

**GROUNDWATER MONITORING PROGRAM
2019 SEMI-ANNUAL (DECEMBER) STATUS REPORT
FOR
REMEDIAL ACTION AT
THE DEFENSE NATIONAL STOCKPILE CENTER SCOTIA
DEPOT
GLENVILLE, NEW YORK**

Prepared For:



U.S. Army Corps of Engineers

Prepared By:



AECOM Technical Services

March 2020

GROUNDWATER MONITORING PROGRAM

2019 QUARTER FOUR STATUS REPORT

FOR

REMEDIAL ACTION AT

THE DEFENSE NATIONAL STOCKPILE CENTER SCOTIA

DEPOT

GLENVILLE, NEW YORK

Prepared For:

U.S. Army Corp of Engineers

Prepared By:

AECOM

Contract No. W912DY-09-D-0059

Task Order No. 0010

March 2020

Table of Contents

1	INTRODUCTION.....	1-1
1.1	Site Description.....	1-1
1.2	Site History	1-1
1.2.1	Summary of Previous Investigations	1-2
1.2.2	Pre-Design Groundwater Investigation – 2013.....	1-3
1.2.3	Baseline Groundwater Investigation.....	1-3
1.3	PRB Design Summary	1-4
1.4	Remedial Action Implementation	1-4
2	GROUNDWATER MONITORING PROGRAM.....	2-1
2.1	Sample Collection Methods.....	2-1
3	RESULTS	3-1
3.1	Hydrogeologic Results.....	3-1
3.2	Groundwater MNA and Field Parameter Results	3-2
3.3	Groundwater VOC Results	3-4
4	SUMMARY AND CONCLUSIONS	4-1
5	REFERENCES.....	5-1

List of Figures

Figure 1-1 Site Location Map

Figure 1-2 Site Layout Map

Figure 2-1A Compliance Wells and PRB Profile – December 2019

Figure 2-1B Groundwater Profile Section B-B' – December 2019

Figure 3-1 Potentiometric Site Map – December 2019

Figure 3-2 Groundwater Results – Quarterly Monitoring Locations – December 2019

List of Tables

Table 2-1	Monitoring Well Sampling Schedule and Guidelines
Table 3-1	Groundwater Elevation Data – December 2019
Table 3-2	Compliance Well Pairs Groundwater Sample Results - December 2019

List of Appendices

Appendix A	Groundwater Sample Collection Field Forms
Appendix B	Field Calibration Forms
Appendix C	Hydraulic Gradient and Velocity Calculations
Appendix D	Full Laboratory Analytical Results
Appendix E	AECOM Data Usability Summary Report (DUSR)
Appendix F	Groundwater Concentration Trend Plots
Appendix G	Mann-Kendall Trend Analysis Results

1 INTRODUCTION

This report has been prepared by AECOM on behalf of the United States Army Corps of Engineers (USACE) and the United States General Services Administration (GSA) to document the groundwater monitoring activities performed at the Former Scotia Navy Depot (FSND) (Site) for the semi-annual period between June 22, 2019 and December 12, 2019. This report presents the results of the eleventh groundwater sampling event after the completion of the construction of the zero valent iron (ZVI) permeable reactive barrier (PRB) which was installed across the volatile organic compound (VOC) plume to remediate groundwater at the Site. The first eight sampling events were conducted on a quarterly basis. Starting with the ninth event, the schedule has changed to a semi-annual basis (December and June) in accordance with the Site Management Plan (SMP) (AECOM, 2017b).

This groundwater sampling event included collection of groundwater samples from 12 monitoring wells. Installation of the PRB was completed from February 2016 to December 2016. The Site is adjacent to the north side of New York State (NYS) Route 5 (Amsterdam Road) in the Town of Glenville, Schenectady County, New York. A Site location map is provided in Figure 1-1.

1.1 Site Description

The Site and adjacent properties are zoned for commercial use. Residential properties are located to the south between Amsterdam Road and the Mohawk River. The Mohawk River is located approximately 1,500 feet west-southwest of the Site and represents the major drainage feature in Schenectady County. The water table beneath the Site is approximately 65 feet below ground surface (bgs), and groundwater beneath the Site flows from northeast to southwest toward the Mohawk River.

The Site overlies a United States Environmental Protection Agency (US EPA) designated Sole Source Aquifer referred to as the Schenectady or Great Flats Aquifer system, which is adjacent to and extends beneath the Mohawk River over a distance of approximately 12 miles in Schenectady County. Relative to a series of four aquifer protection zones established to protect five municipal water supplies relying on the aquifer system, the Site lies in Zone III or the General Aquifer Recharge Area. The Site is located approximately 1,500 feet southwest of the Village of Scotia well field and approximately 1.25 miles north of the Town of Rotterdam and City of Schenectady well fields.

Portions of the original Scotia Naval Depot have been subdivided and sold since 1972 by the United States Government. The Site now consists of several large privately held parcels in addition to a portion of land still administered by the GSA. The private parcels contain a variety of industrial tenants; while the GSA leases its remaining portion to the Defense Logistics Agency/Defense National Stockpile Center and the Navy.

1.2 Site History

The Scotia Depot was built in 1942 and 1943 and was commissioned as a United States Navy facility on March 30, 1943. It served as a storage and supply depot for naval forces along the Atlantic coast and Europe, and as a storage and distribution point for National Stockpile materials.

On January 1, 1960, the Navy turned the facility over to the GSA. During the period between early 1966 and approximately 1973, the USACE/Army Material Command (AMC) leased buildings from the Navy for the fabrication and storage of vehicles as well as other military equipment. Additionally, between 1967 and 1969, the GSA and the Navy leased to the United States Army/Defense Supply Agency, Buildings 202 and 203. The agreement indicates these buildings were used for the preservation and rail loading of trucks; and storage of trucks and vehicles.

