 4 50%	Beaker 5 100%		
After addition, rapid resolution used to adjust the After Rapid Mixing Check for: PH Fe Mn Visual Observations Bkr 1 Bkr 2 Bkr 3 Bkr 4	mix for 5 minut t pH. Beaker 1 Dose: 0%	Beaker 2	Seaker 3 25%	Beaker 4 50%	Beaker 5 100%	125%	
After addition, rapid resolution used to adjust the After Rapid Mixing Check for: PH Fe Mn Visual Observations Bkr 1 Bkr 2 Bkr 3 Bkr 4 Bkr 5	mix for 5 minut t pH. Beaker 1 Dose: 0%	Beaker 2	Seaker 3 25%	Beaker 4 50%	Beaker 5 100%	125%	

ph Adjusted Atter Kingy Dosnge. (xx) coll be due to

excess Kmno y.

JAR TESTING KMnO4 ADDITION

Project Name: S	Shoke Rr	g lity		Date:	6-9-9 ALR / B.	3	
Project No.:	3-1033	-2501		Operator:	ALR / B.	JR	
Project Mgr.:	RSK 10L	m		•			
1. pH of Composite Initial pH:		f < 6, increase		-		``, ±	
Adjusted pH:	,		mls of 10% Na	OH/10% H2SC	4 used.		
2. Fe/Mn Content							
Fe content of composite:	ھ	<u> 5</u>		mg/500ml			
Mn content of composite:	:			mg/500ml			
3. KMnO4 Addition		15 0 17	14-04				
* 1 ml of 0.1 molar KM Stoichiometric Dose of 0.				e for Ee (1 mg E	a · 1ma KMnO	4) alus	-
the stoichiometric dose for		•		e for re (Img r	e : Img AvinO	4) plus	
the stolehometric dose to	or term (Thig ter	in . Zuig Kiviu					
Bkr 1: 0% of stoichiomet	tric dose =		mg/500ml=	0	ml KMnO4		
Bkr 2: 45% of stoichiome	_		mg/500ml=	<u>~</u>		. Ynl of A	ALI S 5
Bkr 3: 25% of stoichiome			mg/500ml=		_	(1)	Aささ 8 2
Bkr 4: -50% of stoichiome			mg/500ml=			7 0 1 1 1	01 80
Bkr 5: 100% of stoichiom			mg/500ml=		1101 01	CI AFRA	0.5
Bkr 6:425% of stoichiom	etric dose =		mg/500ml=	1.0	ml KMnO4	142/05 NA	lor 8 -
After addition, rapid mix	for 5 minutes,	adjust pH if	<6, note amou		ic		
solution used to adjust pH	ł.						
After Rapid Mixing:							1
	1	Beaker 2	Beaker 3 50	Beaker 4		Beaker 6	
Check for: Do	ose: 0%	<u>a</u> 5%	25%	\$0%	100%	125%	
pH	6.4	6.5	6,6	6.6	6.6	66_	
Fe Fe	25	10	0	0	C		C C.
				0		0	100
Visual Observations:	25	0	0	O,	0	.◄	~ 10° WHM 20°
Bkr 1 <i>A</i>	4 100 3 KIN	, 5	ਰ	- Cloud	ly othin	20 mb	5 . 4
—	settlim	χ		<u> </u>	<u> </u>		
Bkr 2		44	clian.				
Bkr 3 C	lear w	it 1-2	. hum	floor flo	atm.		
			7	P	0		
Bkr 4	Jer 6050 -	Fluc set	Hed immed	ately ASI	K Bh Adis	trent sna	11 Amount
	f floating So	leds - Su	- De en tanner	West one	Mily con		
Bkr 5 0	of dos-	Same	an Bkn	4 - Swa	matani	shighth	4
-	colwed.			' 1		<u> </u>	1
Bkr 6	- 64 403 4 -	Floc sell	led after to	Hadj. Sm	all goutile	d that	y notido.
	permaka	<u> L – Blig</u>	ritly colo	red 1		1 0 1	•
NOTES: 25%	KMn04	`	,				
A N.	> R.	Channel M	b be	of settle	ing. No	Quoten	lead
- TYD	C 150	THOSE ()	מאאוז בי		()	7	
11						. 1	Adlin 200