1.2.1 Summary of Previous Investigations

In the late 1980s, trichloroethene (TCE) was detected at low-level concentrations of less than 1 microgram per liter ($\mu\text{g/L}$) in the Town of Rotterdam and City of Schenectady well fields. In an effort to determine the potential source(s) of the TCE, the New York State Department of Health (NYSDOH) performed sampling of private water supply wells in the area during 1991. The private water supply sampling included residences located on NYS Route 5 in the Town of Glenville hydraulically downgradient of the Defense National Stockpile Center Scotia Depot Site. VOCs, including TCE, 1,1,1-trichloroethane (1,1,1-TCA), and tetrachloroethene (PCE), were detected in groundwater collected in some of these residential wells. The sampling results were consistent with the known groundwater contamination concentrations at the Defense National Stockpile Center Scotia Depot Site, including TCE which was detected in the NYS Route 5 residential well water samples at concentrations up to 320 $\mu\text{g/L}$. Following a recommendation by the NYSDOH to connect to public water, the homes on NYS Route 5 were subsequently connected to public water provided by the Town of Glenville. Although the drinking water standard was never exceeded in the City of Schenectady and the Town of Rotterdam municipal water supply wells, increased groundwater quality monitoring was initiated following the identification of the contamination.

Subsequent to the NYSDOH residential groundwater sampling, six subsurface investigations were completed to identify the possible source of TCE in the residential wells and to delineate the extent of the TCE groundwater plume. The investigations were completed between 1995 and 2007 and focused on the assemblage of properties comprising the former 337-acre Defense National Stockpile Center Scotia Depot. The New York State Department of Environmental Conservation (NYSDEC) 2007 Expanded Site Investigation (ESI) (NYSDEC, 2007) provides details on each of these investigations. Investigation data indicated that TCE disposal may have also occurred in the northeast corner of the 401 sub-block and the area near the north corner of the 403 sub-block.

Based on these investigations, a Record of Decision (ROD) specifying a groundwater remedy was approved by the NYSDEC in March 2010 (NYSDEC, 2010). The ROD specified a remedial action for the groundwater plume which included treatment of the plume through the installation of a zero valent iron (ZVI) PRB. During this time investigations were also conducted in relation to a carbon tetrachloride plume that was identified as a source for potential soil vapor intrusion. In addition to the groundwater remedy, the ROD also identified the need for soil vapor intrusion mitigation at the building 201 sub-block. Details on the installation and monitoring of the Soil Vapor Intrusion (SVI) portion of the remedy are provided in the Final Engineering Report (FER) (AECOM, 2017a). A Site Layout Map is provided in Figure 1-2.

1.2.2 Pre-Design Groundwater Investigation – 2013

A pre-design investigation (PDI) was completed by Stone Environmental in 2013 to verify the location and dimensions of the TCE plume to better estimate the appropriate location and depth of the PRB. The PDI was completed as a component of the ROD selected remedy to aid in the PRB design. The pre-design investigation included:

- Baseline groundwater sampling of 24 existing onsite monitoring wells
- Synoptic measurement of groundwater elevations in 35 on-site and off-site monitoring wells
- Vertical groundwater profile of VOC plume at 16 locations (WP-01 to WP-16)
- Installation and development of four on-site monitoring wells (MW-24 through MW-27)
- Hydraulic conductivity measurements
- Geotechnical soil sampling (laboratory sieve, bulk density, and effective porosity analyses)
- ZVI treatability study (bench-scale column test) using Site soil and groundwater

The results of the PDI indicated that the plume location had shifted to the south/southeast from the estimated plume delineation shown in the 2010 ROD (see Figure 3 from the ROD and Figures 6 and 10 from Final PDI Report) (Stone, 2013). The PDI also delineated the vertical and horizontal limits of the plume across a transect of groundwater profile locations, which had not been well defined in previous investigations. The results of the ZVI treatability study indicated that ZVI would be effective in remediating the TCE plume at the detected maximum concentrations and Site-specific geochemical conditions. The PDI evaluated a preliminary PRB design approximately 850-feet long centered on the highest concentration axis of the TCE plume and extending to estimated lateral limits of the plume based on the results of the vertical groundwater profile locations. Subsequent evaluation of the data to maximize effectiveness and efficiency of the remedial design suggested a 700-feet long deep section centered on the TCE plume with a shallower 250-feet long section to treat lower TCE concentrations would be effective at mitigating the groundwater contamination.

1.2.3 Baseline Groundwater Investigation

As part of the remedial design investigation work plan (RDIWP) (AECOM, 2015) various field activities were conducted during the fall of 2015 in order to gather data and information needed to complete the final PRB design. The main components of the remedial design investigation (RDI) field activities that related to the PRB design included:

- Installation and development of four compliance well pairs (MW-28 to MW-35) and one additional monitoring well (MW-36) to confirm upgradient edge of groundwater plume

-
- Collection of 33 baseline groundwater samples
 - Performance of a confirmatory ZVI bench scale test
 - Performance of aquifer tests including slug testing and hydraulic pulse interference testing (HPIT)

Detailed methods and results of these field activities were presented in the Remedial Action Work Plan (PRB-RAWP) (AECOM, 2016) and the 2015 RDI Work Summary Memo presented in Appendix A of the PRB-RAWP.

1.3 PRB Design Summary

The remedial investigation activities at the Site indicated that variable hydraulic conductivity and hydraulic gradient conditions may exist at the Site, resulting in varied groundwater velocity. Therefore, various design cases were analyzed within the range of the measured values to determine the optimum design for the PRB. Three design cases in particular were outlined in the (PRB-RAWP) (AECOM, 2016). These design cases were based on average values from the slug test data and HPIT data from the 2015 RDI activities and historic data from the Stone PDI (Stone, 2013). The three design cases used an average value of 0.004 ft/ft for the hydraulic gradient and varied the hydraulic conductivity from 15.66 ft/day to 193.8 ft/day. Using an assumed porosity of 0.4 this variability of hydraulic conductivity results in an expected range of groundwater velocity at the Site from 0.128 ft/day to 2.83 ft/day. GeoSierra Environmental, Inc. (GeoSierra), the PRB installation subcontractor, performed a sensitivity analysis based on these design cases and the design of the PRB was chosen based on design scenarios that reflected a conservative approach. A full description of the PRB design including details of each design case is presented in the PRB-RAWP (AECOM, 2016).

1.4 Remedial Action Implementation

In accordance with the ROD for the remedial action at the FSND, a ZVI PRB was installed in order to mitigate the impacted groundwater plume at the Site. AECOM, and its subcontractor GeoSierra, performed the installation of the PRB over the course of 10 months in 2016. The design and installation procedures of the PRB are outlined in the PRB-RAWP (AECOM, 2016). The main components of PRB installation were as follows:

- Installation of 77 injection wells
- Installation of 31 Resistivity strings
- Placement of ZVI into the formation via injection wells
- Post PRB installation HPIT testing

The installation of the ZVI PRB was successfully completed in November of 2016. Details of the PRB construction activities of the PRB are provided in the FER (AECOM, 2017a) for the Site.

2 GROUNDWATER MONITORING PROGRAM

Monitoring well locations are shown on Figure 1-2 and are described in Table 2-1. Table 2-1 also provides the monitoring well sampling schedule and analytical information for the groundwater monitoring program. The December 2019 event included groundwater monitoring from 12 locations. The groundwater monitoring program will be carried out in accordance with the schedule and sampling protocol outlined in the SMP.

The eight compliance monitoring wells (MW-28 through MW-35) were installed in pairs so that groundwater quality could be monitored directly upgradient and directly downgradient of the PRB. The four monitoring well pairs are installed 20 feet apart on opposite sides of the wall, one being upgradient and one being downgradient, with corresponding screen depths. Figure 2-1 provides a profile of the compliance monitoring wells showing the screened interval in relation to the PRB. Results from the groundwater monitoring program will be used to evaluate the effectiveness of the remedy at decreasing chlorinated VOC concentrations in groundwater and preventing the migration of contaminated groundwater off-site. The compliance well pairs, in addition to MW-24 (downgradient), MW-26 (downgradient), MW-15 (upgradient) and MW-16 (outside of plume), were to be sampled quarterly for the first two years (eight quarters) then semi-annually thereafter. The first quarterly sampling event was conducted in December 2016. The first semi-annual sampling event was conducted in December 2018. Monitoring well locations are shown on Figure 1-2 and are described in Table 2-1.

2.1 Sample Collection Methods

Prior to sample collection, depth to water measurements were collected with an electronic water level meter from all accessible wells on Site. Depth to water measurements were taken to the hundredth of a foot from a designated measuring point on the well casing.

The groundwater sampling event was performed in accordance with EPA's low stress, often referred to as low-flow, sampling technique (Low-Flow (Minimal Drawdown) Ground-Water Sampling Procedures, EPA/540/S-95/504) (EPA, 2010) and is discussed below.

A bladder pump was used to purge the monitoring wells with the pump intake set at the midpoint of the saturated screened interval. During purging, the pump was operated at a flow rate of approximately 100 to 500 milliliters per minute (mL/min) and water levels were monitored to ensure that the pumping rate caused minimal/no drawdown. Dedicated tubing for each monitoring well was used for groundwater sample collection. Field parameters were recorded on the Well Sampling Forms every five minutes during purging, including:

- Purge rate (mL/min)
- Depth to water (0.01 ft)
- Temperature (degrees Celsius)
- pH

-
- Specific conductance (millisiemens per centimeter [ms/cm])
 - Dissolved Oxygen (DO) (milligrams per liter [mg/L])
 - Oxidation-Reduction Potential (ORP) (millivolts [mV])
 - Turbidity (NTU)

A flow-through cell was used to obtain temperature, pH, specific conductance, DO, and ORP. Turbidity will be measured using a separate instrument. Purging was considered complete when the indicator parameters have stabilized over three consecutive readings, indicating that formation water was being drawn. Stabilization parameters include the following:

- Drawdown: less than 0.3 ft drawdown during purging
- pH: ± 0.1 standard unit
- Specific Conductivity: $\pm 3\%$
- DO: $\pm 10\%$ (mg/L) for values greater than 0.5 mg/L or 3 readings < 0.5 mg/L
- ORP: ± 10 mV
- Turbidity: < 5 NTU or $\pm 10\%$ for readings > 5 NTU

Groundwater sample collection forms with the field parameter readings for each monitoring well are included as Appendix A. Sampling instruments were calibrated daily prior to starting sampling activities, or as needed throughout the day. A log of the field equipment calibration records is provided in Appendix B.

During past groundwater sampling events observed DO readings have been variable between sampling events and monitoring well locations, and sometimes readings were higher than expected. To get a better understanding of DO measurements in the field during the December 2018 and June 2019 events DO readings were collected with the flow through cell using a YSI Pro DSS, as well as a YSI 550A DO meter placed directly in the monitoring well to compare differences in measurements. The YSI Pro DSS has an optical DO meter, and the YSI 550A DO meter is a membrane based DO meter. Periodic in field verification of the DO readings was performed by inserting the probes into a 0 mg/L DO solution. The use of the YSI 550A DO meter was discontinued for the December 2019 sampling event.