1 n3 Km 044 KIAKH.D 2400	DEST MONGE
	JAR TESTING 25% KMnOx
CATION	JAR TESTING JAR TESTING 25% KMDO4 +5mg/L CATIONIC POLYMER SCREENING
Project Name: Share Reglity	
Project Name: Share Reg 1,4y Project No.: 3-1033-250	Date: <u>6.8-83</u> Operator: <u>ACR/BJ</u> R
Project Mgr.: R&K / DL m	
Cationic Polymer: KIAR A.D	2460 Floor Stitled al-The centre as part-
Best Dose KMnO4= 25 % stoich	hiometric dose =mg/500ml =ml KMnO4
 Initial Screening Add	sample, rapid mix 1 minute. s indicated below, noting visual observations w/each addition.
	20 mg/L = 1 ml polymer/ 500 ml (+0.5 ml) Try higher or in
	60 mg/L = 2.5 ml polymer/ 500 ml (+1.5ml) between doses $00 mg/L = 5 ml polymer/ 500 ml (+2.5ml)$ if necessary
	the decessary
Best cationic polymer dose:	
2. Standard Cationic Screening	
Repeat steps 1-3 above using separate beakers f	for each polymer dose (0X to 2X the best dose as determined above). Visual Observations
Bkr 1: 0 ml cationic polymer/500ml =	0 mg/L (control)
	mg/L
Bkr 3:/ ml cationic polymer/500ml = Bkr 4:/2\(\) ml cationic polymer/500ml =	
Bkr 5: sc ml cationic polymer/500ml =	
Bkr 6: > 5 ml cationic polymer/500ml =	15 mg/L
Best cationic polymer dose: Klar A.D.	2400 25ml or smill added Fremlind on North
reafler pH	ml Adj. 1911
NOTES: KMnog +Col	
• 0 0 — 6.2	2 Settledwell, But Small & Settledwell, But Small & Similar Single Constant
3 06.	
- (5) - (4) (4) (6)	as pui floes in the supernatural supernatural supernatural schled Clean
- (5) O	Supernatant. Settled will. A. T. Stylen in the supernet.
- 6 0 <u>a</u>	.3 - 5. 5 Same c. 6

JAR TESTING ANIONIC POLYMER SCREENING

Project Name:	Share Roglity	
Project No.:	3-1033-250	_
Project Mgr.: _	REK/ DLM	

Date: 6-9-93Operator: ACR/BSR

Cationic Polymer: KIALAID 2400
Anionic Polymer: Aaa floc 408

■Best Dose Cationic Polymer = ______mg/L

- 1. Following instructions in work request, add KMnO4, adjust pH and add best determined dose of cationic polymer
- to each of five beakers. Beaker 1 is a control and should NOT have any additions.
- 2. Rapid mix 1 minute, note visual observations.
- 3. Add anionic polymer as instructed in work request, rapid mix 1 minute, slow mix @30rpm for 10 minutes.
- 4. Let settle for 30 minutes, note visual characteristics. Check pH, Fe and Mn content of each sample.

				Visual Observations (while slow mixing)
	Bkr 1: 0 ml anionic polymer/500 ml =	0	mg/L	
	Bkr 2: 0.25 • ml anionic polymer/500 ml =	0.5	mg/L	
-	Bkr 3: 0.5 ml anionic polymer/500 ml =	1.0	mg/L	
	Bkr 4: 1.0 ml anionic polymer/500 ml =	2.0	mg/L	
-	Bkr 5: 2.5 / ml anionic polymer/500 ml =	5.0	mg/L	
	Bkr 6: 5.0 / ml anionic polymer/500 ml =	10	mg/L	

			рН	Fe_	Mn	Visual Observations after Settling
–	Beaker 1	CONTR	.0L			cloudy - No change.
	Beaker 2	25/K	m04+ Img	cat +	==== 1.0m	rschminic - Very clear
	Beaker 3	as/ KT	nneg + 2n	سيار حما-	+2mg/L	Anionic Very Clear
	Beaker 4				\ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \	
	Beaker 5					
	Beaker 6					

Best Anionic Polymer Dose: Accepte 408 soul of longle

NOTES: Large clump of Solid on Bottom After 7.1/2 nm slow nix, supermetent

. 15

25 10 nmnoy 70.201 Ing/c cqt 7005mi

JAR TESTING ANIONIC POLYMER SCREENING

Project Name: Shore Reglity Project No.: 2-1033-250 Project Mgr.:	Date: 6.9-83 Operator: ALX/BSR
Cationic Polymer: Klaraio 2400 Anionic Polymer: Aque Fluc 408	
Best Dose KMnO4=	mg/500ml = <u>O-Q</u> ml 0.1M KMnO4 mg/L

- 1. Following instructions in work request, add KMnO4, adjust pH and add best determined dose of cationic polymer
- to each of five beakers. Beaker 1 is a control and should NOT have any additions.
- 2. Rapid mix 1 minute, note visual observations.
- 3. Add anionic polymer as instructed in work request, rapid mix 1 minute, slow mix @30rpm for 10 minutes.
- 4. Let settle for 30 minutes, note visual characteristics. Check pH, Fe and Mn content of each sample.