Prior to sample collection, the flow-through cell was disconnected from the dedicated sample tubing and the sample was collected directly from the tubing into the laboratory supplied sample containers. The target flow rate during sample collection was approximately 100 mL/min and sample collection was completed within a single bladder pulse for VOC analysis. Once sampling was complete, the purge water was placed in a 55-gallon drum at the conclusion of the sampling event. More detailed procedures for sample collection and handling and waste handling, are

included in Appendix H of the SMP (AECOM, 2017b). Appendix G of the SMP includes the analytical QAPP for the site management activities. Appendix I of the SMP includes the HASP for the site management activities.

Groundwater samples were packaged on ice and delivered to ALS Laboratory daily via FedEx shipping during the sample collection timeframe. Standard chain of custody procedures were used for sample transport. In total, 12 groundwater samples were collected and analyzed for targeted VOCs (EPA method 8260C) and monitored natural attenuation (MNA) parameters including TOC (SM 5310B), alkalinity (SM 2320B), chloride, nitrate, sulfate (EPA Method 300.0), and dissolved gases (methane, ethane, and ethene; Method RSK 175). In addition, five groundwater samples were analyzed for dissolved hydrogen (PACE Method AM20GAX), and eight groundwater samples were analyzed for total and dissolved iron (EPA SW-846 Method 6010C).

3 RESULTS

3.1 Hydrogeologic Results

The groundwater elevations for the Site were determined based on the initial depth to groundwater measurements that were taken prior to sample collection. Table 3-1 shows the groundwater elevation data for the December 2019 sampling event and compares it to the November/December 2015 baseline sampling event and past sampling event levels. A potentiometric Site map indicating the overburden, groundwater elevation and direction of groundwater flow during the December 2019 sampling event is included as Figure 3-1. Observed general groundwater flow direction in December 2019 was from northeast to southwest, which is similar to past sampling events.

Based on observed trends during the past sampling events groundwater elevation at the Site is subject to seasonal variability. The December and March sampling events exhibit lower groundwater elevations than the June and September sampling events. The similarities between the winter events and summer events indicate a seasonal groundwater level trend. Groundwater elevation data for the December 2019 sampling event indicate that groundwater levels are currently above the top of the PRB wall at the southern end of the wall. Similar groundwater levels have been observed during other sampling events both in the summer and the winter, suggesting that the recent average groundwater level at the site may be different from when the wall was designed. The current potentiometric surface in relation to the PRB is shown in profile on Figure 2-1A and in relation to along the axis of the estimate plume in Figure 2-1B. AECOM has reviewed the vertical COC groundwater concentration data from the 2013 PDI and found that there were no detected COCs at elevations above the wall in this area. AECOM does not believe that groundwater being slightly above the top of the wall in this area is cause for concern because the concentrations of COCs in the elevations around the top of the wall were low or non-detect. The majority of the higher COC concentrations were detected at locations in the lower section of the wall.

The hydraulic gradient is the change in hydraulic head, or water level, per unit distance. The average hydraulic gradient at the Site in the vicinity of the PRB, estimated based on the December 2019 hydrogeologic conditions, was determined to be 0.0069 ft/ft. The hydraulic gradient for the past three sampling events was 0.0038 ft/ft in June 2019, 0.0044 ft/ft in December 2018 and 0.0019 ft/ft in September 2018. The groundwater seepage velocity is the rate of solute transport through the open pore space in the soil. Based on the December 2019 hydraulic gradient of 0.0069 ft/ft and the range of hydraulic conductivities evaluated for the PRB design (15.66 ft/day to 193.8 ft/day), and assuming a porosity of 0.4, groundwater seepage velocity at the Site could vary between approximately 0.27 ft/day and 3.33 ft/day, which is slightly higher than the range of estimated groundwater velocities used for the PRB design (0.128 ft/day-2.83 ft/day). Calculations for hydraulic gradient and velocity estimates are included in Appendix C.

Drums containing purge water from the December 2019 sampling event were removed from the Site on January 14, 2020 and their contents were properly disposed of by the environmental waste services contractor.

3.2 Groundwater MNA and Field Parameter Results

Results of groundwater MNA parameters obtained from the baseline sampling event through the December 2019 semi-annual sampling event for the PRB monitoring compliance wells are presented in Table 3-2. MNA parameters were compared between compliance well pairs to observe changes in groundwater quality from upgradient to downgradient of the wall. Field parameter readings are primarily collected to demonstrate stabilization criteria as discussed in Section 2.1, however they can also give insight as to subsurface conditions and served as indicators for reactions taking place.

During previous sampling events DO measurements were variable with some well pairs showing an increase and some pairs showing a decrease from upgradient to downgradient. ORP values downgradient of the wall are expected to be within the -300 to -400 range, and we should expect to see little to no DO in these monitoring locations. To date these expected values have not been observed on a consistent basis and no definitive trends on DO and ORP measurements have been defined.

DO measurements during the December 2019 event were collected with an optical DO meter (YSI Pro DSS) located in a flow through cell. For most compliance well pairs, the downgradient DO readings were very low (i.e. <1.0 mg/L), which is within the range of what we expect to see in downgradient compliance wells. All downgradient compliance wells DO readings were lower than their upgradient compliance well pairing. Periodically during the sampling event, field verification of the DO reading was performed by inserting the probe into a 0 mg/L DO solution, and these checks showed that the instruments were reading 0 mg/L or very close to 0 mg/L during the checks. The DO data is presented in Table 3-2. For past recent sampling events a second DO meter, the YSI 550A, was used to measure DO in the monitoring well. AECOM found that these results were not consistent to the YSI Pro DSS results for unknown reasons, and did not clarify the redox status of the groundwater. The DO measured in the flow-through cell appeared to be more consistent with the expected geochemistry. Therefore, to eliminate confusion on the data and gain efficiency during the field event the use of the downhole YSI 550A was discontinued for the December 2019 event and future sampling events.

During the December 2019 sampling event ORP levels were again variable but measurements showed decreases between three out of four compliance well pairs. However, ORP values were still not showing values that are typically expected downgradient of a ZVI PRB wall. ORP levels decreased significantly from upgradient to downgradient at well pairs MW-28/29 and MW-32/33. Lower ORP values are expected downgradient of the PRB indicating reducing conditions as the groundwater passes through the PRB, however we expect to see ORP levels in the range of the -300 to -400 range for the Beta elimination CVOC reduction to occur.

The December 2019 groundwater results showed a general increase in methane, ethane and ethene in well pairs MW-30/31 and MW-32/33 when compared to their upgradient counterparts. For methane, there was a particularly large increase between upgradient and downgradient in monitoring well pairs MW-30/31 and MW-32/33. The largest increase in methane, was seen in compliance monitoring well pairs in the middle of the PRB. Increased downgradient methane

concentrations began to be observed by mid-2017 and since then the methane levels have been relatively sustained in the mid-section of the wall. Initially methane, ethane and ethene concentrations increased from the breakdown of the ZVI carrier fluids (guar) and served as an indicator for biological reductive dechlorination activity in the subsurface. Elevated levels of methane continue to be observed, shifting from the southern end to the middle of the wall's downgradient compliance wells, suggesting the continued occurrence of anaerobic biological activity in the subsurface in some portions of the wall.

To date, nitrate and sulfate levels have been variable since the 2015 baseline sampling event including for the December 2019 event. During this event most downgradient wells showed a decrease in nitrate and sulfate from their upgradient counterparts. Nitrate and sulfate concentrations are expected to decrease from upgradient to downgradient wells as this would further indicate that bioactivity is occurring.

The continued presence of ethane and ethene in downgradient well pairs is indicative of the β -elimination abiotic reaction of CVOCs with the PRB. These compounds, along with acetylene, are final products from the interaction of the ZVI and CVOCs. The June 2018 results did not indicate the presence of acetylene, however the detection of acetylene is very rare since it is extremely volatile. Acetylene could be produced as a byproduct of abiotic TCE reduction. Its presence would be an indicator of an abiotic TCE reduction, but its absence does not necessarily provide any insight into the reactions that may be occurring within the ZVI PRB. For the 10 locations where acetylene was analyzed for in June 2019, there were no detections. Therefore, acetylene was not analyzed for in the December 2019 event.

The December 2019 event included sampling for dissolved hydrogen at five of the compliance wells to monitor for the expected geochemical conditions present from the CVOC reactions with the ZVI wall. Dissolved hydrogen is a byproduct of the CVOCs degradation reactions taking place in the wall and may serve as an indicator of ZVI reactivity. Results showed very low concentrations of dissolved hydrogen in the vicinity of the wall at both upgradient and downgradient locations; however, the lack of detections of dissolved hydrogen is not unusual since dissolved hydrogen is quickly consumed in the subsurface. The background locations sampled in June 2019 (MW-15 and MW-24) also showed very low dissolved hydrogen concentrations, similar to those within the compliance monitoring well locations.

Dissolved iron and total iron were sampled for at the eight compliance wells during this quarter to monitor for reaction byproducts in the vicinity of the wall. Dissolved iron is another byproduct of ZVI redox reactions with CVOCs and it is expected to see increased concentrations in the subsurface at monitoring wells downgradient of the wall. Neither dissolved nor total iron was detected in concentrations at wells downgradient of the PRB that are indicative of presence of ZVI redox reaction byproducts.

Overall the MNA data does not show consistency in the well pairs throughout the expanse of the PRB. The December 2019 monitoring data may indicate that the groundwater conditions at the site are shifting away from the anaerobic conditions that were created in the wake of the PRB wall installation, to conditions that indicate that abiotic reductive dechlorination reactions are occurring

in the middle of the wall. However, MNA parameters to date have not yet shown signs indicative of the redox reactions that would be expected to take place as the COCs flow through the ZVI PRB wall. ZVI reactions typically show a significant increase in pH to about 9-10 in downgradient monitoring locations depending on the buffer capacity of the soil. However, pH measurements observed to date downgradient of the wall have been mostly the same, typically between 7-9, since the installation of the PRB with only some minor increases throughout time.

3.3 Groundwater VOC Results

The VOC results from the June 2019 sampling event are presented in Table 3-2. This groundwater sampling event included collection of 12 groundwater samples for VOCs. Figure 3-2 provides a summary of the groundwater VOC results for the monitoring well compliance pairs, and semi-annual wells that exceed the NYSDEC Ambient Water Quality Standards (AWQS) and Guidance Values (GV) found in the Technical and Operational Guidance Series (TOGS) 1.1.1 (NYSDEC, 1998) and compares the December 2019 sampling event results to the historic sampling event results. Figure 3-3 shows the Site-wide monitoring locations total CVOC results. Full analytical reports are included in Appendix D.

The laboratory data was validated by an AECOM chemist and a full data usability summary report (DUSR) was prepared. The DUSR, included in Appendix E, indicated that all data points were usable with some qualifications and no data points were rejected. Further details on other data that required qualifications are provided in the DUSR.