ſ				Visual Observations (while slow mixing)	
٦	Bkr 1:	0 ml anionic polymer/500 ml = 0 r	mg/L	snell 2. Aflac	
1	Bkr 2:	0.25, 2 ml anionic polymer/500 ml = 0.5,3 m	mg/L	smalls, a floc close suspended solids a Muslita	,\
			I .	Smallon flow etro-wayded class.	
	Bkr 4:	1.0.7 ml anionic polymer/500 ml = 2.8/8	mg/L	clase no più Plac no Fleating Sussendade	و ک
				class no sus fluc no flexion	
	Bkr 6:	5.8.6 ml anionic polymer/500 ml = 18/12.00	mg/L_	Clerk no pin flow no check flowing	//

	рН	Fe	Mn	Visual Observations after Settling
Beaker 1	8-2	Q		
Beaker 2	8.2	0		
Beaker 3	8.2	Ò		
Beaker 4	8.2	Ò		
Beaker 5	8 2			
Beaker 6	8.2	\bigcirc		

Best Anionic Polymer Dose: A QUA Floc 408	1 3ml OR	16mg/L
—		- ,
NOTES: BOKKER J. Best Dosage.	.4n1 or	· 0 mg/ c

Send simples 5 scaples Alkanity 2-3-+ 1 - Brist Polyard 5 Shore Reality 1 - B-st Kuncy - Aest Knnoy + Best Polyno Received 5581 Sample Instial Fe 25ppm Final - 0:3ppm Initial Ma Oppm Ph. - 6. 4 Adjusted Pin - 8.4 #8.1s of 1.00 NAOH After ph was Adjusted . Ikon existred and fello-t. sample (wis no good. Fe fellout A+ Ph between 7.2 -8.5 nade A olon solution of Km cy (10mlofin solution / 90ml 450) 1 No Kung Added Just KLARRED 2400 (8 force

902mg) 690 mg/L 25% Kmnoy 1.0mg/L 10/120H small cation 0.345HB 500m salmyle I is 1. H 600 1146 0.03579 1) mg/L 500 ml Cationic 1.1488 Img / r. Na O H Tritial weight eludge broduced Studge produced Find wight After duming Waler

SLUDGE PRUDUCTION.

	Machtham & 8.4 Nachtham Img/Lear (0.05ml)	\$5% KMnU4 (8.cm) + 0.5ml NaOH & + 5mg/L Cal (0.25ml)	40.5ml March +0.5ml March +1mglecat (0.05 +0.8mgle Anioni (0.4mg)
Fo bailed map	25	28	25.
صيدع	4		
ox math.	Ö.3	p. 3	6.3
Nach ml	0.1		
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Naloh		0.0	٥٠٥
TO	& ⊗	8.2	8.2.
	Floe furnation good. This/ Air (100. 8th)! Breams dum 1000	Floes surprided during of the minimus of the minimus of the office of the minimus.	Floe furnation group good. Mat of the floca settle of durain
laising roundered	met pusperoins	lesst thereof	most floore

	, i	1.0 m/L.		190g	Volume of water nounter 3,500 ml Switzed Pre SpinkMrv Orgal- Initial pre 8.4	H solution }	Finish forth Finish Fe Min it O. myll.
1044 Number 3-1033 - 250 Contract Cours Cours Cours 6-9-93	Strick Concentration with a superior of the contraction of the contrac	And	Anonic Totaline (4) Anonic Totaline (5) mg/L (5 mg/L 5 mg/L 5 mg/L 15	A Strong Concertation Polymer Name	Non-jonic Polymer Dosage (wat KMnO @	Spip ton Concerning Popmer Name Optimum Dosage:	Strickionatic Combination Polymer Dosage with Kino @ 22 / 4 x - Carbonic Polymer at 1 moly with Kino @ 22 / 4 x - Carbonic Polymer at 1 moly Strip Line Concernation Strip Line Concernation Strip Line Concernation Strip Line Charles

APPENDIX G

Soil Analytical Data

PRE-REMEDIAL DESIGN INVESTIGATION REPORT

for the

Shore Realty Superfund Site Glenwood Landing, New York

Volume II Appendix G

Copy of Appendix G is maintained in both NYSDEC's and RETEC's files.

APPENDIX H

Groundwater Analytical Data

PRE-REMEDIAL DESIGN INVESTIGATION REPORT

for the

Shore Realty Superfund Site Glenwood Landing, New York

Volume III Appendix H

Copy of Appendix H is maintained in both NYSDEC's and RETEC's files.