A narrative summary of the results is presented below:

- TCE, the primary constituent of concern, was detected in all 8 of the compliance wells, 7 of which were above or equal to the AWQS of 5 µg/L. One downgradient monitoring well location, MW-34, for the second consecutive sampling event showed the TCE concentration to be below the AWQS.
- For the December 2019 event three out of the four downgradient wells of the compliance well monitoring pairs showed lower levels of TCE concentrations than their upgradient counterparts. The downgradient TCE concentrations ranged from 2.9 µg/L (MW-34) to 219 µg/L (MW-28).
- TCE was also detected at varying levels in two of the four other wells that were sampled, including elevated levels of TCE at MW-15 (105 µg/L) upgradient of the wall. The TCE concentration at downgradient well MW-24 was below the AWQS. While there may be some seasonal variability in TCE concentrations, at this time there does not seem to be a clear pattern in the results.
- In general detected concentrations of TCE in the 12 wells, as well as other CVOCs, for the December 2019 sampling event were consistent with previous groundwater sample results. However, as previously noted, there appears to be a decreasing trend in TCE concentrations at some downgradient compliance well pair locations.

-
- Wells with detectable levels of tetrachloroethene (PCE) were MW-15, MW-28, MW-29 and MW-31. The concentration of PCE measured in MW-28 and MW-29 were above the AWGS of 5 µg/L and the others were below the AWQS.

Graphs showing concentrations of CVOCS were created for the monitoring well compliance pairs to monitor groundwater concentration trends. Data shown includes the baseline sampling event in December 2015 through the most recent sampling event in December 2019. These trend plots are included in Appendix F as Figures F-1 through F-4. Analysis of concentration trends so far indicates a decreasing trend in concentration of CVOCs in all downgradient compliance monitoring wells except MW-28. Over the last several monitoring events in particular a decrease in total CVOCs from upgradient to downgradient compliance well pair locations has been observed.

4 SUMMARY AND CONCLUSIONS

The December 2019 groundwater monitoring event was the third semi-annual sampling event and the eleventh overall sampling event completed since the installation of the PRB wall. Sampling is now conducted on a semiannual basis for a 3 year period as described in the SMP (AECOM 2017b).

Mann-Kendall trend analysis of the laboratory results for TCE suggests that there is a decreasing trend in TCE concentrations at the downgradient compliance wells located in the deeper portion of the PRB. MW-28, screened in the shallow portion of the wall, does not exhibit a trend. Trend analysis results are included in Appendix G. CVOC concentrations at downgradient well pair locations have in general been lower than their upgradient counterparts for the last several monitoring events. Continued observations of elevated methane and ethane concentrations at some downgradient monitoring wells, particularly in the center of the wall, indicate the presence of anaerobic biological activity within the subsurface in the vicinity of the PRB. While there had been increased TOC concentrations at the monitoring well compliance pairs noted in the previous sampling events it appears TOC has moved toward baseline conditions. Results from the December 2019 event indicate this trend is continuing. Since the installation of the wall downgradient parameters including the presence of ethane, ethene, and methane suggest that the biotic degradation of TCE may be taking place as impacted groundwater flows through the PRB in most well pairs. However, more recent dissolved gasses data indicates that a shift from biotic reductive dechlorination to abiotic reductive dechlorination conditions and reactions may be occurring in the subsurface.

The laboratory and field data from the MW-28/29 well pair at the northern end of the wall is noticeably different from the other three well pairs along the wall and has not yet shown the decreasing trend of CVOCs at the downgradient monitoring location that has been observed in the other three well pairs. This well pair has consistently shown the presence of TCE degradation daughter products which are not very prevalent at any of the other well pairs. Furthermore, field parameters in this well pair are different from other locations along the wall. These differences in data could be due to MW-28/29 well pair being screened in the upper portion of the PRB which is in a different geologic layer of the aquifer that is more transmissive than the lower layer. Groundwater hydrology is also different in this area of the Site and the hydraulic gradient appears to be flat in this area, especially during the summer. AECOM and the USACE are considering how to better evaluate the remedy for the area around the MW-28/29 well pair, and will continue to closely review data to look for trends for this well pair during subsequent sampling events.

Current Site groundwater flow conditions indicate that on average the hydraulic gradient is consistent with the PRB design, with the December 2019 event showing slightly elevated gradients as compared with the other sampling events. The PRB was designed based on a hydraulic gradient of 0.004 ft/ft which is lower than the estimated hydraulic gradient of 0.0069 ft/ft measured in December 2019. Seasonal variability in groundwater elevations and overall hydraulic gradient has been observed throughout the past sampling events. During the winter months when the Mohawk River is lowered, there appears to be an effect on the groundwater elevation at the site, and groundwater elevations are lower than in the summer. In turn, this could contribute to the slightly

higher hydraulic gradients observed during these times. The gradient at the Site is slight and at times there appears to be a reverse gradient in the MW-28/29 well pair at the northern edge of the wall. In actuality this is likely an area with a flat gradient and the minor variability in groundwater levels between the well pair is due to margin of error in the survey of the well casing or with the field measurements. Historic data indicates a range of gradients from 0.001 to 0.006 ft/ft measured at the Site (Stone 2013).

At this time monitoring results have not yet shown the expected decrease in COC concentrations; however, they have shown evidence of groundwater flow through the wall in that increased concentrations of methane have been observed at downgradient locations. As described in the PRB RAWP, expectations are that ZVI PRBs will function for at least 30 years with the possibility of a greater lifetime depending on site conditions. Approximately three years have elapsed since the completion of PRB construction, so the wall is well within the expected operational lifetime. Although concentrations of COCs at downgradient compliance wells are lower than their upgradient counterparts, the observed concentrations are not as low as they were expected to be at this point in the project. Preliminary modeling conducted during the project proposal using the existing conceptual site model data indicated that COC concentrations were originally estimated to be at or below MCL cleanup standards at 10 feet downgradient of the wall approximately 6 months after the PRB wall installation effort was completed. Subsequently, the slower than expected decline in COC concentrations have prompted discussions on possible investigative actions with the USACE in order to work on a path forward to get clarity on groundwater geochemistry and hydrogeology around the PRB wall in order to help evaluate remedy effectiveness.

The next groundwater sampling event is scheduled for June 2020 and will include sampling from 34 locations across the site.

In 2018 dissolved hydrogen, dissolved acetylene and total and dissolved iron were added to the sampling parameters in order to help gain an understanding of the PRB wall performance and groundwater geochemistry. At this time AECOM and the USACE no longer feel that the dissolved acetylene and dissolved hydrogen results are producing data that is helpful and therefore will be eliminating these parameters from future sampling events. The use of the down-well DO meter will also remain discontinued, and the DO will be measured using the YSI Pro DSS in the flow through cell. Iron sampling will be retained.

5 REFERENCES

AECOM, 2015. Remedial Design Investigation Work Plan for the Defense National Stockpile Center Scotia Depot, Town of Glenville, NY. November.

AECOM, 2016. Permeable Reactive Barrier Remedial Action Work Plan for the Defense National Stockpile Center Scotia Depot, Town of Glenville, NY. April.

AECOM, 2017a. Final Engineering Report for the Defense National Stockpile Center Scotia Depot, Town of Glenville, NY.

AECOM, 2017b. Site Management Plan for the Defense National Stockpile Center Scotia Depot, Town of Glenville, NY. Revised, November 5, 2018.

AECOM, 2017c. Groundwater Monitoring Program 2016 Fourth Quarter Status Report for the Defense National Stockpile Center Scotia Depot, Town of Glenville, NY. April.

AECOM, 2017d. Groundwater Monitoring Program 2017 First Quarter Status Report for the Defense National Stockpile Center Scotia Depot, Town of Glenville, NY. May.

AECOM, 2017e. Groundwater Monitoring Program 2016 Second Quarter Status Report for the Defense National Stockpile Center Scotia Depot, Town of Glenville, NY. November.

AECOM, 2018. Groundwater Monitoring Program 2017 Third Quarter Status Report for the Defense National Stockpile Center Scotia Depot, Town of Glenville, NY. February.

AECOM, 2018a. Groundwater Monitoring Program 2017 Fourth Quarter Status Report for the Defense National Stockpile Center Scotia Depot, Town of Glenville, NY. May.

AECOM, 2018b. Groundwater Monitoring Program 2018 First Quarter Status Report for the Defense National Stockpile Center Scotia Depot, Town of Glenville, NY. June.

AECOM, 2018c. Groundwater Monitoring Program 2018 Second Quarter Status Report for the Defense National Stockpile Center Scotia Depot, Town of Glenville, NY. October.

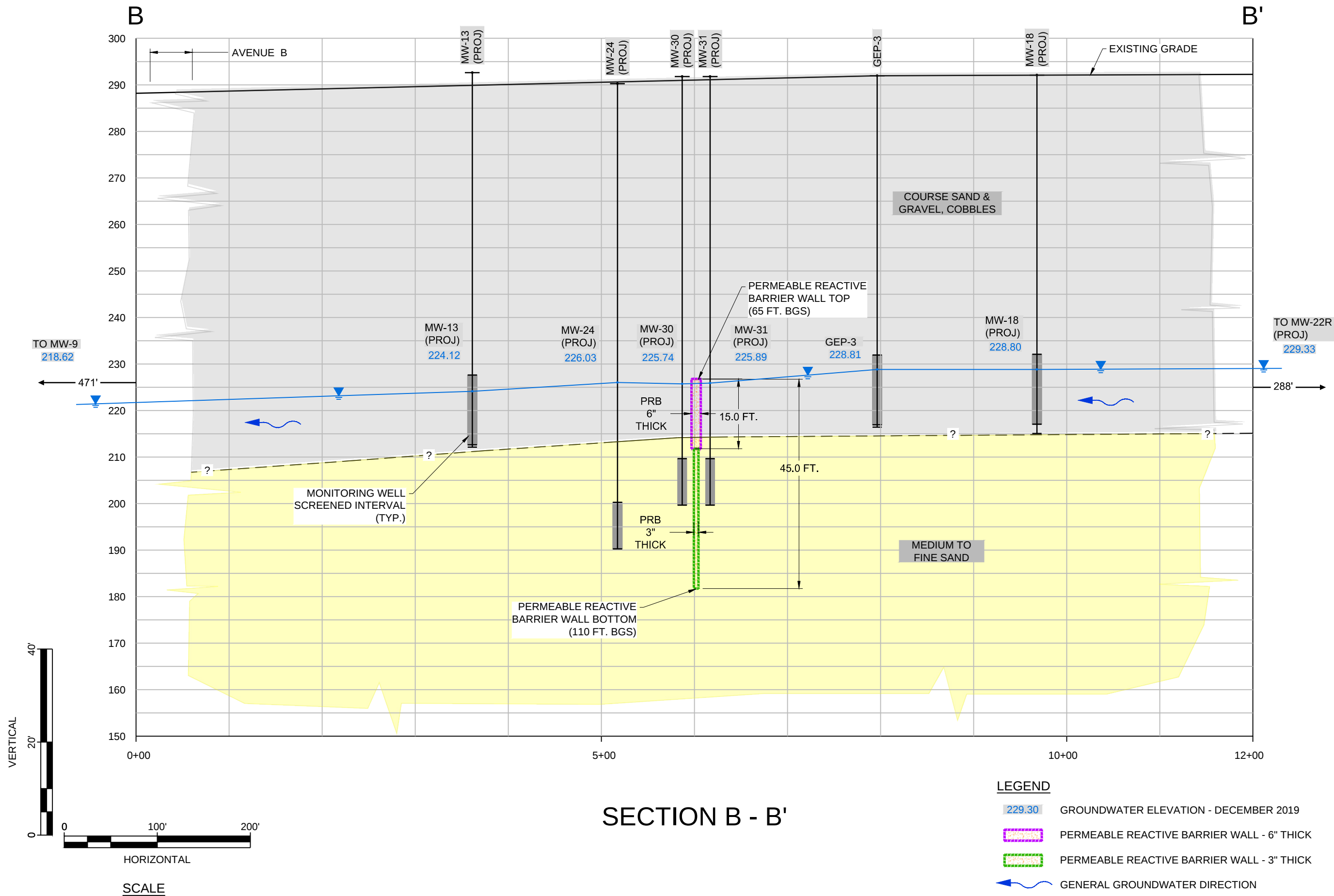
AECOM, 2019. Groundwater Monitoring Program 2018 Third Quarter Status Report for the Defense National Stockpile Center Scotia Depot, Town of Glenville, NY. January.

AECOM, 2029. Groundwater Monitoring Program 2018 Fourth Quarter Status Report for the Defense National Stockpile Center Scotia Depot, Town of Glenville, NY. April.

NYSDEC, 2010. Record of Decision for Defense National Stockpile Center Scotia Depot Site State Superfund Project, Site Number 447023, Town of Glenville, NY, March.

Stone Environmental, 2013. Final Pre-Design Investigation Report, Defense Nation Stockpile Center Scotia Depot Site, Town of Glenville, NY, December.

FIGURES



TABLES

Table 2-1
Monitoring Well Sampling Schedule and Guidelines
The Defense National Stockpile Center Scotia Depot

Monitoring Well ID ¹	Rationale ²	Sampling Frequency	Analytes ³	Screen Interval (ft bgs)
MW-15	Upgradient	Quarterly for 2 years then semi-annually	VOCs/MNA	65-80
MW-16	Outside Plume	Quarterly for 2 years then semi-annually	VOCs/MNA	55-70
MW-24	Downgradient	Quarterly for 2 years then semi-annually	VOCs/MNA	100-110
MW-26	Downgradient	Quarterly for 2 years then semi-annually	VOCs/MNA	100-110
MW-28	Downgradient	Quarterly for 2 years then semi-annually	VOCs/MNA	67-72
MW-29	Upgradient	Quarterly for 2 years then semi-annually	VOCs/MNA	67-72
MW-30	Downgradient	Quarterly for 2 years then semi-annually	VOCs/MNA	82-92
MW-31	Upgradient	Quarterly for 2 years then semi-annually	VOCs/MNA	82-92
MW-32	Downgradient	Quarterly for 2 years then semi-annually	VOCs/MNA	82-92
MW-33	Upgradient	Quarterly for 2 years then semi-annually	VOCs/MNA	82-92
MW-34	Downgradient	Quarterly for 2 years then semi-annually	VOCs/MNA	82-92
MW-35	Upgradient	Quarterly for 2 years then semi-annually	VOCs/MNA	82-92
GEP-3	Upgradient	Annually	VOCs	59.6-74.6
MW-B-3	Outside Plume	Annually	VOCs	47.5-67.5
MW-5	Downgradient	Annually	VOCs	62.5-72.5
MW-6	Downgradient	Annually	VOCs	58.5-68.5
MW-7	Outside Plume	Annually	VOCs	61-71
MW-8	CT Plume	Annually	VOCs	66-76
MW-9	Downgradient	Annually	VOCs	110-120

Monitoring Well ID ¹	Rationale ²	Sampling Frequency	Analytes ³	Screen Interval (ft bgs)
MW-11	CT Plume	Annually	VOCs	65-80
MW-12	CT Plume	Annually	VOCs	65-80
MW-14	Upgradient	Annually	VOCs	65-80
MW-17	Upgradient	Annually	VOCs	60-75
MW-18	Upgradient	Annually	VOCs	60-75
MW-19	Upgradient	Annually	VOCs	62-77
MW-20	Upgradient	Annually	VOCs	63-78
MW-22	Upgradient	Annually	VOCs	63-78
MW-23	Outside Plume	Annually	VOCs	63-78
MW-25	Upgradient	Annually	VOCs	65-75
MW-27	Downgradient	Annually	VOCs	100-110
MW-36	Upgradient	Annually	VOCs	70-80
GEP-2	Upgradient	Annually	VOCs	60.6-75.6
GEP-1	Upgradient	Annually	VOCs	59.6-74.6
GEP-4	Upgradient	Annually	VOCs	60.15-75.15

Notes:

¹ *2015 Compliance monitoring well

² Rationale: Upgradient of PRB wall; Downgradient of PRB wall; Outside of any plume; Within Carbon Tetrachloride (CT) plume

³ Monitored natural attenuation (MNA) parameters include TOC (EPA SM 5310B), alkalinity (EPA SM 2320B), Chloride, nitrate, sulfate (EPA Method 300.0), and Dissolved Gases (Methane, ethane, and ethene; Method RSK 175).

APPENDICES

APPENDIX A: Groundwater Sample Collection Field Forms

